Temporal variation <u>of bacterial community and nutrients</u> in <u>Tibetan</u> glacier snowpack-bacterial communities mediated by nitrogen

Yuying Chen^{1,4}, Keshao Liu^{1,4}, Yongqin Liu^{1,2,4}, Trista J. Vick-Majors³, Feng Wang^{1,4}, Mukan Ji²

¹State Key Laboratory of Tibetan Plateau Earth System, Resources and Environment (TPESRE), Institute of Tibetan Plateau Research, Chinese Academy of Sciences, Beijing 100101, China

²Center for the Pan-third Pole Environment, Lanzhou University, Lanzhou 730000, China ³Department of Biological Sciences, Great Lakes Research Center, Michigan Technological University, Houghton, Michigan 49931, United States

⁴University of Chinese Academy of Sciences, Beijing 100049, China

10

5

Correspondence to: Mukan Ji (jimk@lzu.edu.cn)

Abstract. Global warming accelerates glacier meltmelting across the globe, releasing stored carbon and nitrogen, which fertilizefertilizes downstream ecosystems. DiverseSeveral studies have investigated the seasonal dynamics of nutrients and active-microbial communities mediate biogeochemical cycles in supraglacial snow and are vital to the glacial ecosystem.

- 15 However, but little is known about their temporal changing pattern and the environmental and biotic determinantschanges in snowpacks.fresh snow from a single snowfall. Here, we investigated used Illumina high throughput sequencing of 16S rRNA gene to investigate the bacterial community in the surface and subsurface snow (depth at 0-15 and 15-30 cm, respectively) during a nine-day period immediately following a snowfall in the Dunde Glacier of the Tibetan Plateau, based on Illumina MiSeq of 16S rRNA gene sequences. Our results revealed dynamic-rapid temporal changes in nitrogen (including nitrate and
- 20 <u>ammonium</u>) and bacterial communities in both surface and <u>surface snow</u>, and <u>nitrogen is the key determinant of bacterial diversity</u>, <u>composition</u>, <u>community structure</u>, and <u>biotic interactions.subsurface snow</u>. Nitrate and ammonium <u>concentrationconcentrations</u> increased from 0.44 to 1.15 mg/L and 0.18 to 0.24 mg/L in the <u>surface snow</u> and decreased from <u>3.81 to 1.04 mg/L and 0.53 to 0.25 mg/L</u> in the <u>surface and</u> subsurface snow over time, therefore indicating accumulation and consumption processes, respectively. This is also evidenced by The nitrate concentration covaried with bacterial diversity.
- 25 <u>community structure, and the dominance of organisms</u>-predicted to carry-nitrogen fixation and denitrification-genes in the surface and subsurface layers, respectively.-related genes, suggesting nitrogen could mediate bacterial community changes. The nitrogen limitation and the apparent dominance of the<u>enriched</u> denitrification-related genes in the subsurface snow suggestsuggested stronger environmental and biotic filtering than those in the-surface snow. This was associated with, which may explain the lower bacterial diversity, more pronounced community temporal changes, and stronger biotic interactions than
- 30 in the surface snow. Collectively, these findings significantly advanced advance our understanding of microbial bacterial community variations and bacterial interactions after snow deposition, and revealed the dynamics of provide a possible biological explanation for nitrogen metabolism dynamics in Tibetan snow.

1 Introduction

- 35 Global warming accelerates glacier melting across the globe, and supraglacial snow is particularly vulnerable (Hodson et al., 2008) with the nutrientscarbon and nitrogen stored in glaciers arebeing released into downstream ecosystems in melt waters (Hodson et al., 2005; meltwaters (Wadham et al., 2019).; Hodson et al., 2005). The composition and abundance of nutrients in glaciers, which are typically poor in inorganic nutrients relative to aquatic and soil environments (Ren et al., 2019)supraglacial snows are regulated by glacier-dwelling microorganisms (Hodson et al., 2008). A range of metabolically active bacteria have
- 40 been reported in <u>supraglacial</u> snow<u>environments</u> including Bacteroidetes, Actinobacteria, Firmicutes, and Alphaproteobacteria (Miteva, 2008; Maccario et al., 2019; Carey et al., 2016; Lazzaro et al., 2015; Michaud et al., 2014). These microorganisms perform key ecological functions in biogeochemical <u>eyelescycling</u> such as carbon and nitrogen fixation, which are vital to <u>the</u> nutrient-limited <u>glacial ecosystems.supraglacial ecosystem</u>. Changes in their community composition and activities <u>are expected towill</u> influence the dynamics of <u>glacial</u>-nutrient- storage, transformation, and release. Thus, it is
- 45 crucial to understand how the <u>microbialbacterial</u> community in <u>glacialsupraglacial</u> snow changes across time and to determine whether those changes are associated with the temporal nutrient <u>differences in snow</u>, to fully estimate the ecological consequences of global glacier meltingdynamics in snow.

