Environmental spaces for palsas and peat plateaus are disappearing at a circumpolar scale

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Abstract. The Aenthropogenic climate change threatens northern permafrost environments. This compromises the existence of permafrost landforms, such as palsas and peat plateaus, which have been assessed to be critically endangered habitats. In this study, for the first time we integrated geospatial datasets and statistical methods, to model the distribution of suitable environments for palsas and peat plateaus across the Northern Hemisphere permafrost region. The models were calibrated using data from years 1950–2000. The effects of climate change on the future distribution of palsas-suitable environments for the landforms were assessed by using low, moderate, and high emissions scenarios (Representative Concentration Pathways; RCP2.6, RCP4.5 and RCP8.5, respectively) for two periods (2041–2060 and 2061–2080). Hotspots for palsas and peat plateau environments occurred in Northern Europe, Western Siberia, and subarctic Canada. Climate change was predicted to cause an almost complete loss (decrease of 98.2 %) of suitable environmental spaces under thea high emissions scenario by 2061–2080, while under-a low and moderate emissions scenarios the predicted loss was 76.3 and 89.3 % respectively were predicted to disappear. Our modelling results are in-line with previously published thermokarst data pointing out areas of The comparison with previously published thermokarst data supported our findings regarding the recent degradation of palsa and peat plateau environments. Our results fill the knowledge gaps inprovide new insights into the distribution of the permafrost landforms in less studied areas such as Central and Eastern Siberia. In addition, the predictions provide new understanding of insights into the changing geoecological conditions of the circumpolar region with important implications for greenhouse gas emissions.

1 Introduction

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Northern environments are heavily affected by the climate change (IPCC, 2021). Because of the Aarctic amplification these environments are warming almost two to three times as fast as the global average (You et al., 2021). As the climate changes, permafrost (defined as ground with a temperature of 0 °C or below, for at least two consecutive years; Muller, 1943), is projected to thaw from extensive areas (Wang et al., 2022a), and distinctive permafrost processes and related landforms are threatened (e.g., Liljedahl et al., 2016; Aalto et al., 2017; Borge et al., 2017; Karjalainen et al., 2020). Palsas and peat plateaus are peat hummocks with permafrost cores, which can be primarily found primarily from in regions of sporadic and

discontinuous permafrost (Seppälä, 1988). They mainly differ mainly by in their extent and height (Zoltai and Tarnocai, 1971). Height of palsas varies between 0.5–10 m and their diameter exceed two meters (Washburn, 1983; Pissart, 2002). Peat plateaus have a greater extent, even over one square kilometer, but are usually elevated from their surroundings by only about one meter (Kershaw and Gill, 1979; Zoltai, 1972). Palsas and peat plateaus are not clearly distinguishable from each other and many transitional types between these landforms are found. Peat plateaus can be considered as a morphological class of palsas, so-called palsa plateaus (e.g., Åhman, 1977; Seppälä, 1988).

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According to Hugelius et al. (2020) nearly half of the peatlands of the Northern Hemisphere, including palsa mires and peat plateaus, contain permafrost. Permafrost peatlands, such as palsa mires and peat plateaus, are widespread, and according to Hugelius et al. (2020) nearly half of the peatlands of the Northern Hemisphere contain permafrost. Previous studies have shown that most palsas were formed in considerablymuch colder climate than present (Halsey et al., 1995; Treat et al., 2016a; Vorren, 2017; Fewster et al., 2020). Thus, Palsas and peat plateausthese landforms are highly sensitive to further changes in climate, and many studies have reported rapid degradation of palsas them (e.g., Borge et al., 2017; Mamet et al., 2017; Olymo et al., 2020). As the permafrost thaws from peatlands, changes in the greenhouse gas fluxes are expected (Turetsky et al., 2020; Miner et al., 2022). Increasing CO₂, CH₄ and N₂O -emissions further accelerate the climate change (Marushchak et al., 2011; Schuur et al., 2015; Hugelius et al., 2020).

In addition to their importance in global carbon balance, pPalsas and peat plateaus have an important role in the bio- and geodiversity of Nnorthern environments. Peat controlled permafrost hummocks create topographical and hydrological variability (Seppälä, 1988; Beilman, 2001), resulting different microhabitats for many animal and plant species (Luoto et al., 2004a). For example, palsa mires are well known for their rich bird life (Järvinen and Sammalisto, 1976; CAFF, 2001). Permafrost thaw leads to the collapse of palsas and peat plateaus (Seppälä, 1982, 2006), changes in vegetation (e.g., Malmer et al., 2005; Johansson et al., 2006; Normand et al., 2017) and overall homogenization of peatlands in the end (Swindles et al., 2015). This can lead to changes in bird and invertebrate species composition and affect the functioning of the peatland ecosystem (Luoto et al., 2004a; Markkula, 2014). The European Union classified palsas as critically endangered habitats in 2016, mainly because of the degradation of permafrost (Janssen et al., 2016).

Palsas and peat plateaus in the Northern Europe are relatively well mapped and studied (e.g., Backe, 2014; Ottósson et al., 2016; Ruuhijärvi et al., 2022). There are mapping and monitoring studies conducted also in Canada (e.g., Zoltai et al., 2000; Tam et al., 2014; Mamet et al., 2017) and in the Western parts of Russian (e.g., Barcan, 2010; Kirpotin et al., 2011; Terentieva et al., 2016). However, only a few studies of palsas and peat plateaus are available for Central and Eastern Siberia (Vasil'chuk et al., 2013a, 2013b, 2014). The distribution of palsasthese landforms has been previously modelled mainly at local and regional scales in Northern Fennoscandia (e.g., Luoto et al., 2004b; Fronzek et al., 2006, 2011; Aalto et al., 2017) but recently also in Western Siberia (Fewster et al., 2022), and at a continental scale in North America (Fewster et al., 2020). Previous studies have shown that palsas and peat plateaus require specific climatic conditions (e.g., Luoto et al., 2004b;

Parviainen and Luoto, 2007; Aalto et al., 2017; Fewster et al., 2020, 2022). Besides air temperature and precipitation, wind patterns, vegetation, snow depth, However, topography and soil properties also affect their distribution—of palsas—(Seppälä, 2011). Especially, the role of sufficient peat cover is crucial in marginal permafrost regions, as the thermal properties of peat protect frozen palsa—cores against thawing (Kujala et al., 2008).

In this study, we provide the first predictions of suitable environmental spaces for palsas and peat plateaus over the entire Northern circumpolar region. As the explored landforms are not found from in the Southern Hemisphere (Seppälä, 2011), this study covers a major part of global palsas and peat plateau environments. Our aim is to predict the suitable environmental spaces for palsasthese landforms in a relatively recent period (1950–2000) and in two future periods (2041–2060 and 2061–2080), using two-three Representative Concentration Pathway climate-scenarios (RCP2.6, RCP4.5 and RCP8.5). Moreover, we compare our predictions with thermokarst data produced by Olefeldt et al. (2016), to examine whether our model results are consistent with the development of thermokarst or not. The purpose of the comparison is to assess the consistency of our future predictions as palsas and peat plateaus form thermokarst ponds when ice-rich permafrost thaws (e.g., Seppälä 2011). Consequently, regions with high thermokarst coverage can be assumed to indicate degradation of palsas and peat plateaus. We aim to answer to the following research questions:

- 1) What are the suitable environmental spaces for palsas and peat plateaus in the Northern circumpolar permafrost region?
- 2) How are the suitable environmental spaces for palsas and peat plateaus changing in the future?

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3) Are the changes in suitable environmental spaces for palsas and peat plateaus spatially consistent with observed and predicted thermokarst landscapes?

To address these questions, we used extensive observational data on landform occurrences and high-resolution geospatial data on environmental conditions to predict the suitable environmental spaces with four different statistical modelling methods and their ensemble. Because of the uneven distribution of the landform observations, our results for the Central and Eastern Siberia are extrapolations of models fitted to other parts of the Northern Hemisphere. By extrapolating the models to the Central and Eastern Siberia our Our results offer new insights into on the circumpolar distribution of suitable environments for palsas and peat plateaus, distribution especially for in these poorly mapped regions. Moreover, changes in palsas and peat plateaus the studied landforms can be used as an indicator of the state of sporadic and discontinuous permafrost (Lagarec, 1982; Sollid and Sørbel, 1998). The thermal state of permafrost is an important factor affecting the integrity of arctic transportation and industrial infrastructure (Hjort et al., 2022), release of greenhouse gases (Miner et al., 2022) and ecosystem stability (Goetz et al., 2007; Sim et al., 2021; Magnan et al., 2022). Moreover, future predictions of endangered landforms and habitats can be helpful in conservation actions for different plant and animal species.