During precipitation events, a new snow layer forms above the previous ones, which creates a stratified structure. Each snow layer has unique physical and chemical characteristics (Lazzaro et al., 2015), which may trigger distinct post-depositional

- 50 selective processes on microbial communities (Xiang et al., 2009; Møller et al., 2013). Several studies investigated the dynamics of nutrient and bacterial changes in supraglacial snow during the ablation period. Larose et al. (2013a) revealed that the form of nitrogen varied as a function of time in supraglacial snow during a two-month field study at the Svalbard, Norway and fluctuations in microbial community structure have been reported with the relative abundance of fungi and bacteria (such as Bacteroidetes and Proteobacteria) increased and decreased, relatively. Seasonal shifts in snowpack bacterial communities
- 55 have also been reported in the mountain snow in Japan, rapid microbial growth was observed with increasing snow temperature and meltwater content (Segawa et al., 2005). However, the results of these studies are likely the consequence of these several precipitation events due to the long period of time. During precipitation, a new snow layer forms above the previous ones, which is responsible for the stratified snowpack structure. These different snow layers have distinct physical and chemical characteristics and their age also differed substantially (Lazzaro et al., 2015). Thus, the microbial process across the aged
- 60 snowpack could be complex, whereas focusing on supraglacial snow from a single snowfall event could provide unique insights into the bacterial and nutrient dynamics. Hell et al. (2013) reported bacterial community structure changes during the ablation period across five days in the high Arctic, while the bacterial and nutrient dynamics during the snow accumulation period remain elusive.

Surface and subsurface typically have distinct bacterial community structures due to the environmental filtering from the

vertical profile of temperature, solar radiation intensity, and nutrients (Xiang et al., 2009; Møller et al., 2013; Carey et al., 2016). For example, Cyanobacteria tend to dominate upper snow layers (0-15 cm) (Carey et al., 2016), while their relative

abundance is greatly reduced in the deeper snow layer (Xiang et al., 2009). This is likely due to <u>the</u> lower light intensity in the deeper snow, <u>andwhich</u> favors heterotrophic bacteria such as the Actinobacteria and Firmicutes (Carey et al., 2016). Thus, differences in the snow physicochemical conditions shape their distinct bacterial community structures. However, whether

- 70 microorganisms in the different snowpack layers exhibit similar responses to environmental selection is still largely unknown.
 Differences in physicochemical conditions can also indirectly influence microbialbacterial community structure through impacts on the types of biotic interactions that dominate an environment (Friedman and Gore, 2017; Khan et al., 2018; Bergk Pinto et al., 2019). On Spitsbergen Island in Svalbard forFor example, the addition of organic carbon shifted microbialbacterial interactions from collaboration to competition in Arctic snow (Bergk Pinto et al., 2019). In comparison, intensive collaboration
- can enhance complex organic carbon degradation and mineralization, which are particularly important for oligotrophic environments such as glaciers (Bergk Pinto et al., 2019; Krug et al., 2020). Collaboration is also known to be essential to biological processes such as ammonia oxidation and denitrification, in which various organisms carry out different steps of these processes (Henry et al., 2005; Madsen, 2011; Yuan et al., 2021). These changes in interactions and network complexity can favor or disadvantage certain microbialbacterial groups, thereby changing the microbialbacterial community structure (i.e., 80 biofiltering).

<u>The</u> Tibetan Plateau is the world's third-largest ice reservoir, after those in Antarctica and Greenland (Qiu, 2012). It is warming at a rate twice of the global average (Chen et al., 2015), <u>causing rapid shrinkageand 95%</u> of <u>the Tibetan</u> glaciers and <u>snowretreated between 1990 to 2005</u> (Rauscher et al., 2007; Hall and Fagre, 2003). The glacier-: Yao et al., 2007). Glacier melting <u>leads to increases</u> the <u>enhanced</u> discharge of microorganisms and nutrients <u>in meltwater</u> into downstream aquatic and