2 Materials and methods

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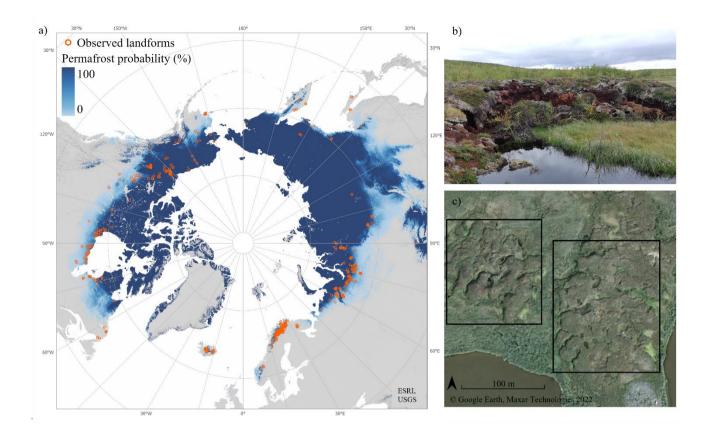
2.1 Palsa and peat plateau observations

Observations of palsas and peat plateaus (Fig. 1a) were <u>mainly</u> collected <u>mainly</u> from available inventories and published studies. In the Scopus and Google Scholar literature search we used search terms 'palsa', 'peat plateau' and 'permafrost peatland' combined with state, province, or region names (e.g., 'Sweden', 'Yukon' and 'Western Siberia') <u>(see Fewster et al., 2022)</u>. Additional observations from less studied areas were compiled from published reports, web pages and Google Earth Pro (version 7.3) and ESRI's ArcGIS Pro (version 2.9.32739) satellite images (the list of used references are provided in Appendix A).

Palsa observations were limited to 'true palsas' (Fig. 1b) based on the description of the landform provided in the used literature (see Appendix A) or a comprehensive interpretation of the satellite images (based on the size, form, and location of the landform and its surroundings for example). whereas 'pPalsa-like' formations and 'lithalsas' (so-called mineral palsas) were excluded, because of the lack of substantial peat cover on these landforms (Pissart, 2002). In this study, peat plateaus were considered as a morphological class of palsas—so called 'palsa plateaus, instead of distinguishing these two landforms from each other, (Zoltai and Tarnocai, 1975; Kershaw and Gill, 1979; Seppälä, 1988; see also Fewster et al., 2020, 2022). All compiled observations were verified by using satellite imagery in Google Earth Pro and ESRI's ArcGIS Pro (Fig. 1c). The final modelling data consisted of 961 grid cells (30 arc sec resolution, ~ 1 km) occupied by palsas or peat plateaus.

Statistical distribution modelling often (i.e., presence absence models) requires absence observations (i.e., presence-absence data) (Brotons et al., 2004; Elith et al., 2006; Zhao et al., 2020). A random sample of 2000 grid cells from areas north of latitude 40° N was taken to compile an absence dataset. All the grid cells were checked using satellite images to ensure that there were no palsas or peat plateaus in the absence grid cells ('true absence'). In total 1496 absence observations were compiled as 504 grid cells were excluded owing to the coarse resolution of satellite images, cloud cover obscuring the view, or the cell occurringed in a water body. The same person (Könönen) collected and validated both presence and absence datasets.

Compiled data were further split into model calibration (N = 2057), and evaluation (N = 400) sets. These sample sizes are expected to be large enough to give robust predictions in geomorphological distribution modelling (see Hjort and Marmion, 2008). The separate evaluation dataset was formed by <u>drawingtaking</u> a random sample from the whole data. The random sample was taken separately from presence and absence observations, to retain the same relative portion (i.e., prevalence) with the original dataset (ca. 40 % presence, and 60 % absence observations) (Hjort and Luoto, 2013). <u>In order to reduce the effect of spatial autocorrelation on model evaluation, o</u>Observations in the evaluation <u>dataset</u> were selected so that they <u>were</u> located at least 50 km from each other <u>(Roberts et al., 2016)</u>. Because of this criteriona, two absence observations had to be removed from the evaluation data, as they <u>were</u> located too close (< 50 km) to presence observations (final N = 398).



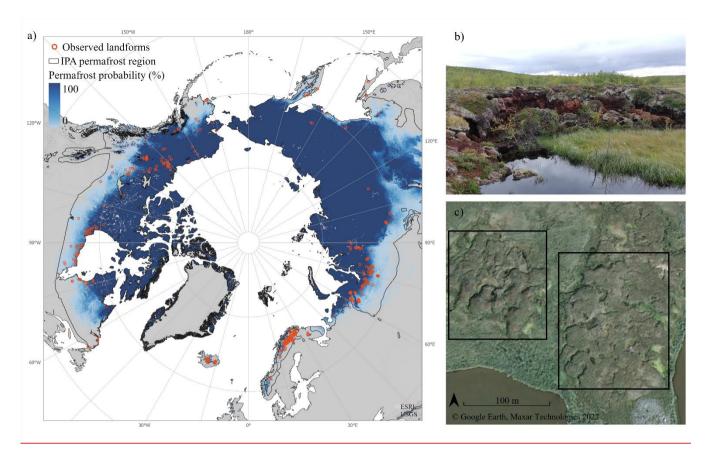


Figure 1: Distribution of the observed palsas and peat plateaus (N = 961) across the Northern Hemisphere, and the probability of the permafrost occurrence (%, Ran et al., 2022) and International Permafrost Association (IPA) permafrost region (Brown et al., 1997) (a). A thawing palsa in Kilpisjärvi, Finland (b), and satellite image of a palsa mire in Kiruna, Sweden (68°28'36.0''N, 20°55'06.0''E), groups of palsas are delimited by have been framed with black lines rectangles (c).

2.2 Environmental data

In this study, we used various geospatial datasets at a 30 arc-second (ca. 1 km) spatial resolution to describe the climateological conditions, soil properties, and topographical variation relevant to palsa mires and peat plateaus. Freezing and thawing degree-days (FDD and TDD, °C-days) were computed to describe seasonal air temperature conditions. Moreover, we used a bioclimatic variable (Bio7 in WorldClim v1.4), which describes the range of annual air temperature (Temp.range, °C, i.e., continentality) by subtracting the minimum temperature of coldest month from the maximum temperature of the warmest month, using the global circulation models (GCMs) available in WorldClim v1.4. Precipitation conditions were considered by calculating sums of liquid (Rainfall, mm) and solid (Snowfall, mm) precipitation. Snowfall was defined as the sum of precipitation for months with average air temperature below 0 °C and rainfall for months with average air temperature over 0 °C (Aalto et al., 2018).

All the environmental variables (hereafter variables) were separately computed separately for different time periods and RCP scenarios, using the WorldClim v1.4 data at 30 arc-second resolution (Hijmans et al., 2005). For these data, the baseline period is 1950–2000, which aligns well with the observations in our presence data. The use of recent climate period instead of the current one better acknowledges the possibility that the compiled landform observations are representing the conditions of past climate. For the climate change scenarios, we used the low emissions scenario RCP2.6, the moderate-emissions scenario RCP4.5 and the high-emissions scenario RCP8.5, and two future periods 2041–2060 and 2061–2080. Climate change projections included in the WorldClim v1.4 database (Hijmans et al., 2005) were derived from an ensemble of 18 global climate models (Taylor et al., 2012).

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Variables to describe the surficial soil conditions (≤ 2 m depth) were obtained from SoilGrids250m 2.0 database (Poggio et al., 2021). Noteworthy, the SoilGrids data layers have been produced using a relatively small amount of data from some permafrost regions, especially Central Siberia and High-Arctic Canada, which may affect the accuracy of data in these regions, especially at finer resolution. SoilGrids were used as alternative soil data were not available in the used (30 arc sec) resolution, and we considered it necessary to account for soil properties, especially soil organic carbon (SOC), which are essential to palsa and peat plateau formation. Here, the SoilGrid variables at the spatial resolution of 7.5 arc-second (~250 meters) were aggregated to the 30 arc-second resolution using bilinear interpolation in ArcGIS Pro. Owing to the lack of high-resolution peat thickness data we used soil organic carbon content (SOC, (g kg⁻¹) as a proxy for the of peat layerto estimate it. In addition to SOC, we used silt content (Silt, g kg⁻¹) and probability of bedrock within two meters from the ground surface (Bedrock, %, Shangguan et al., 2017) to describe the texture and thickness of the soil layer. Topographic Wetness Index (TWI) (Böhner and Selige, 2006) was calculated using the Global Multi-resolution Terrain Elevation Data 2010 (GMTED2010) (Danielson and Gesch, 2011) to characterize the accumulation potential of ground moisture. The grid cell-wise values from each variable were extracted for the presence/absence observations. In case no value from an environmental variable was available at presence or absence location (109 cases), the value was extracted from the neighboring grid cell. If none of the neighboring grid cells had the missing value, the observation was removed from the dataset.