- 85 terrestrial ecosystems, (Kohler et al., 2020), which makes an impact on their substantially impacts the bacterial community and biogeochemical processes, (Liu et al., 2021). Thus, it is crucial to understand the transformation processes of microbial<u>the</u> bacterial community and nutrients in the supraglacial snow. Several studies have investigated the nutrient and bacterial community changes in supraglacial snow, across the winter (Brooks et al., 1998; Liu et al., 2006), but the bacterial and nutrient dynamics of freshly fallen snow have been largely overlooked. These short temporal changes will influence the following post-
- 90 depositional processes after it is buried by the next snowfall, and will ultimately determine the physicochemical properties of the stratified snow in the following year. In the present study, we investigated the bacterial community composition-and interactionsnow physiochemical property changes in the surface and subsurface supraglacial snow-layers during a nine-day period after a single snowfall event at the Dunde Glacier on the northeast of the Tibetan Plateau. Our sampling strategy focused on the ablation zone of the glacier, because it is in this region that mass lost exceeds mass gained. Therefore, ablation zone
- 95 microorganisms and the biogeochemical processes they are associated with are expected to have a greater impact on downstream processes, relative to that of the glacier accumulation zone. In particular, we<u>We</u> aimed to answer the following key questions: 1) do the bacterial community and nutrient changes in a short temporal scale, 2) do the bacterial communities in different snow layers exhibit similar community temporal changes, and 2<u>3</u>) are <u>the</u> temporal changes within each layer<u>in</u> the surface and subsurface snow related to environmental filtering, biotic interactions, or both?

100 2 Materials and methods

2.1 Site description and sample collection

Snow samples were collected from the ablation zone at Dunde glacier (38°06'N, 96°24'E, 5325 m above the sea-level), during the-October and November, 2016 (Supplementary Fig. S1). Dunde glacier is located in the Qilian mountain region on the northeastern Tibetan Plateau, and it is continuously monitored by the Institute of Tibetan Plateau Research, Chinese Academic

- 105 of Sciences. No supraglacial snow was observed on the glacier surface on the 10th of October when first arrived at the camp. Snowfall started on the 18th and ended on the 23rd of October. Sampling was conducted over a nine-day period (after the snowfall stopped on a flat 5 m × 3 m small area to reduce the impact of sample heterogeneity due to spatial variations. Snow samples were collected on the 24th, 25th, 26th, 27th, and 29th of October, and the 2nd November, (which are referred as day 1, 2, 3, 4, 5, 6, and 9 thereafter). No snow was observed on the glacier surface during the summer, so the snowpack observed during
- 110 October and November represented fresh snow accumulated during the autumn months.) until the next snowfall started. This enabled us to follow the development of microbialbacterial communities and the chemical environment through time. <u>after</u> deposition. The ambient air temperature at the sampling period is averaged -8 °C (data available through the European Centre for Medium-Range Weather Forecasts, Supplementary Fig. S2), no snow melting was observed over the nine days.
- On each day, three snow pits were <u>randomly</u> dug within <u>athe</u> 5 m \times 3 m area and any two snow pits were <u>at least 30-50</u> cm apart. Each snow pit was approximately 30 cm deep, and the bottom of <u>then</u> the snow <u>pit was a layer of ice</u>, <u>indicating that the</u> snow pits encompassed the depth of the existing snowpack. Each snow pit-was further divided <u>equally</u> into <u>the</u> surface and subsurface layers (approximately 15 cm deep for each layer) based on the softness and heterogeneity of the snowpack<u>to get</u> <u>enough snow samples to extract DNA</u>, after Carey <u>et al.</u> (2016). For each snow pit, the top 1 cm in contact with the air was removed using a sterile spoon to avoid contamination, and then surface and subsurface snow were collected using a sterilized
- 120 Teflon shovel into 3L separate, 3 L sterile sampling bags. An additional 120 separately. Approximately 100 mL (meltwater volume) of surface and subsurface snow was sampled were used for physiochemical physicochemical analyses, whereas the rest was used for DNA extraction. A total of 36 samples were collected. Tyvek bodysuits and latex gloves were worn during the entire sampling process to minimize the potential for contamination, and gloves were worn during all subsequent handling of samples. Samples were kept frozen during the transportation to the laboratory and stored at -20 °C until analysis.

125 2.2 Environmental characterization of snow

130

The 100 mL snow sample for physicochemical analysis was melted at room temperature for 3 hours before being analysed. For dissolved organic carbon (DOC), total nitrogen (TN),) and major ions measurements, 100 mL of snow meltwater was syringe-filtered through a 0.45 μ m cellulosepolytetrafluoroethylene (PTFE) membrane (Millipore, Billerica, MA, USA<u>filter</u> (Macherey–Nagel) into 20-mL glass bottles-(. The membrane has been pre-treated with 1% HCl-leached, deionized water rinsed, and 450 °C > 3 h combusted). to remove any potential carbon and nitrogen on the membrane, and the initial 10 mL of the filtrate was discarded before collecting the sample for analysis to eliminate any residual compound on the membrane. The DOC-and-TN concentrations were measured with a TOC-VCPH analyzer (Shimadzu Corp., Japan). Major ions (NH₄⁺, NO₃⁻, Na^+ , K^+ , and SO_4^{2-}) were analyzed using a Thermo-Fisher ion chromatography system 900 as described previously (Rice et al., 2012). The precision and accuracy of the TOC-VCPH analyzer were both < 3% and the limit of detection was 0.05 mg L⁻¹.