We compared our predictions to a circumpolar thermokarst dataset by Olefeldt et al. (2016). Their prediction of thermokarst coverage is based on six environmental variables including the permafrost zonation, ground ice content, thickness of the sedimentary overburden, ecoregion, and ruggedness of the topography. These variables were not utilized in our study, and thus our analyses can be considered as independent. The dataset includes different types of thermokarst landscapes, and their areal coverages classified into five classes, ranging from none (0–1 %) to very high (60–100 %). We utilized wetland and lake thermokarst coverages as these types can be assumed to be present in degrading palsa mires (Luoto and Seppälä, 2003; Olefeldt et al., 2016). Regions predicted to become unsuitable for palsas and peat plateaus in future scenarios should have

higher thermokarst coverage than persisting suitable environments. In the dataset, thermokarst areal coverages are classified into five classes, ranging from none (0 1 %) to very high (60 100 %).

In addition, we spatially compared our predictions to four peatland datasets (Treat et al., 2016b; Xu et al., 2018, Hugelius et al., 2020; Olefeldt et al., 2021) and two suitable climate spaces to further assess our results. Comparison with the peatland datasets was conducted to examine how well our predictions coincide with the distribution of peatlands, whereas the suitable climate spaces were estimated to recognize the suitable peatlands (i.e., in less studied parts of Siberia). The suitable climate spaces, based on FDD and TDD values, were estimated by a literature review (see e.g., Seppälä and Hassinen, 1997; Luoto et al., 2004b; Aalto and Luoto, 2014; Fewster et al., 2020) and by interpreting results from regional palsa analyses (results not shown). The first classification (FDD: 1500–4500, and TDD: 500–1500) is stricter interpretation of the suitable climate space, and the second (FDD<5000, and TDD<2000) is a more liberal one.

2.3 Statistical modelling

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Statistical modelling was conducted using the biomod2 (version 3.5.1; Thuiller et al., 2021) package in R (version 4.1.3; R Core Team, 2022). We used four methods which have been previously used in distribution modelling of periglacial landforms and processes (e.g., Aalto et al., 2014, 2017; Rudy et al., 2016; Karjalainen et al., 2020;) and other permafrost characteristics, such as soil organic carbon content (Siewert, 2018; Mishra et al., 2021). By using multiple modelling techniques, the prediction uncertainties can be addressed compared to using only one method (Thuiller et al., 2009). The used Mmethods used were generalized linear model (GLM, Nelder and Wedderburn, 1972), generalized additive model (GAM, Hastie and Tibshirani, 1986), generalized boosted model (GBM, Elith et al., 2008) and random forest (RF, Elith et al., 2005). Multiple modelling techniques were utilized in order to acknowledge the uncertainties associated with individual modelling methods (see Thuiller et al., 2009) and to select the one with the highest model performance. In addition, we built an and ensemble of based on the former methods to find out whether it outperforms the individual modelling techniques in the distribution modelling of suitable environments for palsas and peat plateaus. Ensemble approach has been utilized previously for example to predict ground temperature (Aalto et al., 2018), soil hydrology (Cisty et al., 2014), distribution of plants species (Rissanen et al., 2021) and other periglacial landforms (Karjalainen et al., 2020).

GLMs and their semiparametric extensions GAMs, are popular in statistical modelling because they are relatively easy to use and can be utilized for different types of datasets (Hjort and Luoto, 2013). In the calibration of GLM each explanatory variable including their quadratic terms were inserted into the model to take possible curvilinear relationships into account. Variables for the final GLM were selected in a stepwise fashion using the Bayesian information criteria (BIC, Schwarz, 1978) to find the balance between model complexity and performance. In GAMs we used the generalized cross-validation GCV.Cp smoothing algorithm to optimize limit the degrees of freedom with initial values set to three to avoid overfitting.

GAM formula was generated automatically by using the 's_smoother' argument available in the biomod2. Interactions terms between environmental variables were not included in GLM or GAM.

GBM and RF are machine learning methods which build regression/classification trees to obtain robust estimates of the response (Thuiller et al., 2010). These methods include interactions between environmental variables and address potential overfitting (Elith et al., 2005). Here, we used the following parameters for the GBM models: number of trees = 8150, bag fraction = 0.5, interaction depth = 5, and shrinkage = 0.0038. For RF models we fitted 500 trees at maximum with a node size of 5. Three randomly selected variables were used at each split of a classification tree to assign observations to the nodes. In ensemble modelling, results of two or more related models are combined into a single model in attempt to improve the accuracy and predictive capabilities (Hao et al., 2019, 2020; Kit et al., 2021). In this study, the ensemble model was formed to correspond to the mean probabilities over the produced models (400 = 4 modelling techniques, 100 iterations).

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We performed 100 iterations with random sampling of calibration data for each modelling method and built an ensemble of the produced models at each round by averaging the four modelling algorithms. Moreover, the performed 100-fold random sampling of modelling data allowed us to provide uncertainty measures (standard deviation) for the predicted probability values. Because of the uneven distribution of the landform observations, our results for Central and Eastern Siberia should be interpreted as an extrapolation of model fitted for other parts of the Northern Hemisphere.

Future predictions were performed by using the BIOMOD Projection -function in biomod2 (Thuiller et al., 2021; see Karjalainen et al., 2020). Here, the models calibrated for the recent period (1950–2000) were used to predict future suitable conditions by substituting the climate variables with future climatologies (Hijmans et al., 2005). Predictions were performed for each RCP-scenario (2.6, 4.5 and 8.5) and future period (2041–2060 and 2061–2080). Predictions were extracted to the current extent of the suitable environments because we did not consider it temporally plausible for new palsas and peat plateaus to develop in the region of continuous permafrost during the 21st century.

Finally, The predicted probability surfaces for suitable environments were finally classified to binary distribution maps, which were then used to calculate the areas of the suitable environmental spaces. The classification was performed by the true skill statistic (TSS) cut-off values that were estimated during each model run. From these cut-off values (100 model runs per method) the average cut-off value was computed and used for classification. Used cut-off values are presented in the supplements (Table S1). Binary maps were utilized when we calculated the areas of suitable environments. Average variable importance (VI) values (Breiman, 2001; Fisher et al., 2019) were calculated from the results of 100 modeling runs in order to analyze the contribution of each variable to the models. In VI computation, the values of one variable in its turn are shuffled (i.e., a random sample is drawn from the values recorded at the modelling data grid cells) while the other variables are fixed to their mean values. Then, model predictions are produced by using these variables. Finally, VI is

derived from the Pearson correlation between the predictions from the initial models (all variables having their recorded values) and from the models including the shuffled variable with the following Eq. (1) (Thuiller et al., 2009):

235 VI = 1 - cor(initial model predictions, shuffled model prediction). (1)

The closer a VI is to 1, the larger the influence of a given variable (Thuiller et al., 2021).

Calibrated models were evaluated in two ways (Araújo et al., 2005). First, 100-fold cross-validation was performed with the calibration data. At each validation run a random sample of 70 % (N = 1440) of observations were used to calibrate models and the remaining 30 % (N = 617) was used to validate them. Secondly, semi-independent evaluation was conducted by using the separate evaluation data (N = 398), which was set aside before model calibration (as described in the section 2.1). Performance of the models was evaluated with two prevalence-independent statistical measures of classification accuracy, area under the receiving operating characteristic curve (AUC) and TSS (Allouche et al., 2006). Model reliability was also evaluated spatially by exploring the modelling agreement between four independent modelling methods (Luoto et al., 2010), and by the comparison of our predictions and the four peatland datasets (Treat et al., 2016b; Xu et al., 2018, Hugelius et al., 2020; Olefeldt et al., 2021).

3 Results

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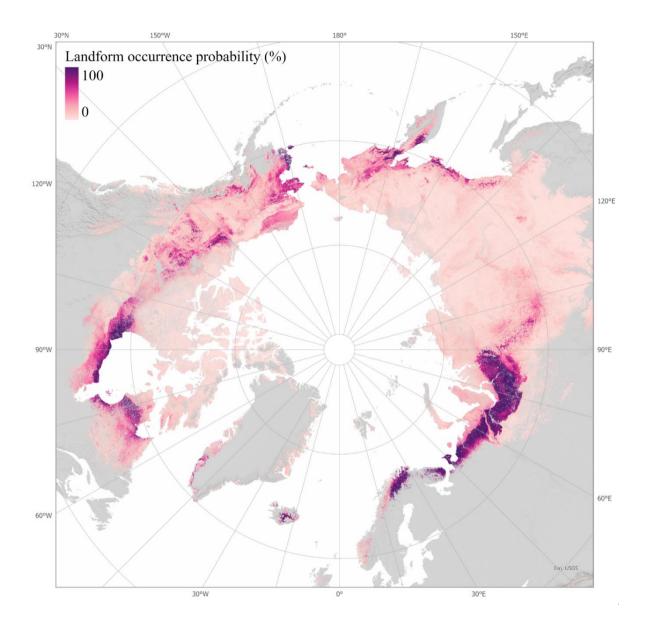
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3.1 Suitable environments in recent conditions

RF had the highest evaluation scores in the model evaluation The presented results are based on the random forest (RF) models, which yielded the highest evaluation statistics among the used methods (see sections 2.3 and 3.4 and supplementary material). and thus the presented results are based on RF. RF model for the recent period 1950–2000 predicted suitable environmental conditions for an area of 1.58 million km². Largest continuous suitable environments located in the Western Siberia, in Canada around the Hudson Bay and Quebec, and in Northern Fennoscandia (Fig. 2). These regions and Iceland had the highest landform occurrence probabilities with only little spatial variation. Other, environments with high landform occurrence probabilities were found from in the Northwest Territories of Canada, west coast of Alaska and east coast of Russiathe Russian Far East. The probability of palsa and peat plateau occurrence decreased relatively sharply outside the most suitable environments (Fig. 2).



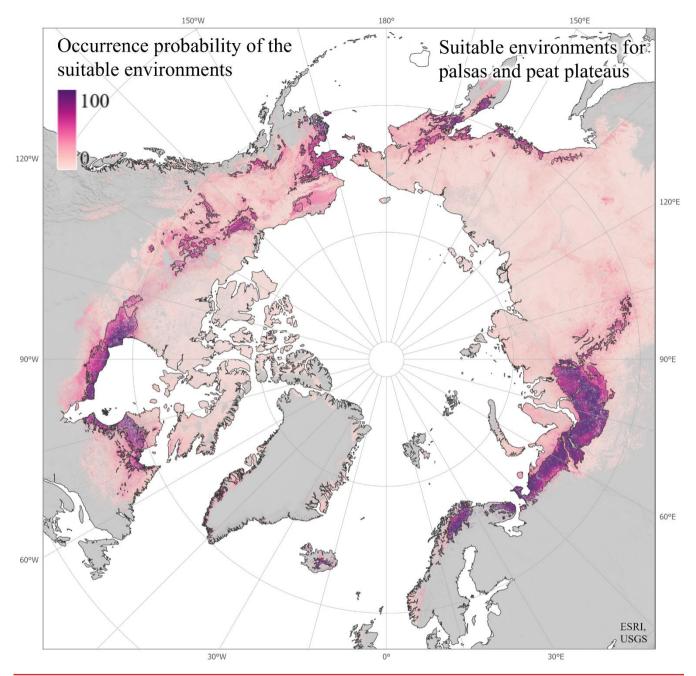


Figure 2: The probability (%) of thefor suitable environments for palsas and peat plateaus occurrence-illustrated with a color gradient from light pink to dark purple, Black outlines delineate the palsa and peat plateau domain, and grey color illustrates areas outside the permafrost region and glaciers. Results of the random forest model are provided for the permafrost region (Ran et al., 2022).

Based on the variable importance (VI) values, TDD was the most important variable (VI = 0.37) to characterize the suitable environments for palsas and peat plateaus in RF models (Fig. 3a., other modelling methods in supplementary material Fig. S3). After TDD, four variables (TWI, SOC, snowfall_a and rainfall) had moderately equal VI values (ranging from 0.062 to 0.077). FDD, temperature range, bedrock, and silt had only a minor importance (VI < 0.021) in the RF models.

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The response curve of TDD shows the optimal summer air temperatures to be around 1000 °C-days, with a steep drop in the probability of occurrence beyond ~1250 °C-days (Fig. 3b). The response of TWI shows a higher occurrence probability increases with higher TWI values, indicating that palsas and peat plateaus are most likely found from flat or gently sloping basin environments with abundant soil moisture. RF model shows the highest probabilities for regions with a SOC content of over 70 g kg⁻¹. The snowfall variable shows also a clear optimum optimal environmental space for the landforms around 200 mm, whereas the pattern is not equally clear for the rainfall (Fig. 3b). Response curves for all used modelling methods are provided in the supplementary material (Fig. S4).

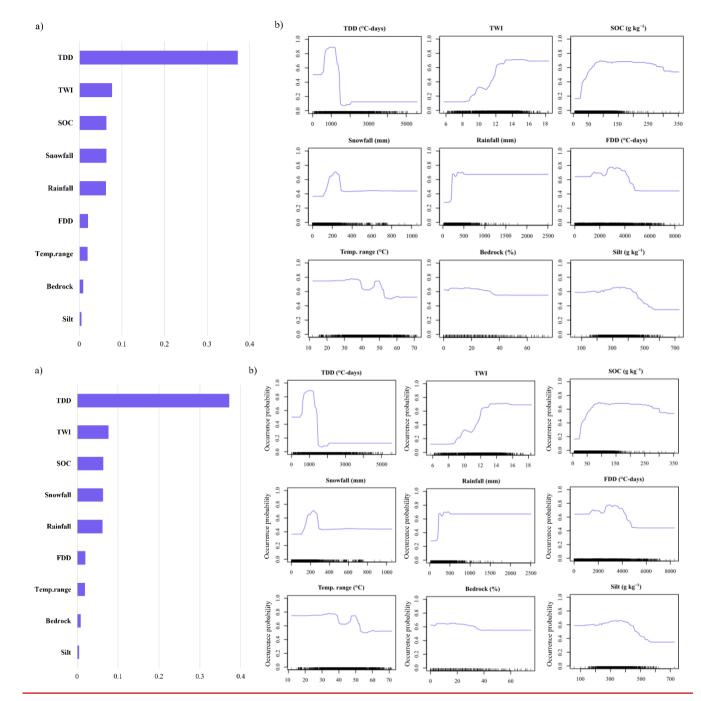


Figure 3: Variable importance values (a) and response curves (b) for the environmental variables based on the <u>random forest RF model</u>. Abbreviated variables are thawing- and freezing degree-days (TDD and FDD, °C-days), annual air temperature range (Temp.range, °C), topographic wetness index (TWI), soil organic carbon content (SOC, g kg⁻¹), silt content (Silt, g kg⁻¹) and bedrock probability within two meters (Bedrock, %).

3.2 Future changes in the circumpolar suitable environments

Our results suggest a dramatic loss of suitable environments for palsas and peat plateaus already by the mid of the 21st century (Table 1.; Table S2 in supplements). The high-emissions scenario RCP8.5 predicted even larger losses by the period 2041–2060, than thelow and moderate_emissions scenarios (RCP2.6 and RCP4.5) predicted by the second period 2061–2080. RCP8.5 scenario for 2061–2080 showed almost complete loss (-98,2 %) of recently suitable environments (Table 1e).

Table 1: Predicted Sguitable environmental spaces (in km²) for palsas and peat plateaus at different time periods and RCP climate change scenarios in Nordic countries (a), Western Siberia (b), Central and Eastern Siberia (c), North America (d), and entire circumpolar permafrost region (e). Areal percentage changes are given in relation to the modelled area for the period 1950–2000.

Results are based on random forest modelling. Three True Skill Statistic cut-off values of occurrence probability were used (an average cut-off based on 100 model iterations and average cut-off±1 standard deviation) to classify continuous occurrence probabilities into binary maps.