The precision and accuracy of the ion chromatography system 900 were < 5% and 0.1 mg L⁻¹, and the limit of detection was 135 0.01 mg L⁻¹. The concentrations of TN, NH₄⁺, and NO₃⁻ ions were quantified using different methods. The ion chromatography method for quantifying NH4⁺ and NO₃⁻ ions is susceptible to interference from carbonate and bicarbonate ions in samples (Novic et al., 1997), this may lead to an overestimation of nitrate ion concentration, which does not allow direct concentration comparison between TN and NO_3^{-1} ions. Nevertheless, the standard curves were linear with r > 0.99, thus allows comparison among different time points (Supplementary Fig S2 (Supplementary Fig. S3). 140

2.3 DNA extraction

For assessing the bacterial community composition, each of the melted snow samples (3 L) was were melted at 4 °C overnight and filtered onto a sterile 0.22 µm polycarbonate membrane (Millipore, USA) with a vacuum pump (Ntengwe 2005). Bacterial community DNA was extracted from the biomass retained in respective filters using a Fast DNA®SPIN Kit for Soil (MP

145 Biomedicals, Santa Ana, CA, USA) according to the manufacturer's instructions. DNA extraction with no sample added were was performed in parallel, and the final elute was used as a negative control.

The raw DNA was checked by electrophoresis in 1% (w/v) agarose gel, and purified from the gel using an Agarose Gel DNA purification kit (TaKaRa, Japan). The concentration and purity of the DNA extracts were measured using a NanoDrop 1000 spectrophotometer (Thermo-Scientific, Wilmington, DE, USA). The extracted DNA was stored at -80 °C until amplification.

150

2.4 Bacterial 16S rRNA amplification and Illumina MiSeq sequencing

In total, 36 DNA samples and one negative control were subjected to amplicon sequencing. Universal primers 515F (5'-GTGCCAGCMGCCGCGGTAA-3') and 806R (5'-GGACTACHVGGGTWTCTAAT-3') (Caporaso et al., 2012) with 12 nt unique barcodes were used to amplify the V4 hyper-variable regions of the bacterial 16S rRNA gene. Polymerase chain 155 reaction (PCR) was performed under the following conditions: 94°C for 5 minutes, 30 cycles of 94°C for 30 seconds, 52°C for 30 seconds, 72°C for 30 seconds; followed by a final cycle of 10 minutes at 72°C. Each PCR reaction contained 12.5 µL 2x Premix Tag DNA polymerase (Takara Biotechnology, Dalian Co. Ltd., China), 1 µL each primer (0.4 µM final concentration), and 8.5 μ L nuclease-free water, 2 μ L DNA template (20 ng μ L⁻¹) or 2 μ L sterile water for the PCR negative controls. PCR products were confirmed using agarose gel electrophoresis, and no PCR band was detected in PCR negative 160 controls. To minimise PCR batch-to-batch variations and maximise the quantity of PCR product, triplicate PCR reactions were performed for each sample, and PCR products were pooled for purification using the OMEGA Gel Extraction Kit (Omega Bio-Tek, Norcross, GA, USA) following electrophoresis. PCR products from different samples were pooled in equal molar amounts, and then used for 2×250 bp paired-ends sequencing on a MiSeq machine (Illumina, San Diego, CA).

2.5 Processing of Illumina sequencing data

- 165 MiSeq sequence data were processed using the QIIME 2 pipeline version 2018.8 (Bolyen et al., 2018) following the recommended tutorials (https://docs.qiime2.org/2018.8/tutorials/) and using the plugin demux to visualize interactive quality diagrams and check read quality. Plugin DADA2 (Callahan et al., 2016) was applied to remove primers, truncate poor-quality bases, conduct dereplication, identify chimeras, and merge paired-end reads. Commands included in the feature table (McDonald et al., 2012) generated the summary statistics of sequences related to the samples. Further, we trained a Naïve
- 170 Bayes Classifier with the feature-classifier plugin using the 16S rRNA gene database at 99% similarity of the SILVA 132 QIIME release and based on the 515F/806R primer pair as used for the PCR. Finally, the taxa plugin was used to filter mitochondrial and chloroplast sequences, as well as to generate absolute read count tables of all taxa for each sample. Data were analyzed at the level of amplicon sequence variant (ASV), where ASVs are delineated by 100% sequence identity (Callahan et al., 2017).
- After removing singletons, a total of 1,685,186 high-quality reads were obtained, representing 9178 ASVs. Before statistical analysis, the dataset was rarefied to 45,000 reads per sample, which is the lowest read count among samples. Rarefaction curves reached an asymptote before the subsampling, which confirmed that this depth was sufficient to detect the diversity present (Supplementary Fig-S3, S4).