· · · · · · · · · · · · · · · · · · ·	Suitable area (km²)			Percentage change (%)		
	Average	<u>+1 s.d.</u>	<u>-1 s.d.</u>	Average	<u>+ 1 s.d.</u>	<u>- 1 s.d.</u>
a) Nordic countries						
<u>1950–2000</u>	<u>77 540</u>	<u>72 900</u>	<u>84 060</u>			
RCP2.6 2041–2060	<u>20 100</u>	<u>16 380</u>	<u>25 750</u>	<u>-74.1</u>	<u>-77.5</u>	<u>-69.4</u>
RCP2.6 2061–2080	<u>20 750</u>	<u>16 862</u>	<u>26 410</u>	<u>-73.2</u>	<u>-76.9</u>	<u>-68.6</u>
RCP4.5 2041–2060	<u>17 120</u>	<u>14 020</u>	<u>21 280</u>	<u>-77.9</u>	<u>-80.8</u>	<u>-74.7</u>
RCP4.5 2061–2080	<u>12 140</u>	<u>7 520</u>	<u>14 220</u>	<u>-84.3</u>	<u>-89.7</u>	<u>-83.1</u>
RCP8.5 2041–2060	<u>11 890</u>	<u>10 040</u>	<u>14 030</u>	<u>-84.7</u>	<u>-86.2</u>	<u>-83.3</u>
RCP8.5 2061–2080	<u>5 600</u>	<u>4470</u>	<u>6630</u>	<u>-92.8</u>	<u>-93.9</u>	<u>-92.1</u>
b) Western Siberia						
<u>1950–2000</u>	<u>562 750</u>	<u>537 930</u>	<u>598 390</u>			
RCP2.6 2041–2060	<u>78 930</u>	<u>58 790</u>	<u>89 150</u>	<u>-86.0</u>	<u>-89.1</u>	<u>-85.1</u>
RCP2.6 2061–2080	<u>67 190</u>	<u>54 740</u>	<u>72 450</u>	<u>-88.1</u>	<u>-89.8</u>	<u>-87.9</u>
RCP4.5 2041–2060	<u>48 270</u>	<u>36 600</u>	<u>53 100</u>	<u>-91.4</u>	<u>-93.2</u>	<u>-91.1</u>
RCP4.5 2061–2080	<u>14 930</u>	<u>7 280</u>	<u>18 860</u>	<u>-97.3</u>	<u>-98.6</u>	<u>-96.8</u>
RCP8.5 2041–2060	<u>10 110</u>	<u>1 940</u>	<u>15 252</u>	<u>-98.2</u>	<u>-99.6</u>	<u>-97.5</u>
RCP8.5 2061–2080	<u>10</u>	<u>8</u>	<u>33</u>	<u>-99.9</u>	<u>-99.9</u>	<u>-99.9</u>
c) Central and Eastern Siberia						
<u>1950–2000</u>	<u>305 580</u>	<u>226 520</u>	<u>415 610</u>			
RCP2.6 2041–2060	<u>91 360</u>	<u>72 360</u>	<u>109 880</u>	<u>-70.1</u>	<u>-68.1</u>	<u>-73.6</u>
RCP2.6 2061–2080	<u>80 290</u>	<u>65 940</u>	<u>97 410</u>	<u>-73.7</u>	<u>-70.9</u>	<u>-76.6</u>
RCP4.5 2041–2060	<u>62 550</u>	<u>50 260</u>	<u>78 110</u>	<u>-79.5</u>	<u>-77.8</u>	<u>-81.2</u>
RCP4.5 2061–2080	<u>30 350</u>	<u>25 330</u>	<u>39 050</u>	<u>-90.1</u>	<u>-88.8</u>	<u>-90.6</u>
RCP8.5 2041–2060	<u>25 140</u>	<u>19 710</u>	<u>34 270</u>	<u>-91.8</u>	<u>-91.3</u>	<u>-91.8</u>
RCP8.5 2061–2080	<u>1 070</u>	<u>420</u>	<u>3 270</u>	<u>-99.6</u>	<u>-99.8</u>	<u>-99.2</u>
d) North America						

<u>1950–2000</u>	<u>640 070</u>	<u>459 730</u>	913 453			
RCP2.6 2041–2060	225 830	<u>164 440</u>	<u>311 950</u>	<u>-64.7</u>	<u>-64.2</u>	<u>-65.8</u>
RCP2.6 2061–2080	207 690	146 350	<u>291 890</u>	<u>-67.6</u>	<u>-68.2</u>	<u>-68.1</u>
RCP4.5 2041–2060	<u>167 430</u>	<u>115 180</u>	<u>255 590</u>	<u>-73.8</u>	<u>-74.9</u>	<u>-72.0</u>
RCP4.5 2061–2080	111 450	<u>81 070</u>	<u>173 390</u>	<u>-82.6</u>	<u>-82.4</u>	<u>-81.0</u>
RCP8.5 2041–2060	104 310	<u>74 210</u>	<u>165 190</u>	<u>-83.7</u>	<u>-83.9</u>	<u>-81.9</u>
RCP8.5 2061–2080	<u>21 570</u>	<u>13 740</u>	43 480	<u>-96.6</u>	<u>-97.0</u>	<u>-95.2</u>
e) Permafrost region						
<u>1950–2000</u>	1 587 360	1 298 230	2 013 390			
RCP2.6 2041–2060	<u>416 910</u>	<u>312 610</u>	<u>537 560</u>	<u>-73.7</u>	<u>-75.9</u>	<u>-73.3</u>
RCP2.6 2061–2080	<u>376 650</u>	<u>284 540</u>	<u>488 980</u>	<u>-76.3</u>	<u>-78.1</u>	<u>-75.7</u>
RCP4.5 2041–2060	<u>296 110</u>	<u>216 610</u>	<u>408 830</u>	<u>-81.4</u>	<u>-83.3</u>	<u>-79.7</u>
RCP4.5 2061-2080	<u>169 460</u>	<u>124 440</u>	<u>246 150</u>	<u>-89.3</u>	<u>-90.4</u>	<u>-87.8</u>
RCP8.5 2041–2060	<u>151 990</u>	<u>106 300</u>	<u>229 190</u>	<u>-90.4</u>	<u>-91.8</u>	<u>-88.6</u>
RCP8.5 2061–2080	<u>28 590</u>	<u>18 860</u>	<u>53 790</u>	<u>-98.2</u>	<u>-95.9</u>	<u>-97.3</u>
'	Suita	able area (km²)		Percentage c	hange (%)	
a) Nordic countries						
1950 2000			77 540			
RCP4.5 2041 2060			17-120			-77.9
RCP4.5 2061 2080			12 140			-84.3
RCP8.5 2041 2060			11 890			-84.7
RCP8.5 2061 2080			5-600			-92.8
b) Western Siberia						
1950 2000		ŧ	562 750			
RCP4.5 2041 2060	48 270				-91.4	
RCP4.5 2061 2080	14 930				-97.3	
RCP8.5 2041 2060	10 110				-98.2	
RCP8.5 2061 2080			10			-99.9
c) Central and Eastern Siberia						
1950 2000		÷	305-580			
RCP4.5 2041 2060			62 550			-79.5
RCP4.5 2061 2080	30 350					-90.1
RCP8.5 2041 2060	25 140				-91.8	
RCP8.5 2061 2080			1 070			-99.6
d) North America						
1950-2000		•	640 070			
RCP4.5 2041 2060		-	167-430			-73.8
RCP4.5 2061 2080		-	111 450			-82.6
	•					

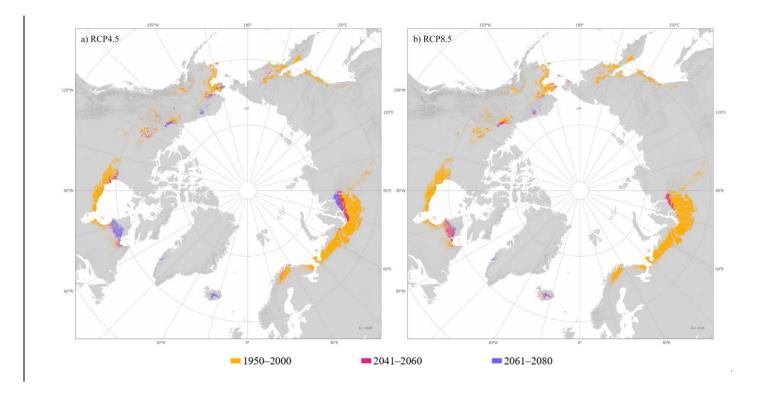
RCP8.5 2041 2060 RCP8.5 2061 2080	104 310 21 570	-83.7 -96.6
e) Permafrost region		
1950 2000	1 587 360	
RCP4.5 2041 2060	296 110	-81.4
RCP4.5 2061 2080	169 460	89.3
RCP8.5 2041 2060	151 990	-90. 4
RCP8.5 2061 2080	28 590	-98.2

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Remarkable reductions in suitable environments were predicted even under the low emissions scenario (RCP2.6) (Table 1; Fig. 4a). Reductions in suitable areas between the future periods in this scenario were minimal; difference in the average changes was 2.6 % for the whole permafrost region and varied from 0.9 to 3.6 % between regions. The predictions based on RCP2.6 scenario for 2061–2080 were more extensive than the RCP4.5 -predictions for the period 2041–2060. According to our results, tThe suitable environments for palsas and peat plateaus will disappear almost all around the Northern Hemisphere even under the moderate emissions climate change scenarios (Table 1; Fig. 4ba). The suitable environments would disappear almost completely from Northern Fennoscandia, West Siberia, coasts of Hudson Bay, western Alaska, and eastern Siberia already by the period 2041-2060. Only in Iceland and Greenland, the modelled palsa and peat plateau environments seemed to persist without major distributional changes. In addition to these regions, the suitable environments could persistremain in North America in the Northwest Territories, northern Quebec, and Alaska during the period 2061-2080. In Russia, palsas and peat plateaus could be foundpersist mainly infrom the Yamal Nenets Autonomous Okrug and Turukhansky district (Table 1b-c; Fig. 4ba). The predictions based on the RCP8.5 scenario for period 2041–2060 present similar distribution of suitable environmental spaces, compared with the RCP4.5 scenario for 2061–2080 (Fig. 4ab-cb). Only 1.8 % of the recently suitable environments would persist under RCP8.5 scenario in the period 2061–2080 (Table 1). However, predicted areas for the suitable environments are sensitive to the used occurrence probability cut-off values (Table <u>1).</u>