2.6 Network analysis

- 180 The ASV-ASV associations within the surface and subsurface bacterial communities were explored using Molecular Ecological Network Analyses Pipeline (http://129.15.40.240/mena/) (Deng et al., 2012). The ASVs that occurred in at least 50% of the samples from the surface or subsurface group were selected to construct the network. Spearman's rank correlation coefficient (ρ) was calculated to reflect the strength of association between species. The false discovery rates (Q-values) were calculated from the observed P-value distribution. The resulting correlation matrix was analyzed with the Random Matrix
- 185 Theory (RMT)-based network approach to determine the correlation threshold for network construction, and the same threshold was used for both the surface and subsurface network, so the topological properties of the surface and subsurface networks are comparable.

2.7 Statistical analysis

Shannon-Wiener and Chao1 indices, which were used to estimate the species richness in the snow community, were calculated using the <u>"diversity</u>" function in the R package <u>"vegan</u>" (Oksanen et al., 2010). Functional profiling of bacterial taxa was carried out using the package "*Tax4Fun2*" in R (Wemheuer et al., 2020). While the application of functional profiles predicted from 16S rRNA gene based community composition data is limited by the functional information available in databases, we present these data as one possible interpretation of the patterns we detected, and note that the "*Tax4Fun2*" package performed well compared to older widely used programs (Wemheuer et al., 2020). Tax4Fun2 has been widely used in the literatures and

- 195 the results has been been shown to correlate with the actual metagenomic datasets well (Wemheuer et al., 2020). However, its accuracy is limited by the functional information available in databases. In the present study, nitrogen cycling-related genes were predicted, and the prediction results were used to provide a possible explanation for the nitrate and ammonium changing pattern observed. The pairwise Wilcoxon rank-sum test was used to compare the depth-horizon differences in environmental variables, alpha-diversity, and the relative abundance of taxonomic groups at the phylum level, and the relative abundance of
- putative functional genes. Linear regression models were used modelling was implemented in R using the "lm" function to 200estimate the trend of changes over time. The bacterial community structure was subjected to principal coordinate analysis (PCoA) carried out using the "pcoa" function of the "ape" package in R. Principal component analysis (PCA) was performed in R using the "prcomp" function. The significance of dissimilarity of community composition among samples was tested using permutational multivariate analysis of variance (PERMANOVA) based on Bray-Curtis distance metrics with the "adonis"
- 205 function in the R package ""vegan²" (Oksanen et al., 2010). Test results with P < 0.05 were considered statistically significant. The correlations between bacterial diversity and snow variables were calculated using Spearman's rank correlation, as implemented in the 'ppcor' R package. Mantel test based on Spearman's rank correlations was performed using the bacterial dissimilarity and environmental dissimilarity matrix, calculated based on the Bray-Curtis distance metrics and Euclidean distance metrics in the <u>"vegan</u>" R package, respectively. The normalized stochasticity ratio (NST) based on the Bray–Curtis

210dissimilarity was calculated using the "NST" package in R to estimate the determinacy and stochasticity of the bacterial assembly processes with high accuracy and precision (Ning et al., 2019). The NST index used 50% as the boundary point between more deterministic (<50%) and more stochastic (>50%) assembly processes. All environmental variables were normalized before the calculation. All statistical analyses were executed in R version 3.4.3 (R Core Team, 2017).

3 Results

215 **3.1 Environmental characteristics of the snowpack**

The concentrations of TN, NO₃, and NH₄⁺ ranged from 0.18 to 1.26 mg L⁺¹, 0.44 to 5.09 mg L⁻¹, and 0.17 to 0.62 mg L⁻¹, respectively (Fig. 1a, Supplementary Table S1). The concentrations of TN, NO $_{2,7}$, and NH₄+they were both significantly higher in the subsurface layer than in the surface layer snow (Wilcoxon rank-sum test; all P < 0.001, Fig. 1a). K⁺ and SO₄²⁻ ions in the subsurface layersnow were also significantly higher (-0.29 ± 0.13 and 6.09 ± 3.18 mg L⁻¹, respectively) than those in the surface have show (0.12 \pm 0.08 and 3.71 \pm 1.64 mg L⁻¹, respectively; Wilcoxon rank-sum test; P < 0.001, and P = 0.015, respectively). 220 The concentrations of DOC ranged from 0.46 to 5.89 mg L^{-1} and exhibited no significant difference in the surface and subsurface layers now (Wilcoxon rank-sum test; P = 0.310). The concentrations of Na⁺ ion ranged from 0.35 to 7.34 mg L⁻¹. which also exhibited no significant difference in the surface and subsurface layers snow (Wilcoxon rank-sum test; P = 0.079). The concentration of NO₃, and NH₄⁺ ions in the surface layer significantly snow increased with time ($F_{1,16} = 5.97$, P = 0.027, $R^2 = 0.27$ and $F_{1.16} = 8.58$, P = 0.010, $R^2 = 0.35$, respectively, Fig. 1b). In comparison, the concentrations of TN, NO₃, and 225 NH_4^+ in the subsurface layer significantly they decreased with time- in the subsurface snow ($F_{1,16} = 40.66$, P < 0.001, $R^2 = 0.72$ and $F_{1,16} = 50.74$, P < 0.001, $R^2 = 0.76$, respectively). Other environmental factors showed exhibited no significant eorrelation changes with time-in-either layer.

3.2 Diversity and composition of bacterial community from the snowpack

- 230 The bacterial Shannon and Chao1 indices in the surface layer were 5.61 ± 0.39 and 744 ± 199, respectively, which were not significantly different from those in the subsurface layer (5.52 ± 0.68 and 705 ± 269, respectively) (Fig 2a). In the surface layer, the Shannon and Chao1 indices did not change significantly with time (Pearson correlation; r = 0.15, P = 0.553; r = 0.02, P = 0.939, respectively, Fig 2b). In contrast, both the Shannon and Chao1 indices both significantly decreased with time in the subsurface layer (Pearson correlation; r = 0.63, P = 0.003; r = 0.56, P = 0.009, respectively). In the surface layer, the
- 235 Shannon indices positively correlated with the concentration of DOC⁻(Pearson correlation; r= 0.48, P = 0.04, Fig-3a).-In the subsurface layer, the Shannon and Chao1 indices were positively correlated with the concentration of TN, NO₃⁻ and NH₄⁺ (Pearson correlation; P < 0.05, Fig 3b and Supplementary Fig S4).

The surface and subsurface snow layers-were both dominated by Alphaproteobacteria $(25\%)_{7_2}$ Actinobacteria $(20\%)_{7_2}$ Cyanobacteria $(15\%)_{7_2}$ Gammaproteobacteria $(15\%)_{7_2}$ Bacteroidetes $(11\%)_{7_2}$ Firmicutes $(4\%)_{7_2}$ Chloroflexi $(2\%)_{7_2}$ Gemmatimonadetes $(2\%)_{7_2}$ Planctomycetes $(1\%)_{7_2}$ Acidobacteria $(1\%)_{7_2}$ Deltaproteobacteria $(1\%)_{7_2}$ and Deinococcus-Thermus (1%) (Fig-4. 2). The relative abundance of most of these phyla was not significantly differed in the two snow layers, except the Gemmatimonadetes, Planctomycetes, and Acidobacteria, which exhibited significantly higher relative abundance in the surface layer than in the subsurface layer (all P < 0.05, Wilcoxon rank-sum test; Supplementary Fig-S5). In the surface layer, negative associations were apparent in the relative abundances and ASV number of Alphaproteobacteria,

- 245 <u>Gammaproteobacteria, and Firmicutes with time ($F_{1,16} = 6.97$, P = 0.018, $R^2 = 0.30$; $F_{1,16} = 23.8$, P < 0.001, $R^2 = 0.60$, and $F_{1,16} = 22.28$, P < 0.001, $R^2 = 0.58$ in relative abundance; $F_{1,16} = 7.56$, P = 0.014, $R^2 = 0.32$; $F_{1,16} = 27.12$, P < 0.001, $R^2 = 0.63$, and $F_{1,16} = 16.68$, P = 0.001, $R^2 = 0.51$ in ASV number, respectively), while positive associations were apparent in the relative abundances and ASV number of Cyanobacteria and Deinococcus-Thermus with time ($F_{1,16} = 6.94$, P = 0.018, $R^2 = 0.30$ and $F_{1,16} = 13.10$, P = 0.002, $R^2 = 0.45$ in relative abundance; $F_{1,16} = 3.42$, P = 0.083, $R^2 = 0.18$ and $F_{1,16} = 4.07$, P = 0.061, $R^2 = 0.20$ </u>
- 250 in ASV number, respectively; Supplementary Fig. S6). In the subsurface layer, negative associations were apparent in the relative abundance and ASV number of Alphaproteobacteria and Firmicutes with time ($F_{1,16} = 15.17$, P = 0.001, $R^2 = 0.49$ and $F_{1,16} = 15.43$, P = 0.001, $R^2 = 0.49$ in relative abundance; $F_{1,16} = 18.98$, P = 0.083, $R^2 = 0.54$ and $F_{1,16} = 15.17$, P = 0.001, $R^2 = 0.001$, $R^2 = 0.001$,
- 255 <u>P = 0.003, R² = 0.44 in relative abundance; F_{1,16} = 5.34, P = 0.034, R² = 0.25 and F_{1,16} = 14.49, P = 0.002, R² = 0.47 in ASV number, respectively).</u>