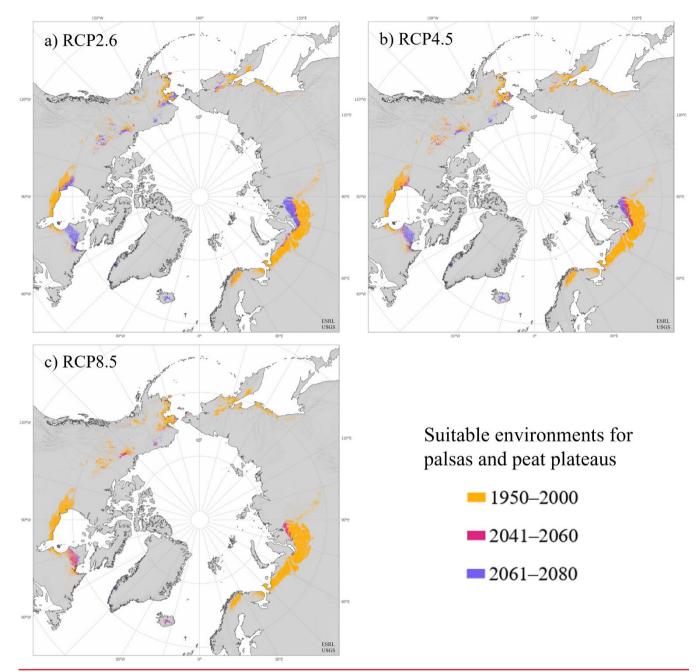


Figure 4: Predicted distributions of suitable environmental spaces for palsas and peat plateaus for different time periods under low (a), moderate (ba), and high emissions (cb) Representative Concentration Pathway climate change scenarios (RCP2.6, RCP4.5 and RCP8.5). Modelling results are presented for periods 1950–2000, 2041–2060 and 2061–2080. Random forest model results are provided for the permafrost region (Ran et al., 2022) and for the future periods for the extent of the period 1950–2000.

3.3 Thermokarst coverage of the suitable environments

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When our modelling results were compared against the thermokarst map (Olefeldt et al., 2016), areas with high or very high wetland thermokarst coverages were the first ones to become unsuitable based on our models (Fig. 4a–b & 5a). These regions located mainly at the coasts of the Hudson Bay, the Northwest Territories, west coast of Alaska and parts of Western Siberia. In turn, wide areas with none to low wetland thermokarst coverage were found from northern Quebec and Turukhansky district in Russia (Fig. 5a). These coincided well with the regions persisting as suitable environments in our climate change projections (Fig. 4a–b). The comparison with lake thermokarst presented comparable results for most of the regions with some exceptions in the Western Siberia (Fig. 5a–b).

The clearest conflict between our results and the thermokarst map located in northern Fennoscandia, where we predict major losses in suitable environmental spaces for permafrost peatlands. In contrast, Olefeldt et al. (2016) classified northern Fennoscandia to have no or low coverage of both wetland and lake thermokarst (Fig. 5a–b). Especially, lake thermokarst was rare in the area. Besides the northern Fennoscandia, the consistency between the predicted suitable environments and the thermokarst data was not so clear in the eastern parts of the Siberia.

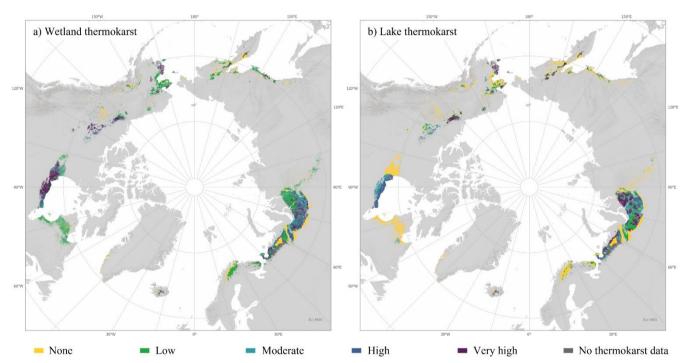


Figure 5: Suitable environmental spaces for palsas and peat plateaus (for period 1950–2000) against thermokarst classification by Olefeldt et al. (2016). Wetland (a) and lake (b) thermokarst coverages range from none (yellow) to very high (dark violet). Dark grey illustrates regions for which the thermokarst data are not available. Thermokarst coverages are classified as following, none (0–1 %), low (1–10 %), moderate (10–30 %), high (30–60 %) and very high (60–100 %). Results are provided for the permafrost region (Ran et al., 2022).

Overall, the wetland thermokarst had a closer spatial match with the projected lost and remaining suitable environmental spaces for palsas and peat plateaus in the circumpolar permafrost region. From the regions that our models predicted to become unsuitable for palsas and peat plateausthe landforms first (by 2040–2061 in RCP4.52.6 scenario), 40.7 36.5 we presented high or very high wetland thermokarst coverage, whereas 43.1 48.0 % had none or low coverage. This indicated slightly conflicting spatial relationship. However, when compared to the regions that were predicted to remain suitable during the period 2040–2061 in RCP2.6, RCP4.5 and RCP8.5 scenarios, a clear relationship was observed. These persisting palsa and peat plateau environments represented lower wetland thermokarst coverage than the degrading regions (Fig. S6).

3.4 Statistical and spatial evaluation of the models

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The 100-fold cross-validation based on the split-sample approach yielded high AUC and TSS values (Fig. 6), indicating very good predictive performance of the models. RF and GBM best predicted the distribution of suitable environments, and the ensemble model performed slightly weaker. Between-model variability within the 100 modelling runs was estimated by the standard deviations. For AUC the standard deviation range was 0.003–0.006, and for TSS 0.014–0.019. Evaluation with the separate dataset resulted in lower AUC and TSS values than the cross-validation, but still indicated excellent predictive performance based on AUC values (AUC > 0.9) (Fig. 6). All models were also good (TSS > 0.5) in predicting the occurrence of landforms. Inwith the separate evaluation data, RF and GBM had the highest TSS values, whereas the ensemble and RF models had the best AUC score. Compared to the cross-validation, larger between-model variability was observed in TSS values (sd. 0.017–0.04), whereas variability in AUC values (sd. 0.001–0.004) was smaller. GLM and GAM were the weakest models in both evaluations. Based on the evaluation metrics, response curves, and predictive maps RF was considered the best method for predicting suitable environments for palsas and peat plateaus.

The total area for which all four models predicted occurrence of palsas, and peat plateaus was 0.8 million km². This represents approximately 28 % of the area for which at least one model predicted occurrence of the landforms (3.1 million km²), indicating relatively low model agreement. The highest model agreement (Fig. S7) was found in Northern Fennoscandia, Western Siberia and Northern Canada around the Hudson Bay and Quebec. Also, west coast of Alaska, regions around Kamchatka Peninsula and Iceland had a high model agreement.

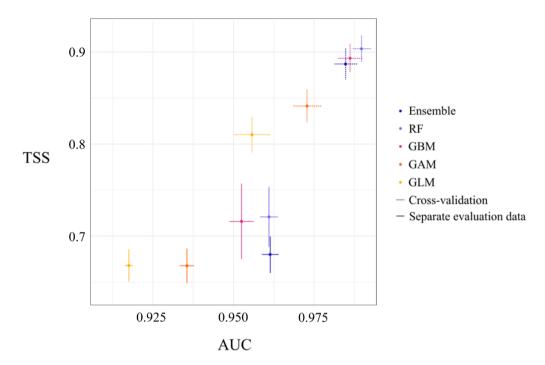


Figure 6: Statistical evaluation metrics for different modelling techniques. The averages of AUC and TSS values with one standard deviation (whiskers) are based on 100 runs of the models. Evaluation results for the 100-fold cross validation with the calibration data are presented with dashed line and for the separate evaluation data, results are presented with solid line.

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The comparison between our predictions and different peatland datasets show that the predicted suitable environments are more extensive than previously assessed permafrost peatland extent in Northern Fennoscandia, Iceland, Labrador Peninsula and Seward Peninsula (Fig. 7). Comparison with the PEATMAP also indicate overestimations in Western Siberia and in Eastern Siberia (Fig. 7c). Possible underestimations are found across the Canadian Shield also, where, however, the underestimations mostly occurred beyond the strict climate space (Fig. 7). According to the suitable climate spaces and used peatland datasets the suitable peatlands for palsas and peat plateaus are limited in Central Siberia (Fig. 7). Most of the region's permafrost peatlands locate north of the suitable climate space.