The bacterial Shannon and Chao1 indices in the surface snow were 5.61 ± 0.39 and 744 ± 199 , respectively, which were not significantly different from those in the subsurface layer (5.52 ± 0.68 and 705 ± 269 , respectively) (P = 0.81 and 0.57, respectively) (Fig. 3a). In the surface layer, the relative abundances of Alphaproteobacteria, Gammaproteobacteria, and

- Firmicutes significantly decreased with time (Supplementary Fig S6), while those of Cyanobacteria and Deinococcus-Thermus significantly increased (all *P* < 0.05). In the subsurface layer, the relative abundance of Alphaproteobacteria and Firmicutes significantly decreased with time, while Cyanobacteria and Chloroflexi significantly increased (all *P* < 0.05). In the surface snow, the Shannon and Chao1 indices were similar across the nine days (F_{1.16} = 0.37, *P* = 0.553, R² = 0.02 and F_{1.16} = 0.01, *P* = 0.939, R²= 0.001, respectively; Fig. 3b). In comparison, negative associations were observed in both Shannon and Chao1 indices with time in the subsurface snow (F_{1.16} = 12.33, *P* = 0.003, R² = 0.44 and F_{1.16} = 8.73, *P* = 0.009, R² = 0.35, respectively). In the surface layer, the positive correlations of Shannon and Chao1 indices with the DOC and sodium ions were apparent (F_{1.16} = 4.90, *P* = 0.042, R² = 0.23 and F_{1.16} = 4.91, *P* = 0.042, R² = 0.24, respectively; Fig. 4a,b). In the subsurface snow, the positive correlations of Shannon and Chao1 indices with the concentrations of NO₃² and NH₄⁺ were apparent (Shannon diversity: F_{1.16} = 9.13, *P* = 0.008, R² = 0.36 and F_{1.16} = 5.17, *P* = 0.037, R² = 0.24, respectively; Chao1 index: F_{1.16} = 8.60, *P* = 0.009, R² = 0.36 and F_{1.16} = 5.32, *P* = 0.035, R² = 0.25, respectively; Fig. 4cd). This is consistent with the random forest
- analysis results, which identified the concentrations of NO_{3} and NH_{4} as the significant determinants of bacterial Shannon diversity in the subsurface layer (Supplementary Fig. S8).

3.3 Bacterial community structure and functional genes

- The bacterial community structure <u>at the ASV level</u> significantly differed in the surface and subsurface snow (PERMANOVA, F = 2.78, *P* < 0.001, Fig. 5a), as well as among the different sampling times (PERMANOVA, F = 3.31, *P* < 0.001; and F = 2.17, *P* < 0.001, respectively). Additionally, a significant interactive effect was detected between the depth and time (PERMANOVA, F = 2.68, *P* < 0.001), indicating that the depth influenced the temporal pattern of bacterial community structure changes. Specifically, only the second principal coordinate (PCoA2) values of the surface layersnow significantly correlatedvaried with time (Pearson correlation; $r=0.94F_{1,16} = 141.8$, *P* < 0.001, $R^2 = 0.89$, Fig. 5b), while the PCoA1 values of the surface layersnow did not. Furthermore, PCoA1 and PCoA2 of the surface snow exhibited no significant correlation with the measured environmental factors (Supplementary Fig. S9 and S10). In comparison, both PCoA1 and PCoA2 values of the subsurface layer correlatedsnow co-varied with time significantly (Pearson correlation; r=0.53($F_{1.16} = 6.35$, *P* = 0.02; r_{023} , $R^2 = 0.5828$ and $F_{1.16} = 8.38$, *P* = 0.04011, $R^2 = 0.34$, respectively); Fig. 5b), while the PCoA2 also demonstrated significant association with nitrate, ammonium, potassium, sulfate, and DOC concentrations (Supplementary Fig. S10).
- 285 Normalized stochasticity ratio (NST) was used to examine the relative contributions of stochasticity and determinism in shaping bacterial communities. The average NST values were 74% and 46% in the surface and subsurface snow layers, and the contribution of stochasticity was significantly higher in the surface than in the subsurface layers (*P* < 0.001; Supplementary Fig-S7. S11).</p>
- Mantel tests were performed to evaluate the effects of environmental <u>parametersfactors</u> on bacterial community 290 <u>compositionstructure</u> within each layer. No significant correlation was identified between the measured environmental factors and the bacterial community structure in the surface <u>layersnow</u>. However, <u>significant positive correlations were apparent in</u> <u>the subsurface snow with the concentrations of TN. NO₃⁻⁻⁻ and NH₄⁺ were significantly correlated with bacterial community</u>

composition in the subsurface layer (P < 0.05 = 0.005 and 0.01, respectively) (Table 1). The relative abundance of nitrogencycling associated functional genes werewas predicted in the surface and subsurface layers now. The relative abundance of