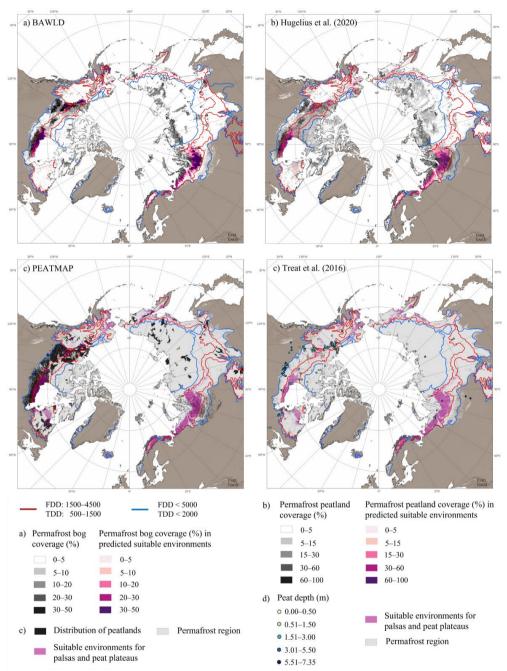


Figure 7: Comparison between the predicted suitable environmental spaces (for period 1950–2000), suitable climate spaces, and peatland datasets; (a) Permafrost bog coverage percent (0–50 %), in each grid cell according to the Boreal–Arctic Wetland and Lake Dataset (BAWLD; Olefeldt et al., 2021), (b) permafrost peatland coverage percent (0–100 %) (Hugelius et al., 2020), (c) distribution of peatlands according to PEATMAP (Xu et al., 2018), and (d) peat depth according to Treat et al. (2016b) in point symbols (yellow to blue scale). Permafrost regions are illustrated in light gray (Ran et al., 2022) and regions outside permafrost region with brown. The suitable environmental spaces are illustrated with pinkish colors, whereas areas outside our binary predictions are illustrated with grey scale. Two classification of suitable climate spaces are illustrated with red (a strict interpretation) and blue (a liberal interpretation) lines.

The total area for which all four models predicted occurrence of palsas, and peat plateaus was 0.8 million km². This represents approximately 28 % of the area for which at least one model predicted occurrence of the landforms (3.1 million km²), indicating relatively low model agreement. The highest model agreement (Fig. S7) was found in Northern Fennoscandia, Western Siberia and Northern Canada around the Hudson Bay and Quebec. Also, west coast of Alaska, regions around Kamehatka Peninsula and Iceland had a high model agreement.

4 Discussion

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Our models predicted drastic loss of suitable environments for palsas and peat plateaus, indicating almost complete loss of the landforms by 2080 according to the moderate and high emissions scenarios (RCP4.5 and RCP 8.5). Previous modelling studies have shown similar trends for restricted study areas (e.g., Aalto et al., 2017; Fewster et al., 2022) and many monitoring studies have documented palsas and peat plateaus already degrading around the Northern Hemisphere (Payette et al., 2004; Borge et al., 2017; Mamet et al., 2017; Olvmo et al., 2020). Here, we presented predicted the disappearance of suitable environments hese degradation trends for the whole northern circumpolar permafrost region for the first time. The shrinkage of suitable environments continued in all used emissions scenarios (Table 1). This might indicate that even the strong climate change mitigation is not enough to slow the degradation of suitable palsa and peat plateau environments down. Such major changes in the permafrost peatlands as projected here, may have an influence on the future carbon cycling and potential to further accelerate the climate change (Turetsky et al., 2020; Miner et al., 2022). Moreover, the predicted permafrost degradation may have a major impact on both geodiversity and biodiversity of current Northern Hemisphere permafrost region. Thus, comprehensive understanding of the current distribution of studied landforms is needed.

Based on the recent period, the suitable environments for palsas and peat plateaus—were predicted to occur around the Northern Hemisphere permafrost region, with hotspots at the most studied regions (i.e., Northern Fennoscandia, Western Siberia, and subarctic Canada). Our results from the Northern Fennoscandia and Western Siberia coincide well with the results of Fewster et al. (2022). Especially, suitable environments of the Northern Finland, Norway, and Sweden agreed well with the previous mappings and modellings (Fronzek et al., 2006; Backe, 2014; Metsähallitus, 2019; Fewster et al., 2022). In contrast, our models predicted more extensive palsa and peat plateau occurrences in European Russia (e.g., Kola Peninsula and Komi Republic) than expected, and RF did not predict landforms to occur as far north in Yamal-Nenets than Fewster et al. (2022). The comparison with the peatland datasets (Fig. 7) indicated more extensive suitable environments especially in the Northern Fennoscandia, but this might have been caused by the coarse resolution of some of the peatland datasets (Fig. 7a–b), which did not allow for reproducing small-scale variation in sporadic permafrost in the region.

For previously unmapped regions of Central and Eastern Siberia oOur model extrapolationsmodels predicted suitable environments in the east coast of Russia and inside a zone stretching southeast from the Western Siberia. Our RF models did not predict palsas or peat plateaus in the central parts of Siberia (Fig. 2). This might indicate too dry and continental climate, or otherwise unsuitable environmental conditions conditions favoring other periglacial landforms, such as pingos and polygon mires (French, 2017), or too thin peat layers (Seppälä, 1988; Hugelius et al., 2020). Indeed, the comparison between peatland datasets and suitable climate spaces, showed that most of the peatlands of Central and Eastern Siberia are located north of the estimated suitable climate spaces (Fig. 7). Whatsoever, our models likely failed to recognize suitability of some of the relatively few peatlands inside the suitable climate space. It is possible that the limited distribution of suitable environments is caused or might have been caused by insufficient number of presence observations to characterize the possibly suitable environmental conditions for palsas and peat plateaus in the region (Fig. 1). However, based on the stricter suitable climate space the region for these landforms is narrow, indicating that our extrapolations likely did not underestimate the distribution of suitable environmental spaces as much as one could expect based on the limited number of landform observations in the region.

For North America our results aligned well with Fewster et al. (2020) around the Hudson Bay, Quebec, the Northwest Territories, and Yukon. However, our results indicated more variability spatially resolved in landform occurrence especially in the Northwest Territories and Yukon, showing e.g., a gap in suitable environments east of the Great Slave Lake, and clear differences in suitable environments of Alaska. For Alaska, our models predicted occurrence of landforms suitable environments in the coastal regions whereas Fewster et al. (2020) predicted them in more central parts. The contrasting results may be partly explained by the inclusion of non-climatic variables to our models, and the substantially finer spatial resolution of our study, which both can facilitate the differentiation of environmental suitability at an improved accuracy. Comparison with peatland datasets show possible overprediction of suitable environments in North Slope and Seward Peninsula, for example. In Canada the most striking overprediction occurred in the Labrador Peninsula. However, a recent study from the region showed that previous studies have underestimated distribution of peatland permafrost along the coast of the peninsula (Wang et al., 2022b).

The optimal climatological conditions coincide quite well with the climate envelopes determined by previous studies for Fennoscandia (e.g., Luoto et al., 2004b; Aalto and Luoto, 2014) and Northern America (Fewster et al., 2020). For the Central and Eastern Siberia no optimal climatic ranges for palsas and peat plateaus have been determined before. TDD was the most important variable while FDD had only a minor importance in our models, indicating that the summer conditions thawing season conditions during the thawing season have a greater effect on the distribution of suitable environments for palsas and peat plateaus. Recent studies have similarly highlighted the importance of thawing season conditions for the occurrence of permafrost (Mekonnen et al., 2021), other periglacial landforms (Karjalainen et al., 2020), and active layer (Peng et al., 2018; Karjalainen et al., 2019).

The negative effect of higher air temperatures on the distribution of palsas and peat plateaus can be observed from the response curve for TDD (Fig. 3a). Moreover, increasing rainfall can cause permafrost to thaw even faster than it would only because of the temperature rise (Magnússon et al., 2022). According to these results, warmer and wetter summers might be most influential in accelerating the degradation of palsas and peat plateaus and decreasing the extent of the suitable environments in our future scenarios. In addition, thicker snowpack can also lead to the permafrost thaw in the discontinuous permafrost region (Biskaborn et al., 2019) as the thick snowpack acts as an insulator preventing the frost from penetrating deep into the soil (Seppälä, 1990, 1994; Ge and Gong, 2010; Sannel, 2020). Increasing snowpack also leads to moister conditions during late spring due to abundant meltwaters (Sannel, 2020), and thus alters the thermal properties of peat (Kujala et al., 2008). Indeed, the thin snow cover has been tested in the field to be a key factor in the formation of palsas (Seppälä, 1982), and the negative effect of increasing snowfall (Johansson et al., 2013) can be recognized from our results (Fig. 3a). However, there might be a delay before the increasing snow depth will affect to the thaw depth (Sannel et al., 2016).

Other environmental conditions beside the climate <u>also</u> affect also the distribution of peat dominated permafrost mounds and other periglacial landforms (Seppälä, 2011; Karjalainen et al., 2020). Our models indicate that palsas and peat plateaus require sufficient SOC content, which is logical, as the sufficient peat cover is considered crucial for the formation and occurrence of these landforms at the marginal permafrost regions (Kujala et al., 2008; Seppälä, 2011). Most of the previous modelling studies of palsas and peat plateaus (e.g., Fronzek et al., 2006; Parviainen and Luoto, 2007; Fewster et al., 2020, 2022) have used only climate envelope models. Our results show that incorporating other environmental variables may enhance the predictive performance of the models as SOC and TWI had the highest VI values after TDD (Fig. 3). In addition, compared to previous broad-scale studies (Fewster et al., 2022, 2020) the used high spatial resolution allows for identifying unsuitable areas within the climatically suitable envelopes and thereby reduces the risk of overestimating <u>suitable</u> environments <u>palsa and peat plateau distribution</u>.

Even though our models predicted major losses in the suitable environments, the peat cover of the landforms may cause time delay in the degradation process, as dry peat acts as an effective insulator for permafrost cores of the mounds (Kujala et al., 2008). Thus, the actual degradation of palsas and peat plateaus might happen later than our statistical models predict. The documented areal degradation rates varying from 0.5–1 % a⁻¹ (e.g., Borge et al., 2017; Mamet et al., 2017), however, support the rapid degradation of the landforms. Moreover, not all palsas and peat plateaus are equal when comparing the rate of degradation. Landforms which are already imbalanced with the current climate (e.g., Olvmo et al., 2020) will degrade faster than landforms in equilibrium with climate (e.g., Tam et al. 2014). For these reasons we preferred to discuss about the distribution of suitable environments rather than claiming that our results would represent the actual distribution and degradation of the studied landforms. Although Despite the possible delayed or varying degradation rate between regions the

timing of the degradation might be delayed, our predictions can be used to estimate the forthcoming spatial changes in the distribution of the suitable environments.

It is important to note that most of palsas and peat plateaus were formed during the Little Ice Age (the 16th to mid-19th century period), in colder climates than we are currently facing (Halsey et al., 1995; Treat et al., 2016a; Vorren, 2017; Fewster et al., 2020). Because of this not all palsas and peat plateaus are in equilibrium with the current climate anymore. We did not model the historical distribution of suitable environments by using climate data from the landform initiation period, because it would have required historical data describing the other environmental factors also. Especially, the peat initiation data would have been necessary, and the compilation of the data for the whole pan-Arctic region was beyond the scope of this study. However, we acknowledged the role of past climates by choosing the 1950–2000 to be our baseline period as it was considered to better resemble the conditions of the period of palsa and peat plateau formation than the current climatology (e.g., 1991–2020; Seppälä, 2003; Saemundsson et al., 2012).

Previous studies have suggested that the development of thermokarst can be used as an indicator of the former distribution of palsas (Matthews et al., 1997; Luoto and Seppälä, 2003). Thermokarst ponds are relatively common in Fennoscandian palsa mires (e.g., Luoto and Seppälä, 2003), although Olefeldt et al. (2016) classified this region to have a low thermokarst coverage. This mismatch was noticed also here, as we predicted extensive degradation of the landforms in the region. Overall, our results showed that regions with higher thermokarst coverage are in a higher risk to become unsuitable environments for palsas and peat plateaus, compared to the regions with lower thermokarst coverage. This indicated clear spatial relationship with our results and the thermokarst coverage. Recognition of the regions with a high degradation risk is useful in the estimation of future greenhouse gas fluxes from permafrost wetlands (Swindles et al., 2015; Miner et al., 2022) and establishing conservation actions for these endangered ecosystems and habitats (Janssen et al., 2016).

Owing to the extensive study area with heterogeneous environmental conditions our aim was to compile as representative landform observation dataset as possible for model calibration. Due to this, no fully independent landform datasets were available for the evaluation of our models. This was however compensated in multiple ways. First, we divided the compiled landform observations into semi-independent calibration and evaluation datasets and performed evaluation both by the 100-fold cross-validation (with calibration dataset) and with the separate evaluation dataset (Fig. 6). Second, our predictions were compared with four peatland datasets (Treat et al., 2016b; Xu et al., 2018; Hugelius et al., 2020; Olefeldt et al., 2021) and suitable climate spaces to assess our predictions for the recent period (1950–2000) (Fig. 7). Third, our future predictions were compared to the independent thermokarst data (Olefeldt et al., 2016), to examine whether the degrading regions coincide with high thermokarst coverage. All these evaluations supported our predictions even in the regions with low number of landform observations. We recognize that our results for Central and Eastern Siberia are extrapolations of models fitted elsewhere as the landform observations from the region are scarce and this must be considered when interpreting the results. In addition, it is important to acknowledge the uncertainties related to statistical modelling itself (see e.g., Heikkinen

et al., 2006; Hjort and Luoto, 2013). To overcome these methodological caveats, we used multiple modelling methods and quantified the prediction uncertainty.

To further develop the statistical models for the prediction of <u>suitable environmental spaces for palsas</u> and peat plateaus distribution we would need more spatially resolved datasets describing the thickness of the snowpack and peat layer instead of the indirect snowfall and SOC variables. <u>Incorporating variables describing, e.g., average wind conditions and vegetation structure could improve the performance of the models, but their applicability is hampered by the lack of suitable high-resolution data depicting future conditions. Although our study gave new insights into the distribution of <u>suitable environments</u> for palsas and peat plateaus in Central and Eastern Siberia, future research should focus on acquiring better</u>

knowledge of the current distribution of palsas and peat plateausthe landforms in these regions.

5 Conclusions

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From the results of this study, we conclude that:

- 35 Based on a recent period, sSuitable environments for palsas and peat plateaus can be found across the Northern Hemisphere, with occurrence hotspots in Northern Europe, Western Siberia and around the Hudson Bay, Quebec, and the Northwest Territories of Canada.
- A dramatic loss of the suitable environments for palsas and peat plateaus is predicted to occur already by 2041–2060 and almost complete loss by 2061–2080.
 - Under <u>alow and moderate emissions scenarios</u> (<u>RCP2.6 and RCP4.5</u>), landforms can persist in the coldest parts of the recent distribution area, but if the climate change mitigation fails (RCP8.5), suitable environments are predicted to be lost from almost the entire Northern Hemisphere.
- Further research on the distribution of the landforms is needed especially in Central and Eastern Siberia, for more reliably assessing their current state and potential future development.
 - The pronounced importance of thawing-season climate conditions for circumpolar palsa and peat plateau occurrence suggests that the projected increases in summer temperatures and rainfall may strongly affect the stability of permafrost peatland landforms.
- In addition to the climatic variables, soil organic carbon, and moisture accumulation potential of the soil affect the landform occurrence at circumpolar scale, and they need to be considered in order to draw a detailed picture of the landforms' distribution.
 - Projected loss of the suitable palsa and peat plateau environments overlapped with the regions having a high thermokarst coverage. In the future, it is likely to find increasing amount of thermokarst from recent palsa and peat plateau environments.

The degradation of permafrost peatlands will potentially have an influence on the diversity of subarctic nature and the carbon balance of Earth. Thus, the predicted changes in the suitable environmental spaces for palsas and peat plateaus should be taken into consideration when the estimating the pace and impacts of the climate change over northern regions.

Appendix A

List of references used in the compilation of palsa and peat plateau data.

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Data availability

Used climate data were obtained from WorldClim – Global Climatic Data, available at https://worldclim.org/ (Hijmans et al., 2005). Variables describing the soil properties were calculated from data in SoilGrids – global gridded soil information database https://files.isric.org/soilgrids (Poggio et al., 2021). Global Multi-resolution Terrain Elevation Data 2010 (GMTED2010; Danielson and Gesch, 2011) https://doi.org/10.5066/F7J38R2N were used to calculate TWI. Thermokarst dataset (Olefeldt et al., 2016) can be obtained from https://daac.ornl.gov/cgi-bin/dsviewer.pl?ds_id=1332. BAWLD - database (Olefeldt et al., 2021) is available at https://arcticdata.io/catalog/view/doi:10.18739/A2C824F9X, Hugelius et al. (2020) data at Bolin Centre for Climate (https://doi.org/10.17043/hugelius-2020-peatland-2), PEATMAP (Xu et al., 2018) at https://doi.org/10.5518/252, and Treat et al. (2016b) data at https://doi.org/10.1594/PANGAEA.863689. The coordinates of the Llandform observations compiled in this study are available in the supplementary materials from the corresponding author upon reasonable request. List of references used in the compilation is provided in Appendix Aappendices.

Author contribution

OHK, OK and JH conceptualized the research idea. OHK collected the observational data and complied it with the environmental data. OK led the compilation of the environmental datasets and their geospatial processing with OHK and JA. OHK performed the statistical analysis with contribution from OK and JH. OHK wrote the paper with contributions from all the authors.

Disclaimer

785 The authors declare that they have no conflict of interests.

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