- 295 nitrogen-fixation marker gene (*nifH*) significantly increasedpositively associated with time in the surface layer, butwhile no clear pattern was observed in the subsurface layer ($\underline{F}_{1,16} = 7.76$, $P \leq 0.05013$, $R^2 = 0.33$ and $\underline{F}_{1,16} = 0.57$, P = 0.46461, $R^2 = 0.01$, respectively, Supplementary Fig-S8. S12). The relative abundance of *narG* gene, which is associated with involved in the denitrification process, reduced significantly exhibited negative and positive associations with time in the surface layer with time, while increased significantly in the and subsurface layer (, respectively ($\underline{F}_{1,16} = 4.69$, $P \leq 0.05046$, $R^2 = 0.23$ and $\underline{F}_{1,16} = 0.57$).
- 300 = 11.24, $P \le 0.01004$, $R^2 = 0.41$, respectively). The *nirK* gene, which is also associated with<u>involved in</u> the denitrification process, reduced significantly with time in the surface layer, while no significant change was identified in the subsurface layer ($\underline{F}_{1,16} = 10.39$, $P \le 0.01005$, $R^2 = 0.39$ and $\underline{F}_{1,16} = 1.98$, P = 0.18179, $R^2 = 0.05$, respectively).

3.4 Interspecies interactions at the surface and subsurface layers

- Co-occurrence networks were constructed for the surface and subsurface bacterial communities to infer the biotic interactions among species (Fig. 6). The surface network comprised a higher number of nodes (each indicating one ASV, n==197) but a lower number of edges (each indicating a significant correlation between two ASVs, n==436) than the subsurface network (n==140 and 523, respectively, Table 2). The network in the subsurface snow demonstrated a higher number of edges per node (3.73 and 2.21, respectively), higher average connectivity (avgK, 7.57 and 4.43, respectively), and lower average path distance (GD, 4.72 and 5.51, respectively), which indicates a substantially more complex network topology. Both networks
- 310 were dominated by positive (co-presence) relationships, and the subsurface network exhibited a higher positive-to-total interaction ratio (95%) than the surface network (83%).

Modularity, average clustering coefficient (avgCC), and graph density of the surface and subsurface bacterial community networks were all higher than those of random networks (Supplementary Table S2), indicating that snowpack bacterial networks showed non-randomly assemblage and exhibited modular structures. The subsurface networks showed higher values

of avgCC (0.39), transitivity (0.49), and connectedness (0.86) than the surface bacterial community network (0.31, 0.45, 0.71, respectively), indicating a greater degree of connectivity (Table 2).

4 Discussion

4.14.1 Rapid shifts of bacterial community structure across a short temporal scale

- 320 The surface and subsurface snow were both dominated by Alphaproteobacteria, Actinobacteria, Cyanobacteria, Gammaproteobacteria, and Bacteroidetes (Fig. 2). Despite differences in sampling season, the bacterial taxa detected were consistent with previous studies on snow in the Arctic and Antarctic (Larose et al., 2010; Carpenter et al., 2000; Amato et al., 2007; Lopatina et al., 2013; Møller et al., 2013). Bacterial richness and diversity remained consistent throughout the nine days in the surface snow layer, while they exhibited a reduction trend in the subsurface snow layer (Fig. 3b). This indicates that the
- 325 <u>microbiome in the subsurface snow may be subjected to greater environmental filtering than those in the surface snow (Xiang et al., 2009).</u> Among all environmental factors measured, nitrate and ammonium were the only measured environmental factors that changed across the nine days. The nitrate and ammonium concentrations both exhibited an R² value of greater than 0.7 and reduced with time, therefore indicating a consumption process (Fig. 1b). Despite the R² value being weak, both nitrate and ammonium concentrations covaried with bacteria richness and diversity in subsurface snow, which is not observed in the
- 330 surface snow (Fig. 4). Furthermore, random forest analysis also identified nitrate and ammonium to be the dominant driver of bacteria Shannon diversity in the subsurface layer (Supplementary Fig. S8). Thus, these results suggest that nitrate and ammonium could play a more important role in influencing bacterial diversity in subsurface snow than that in surface snow. Nitrogen is an essential nutrient for microbial growth and plays important role in controlling microbial diversity and ecosystem productivity (Vitousek et al., 2002; Xia et al., 2008; Sun et al., 2014). The positive associations between nitrogen concentration
- and alpha diversity indices have been typically inferred as nitrogen limitation (Telling et al., 2011). Thus, these results hint that nitrogen limitation could occur in subsurface snow and influence bacteria diversity. In comparison, the surface layer is unlikely to be subjected to nitrogen-limitation as evidenced by the lack of association between nitrogen and bacterial diversity (Telling et al., 2012). This is consistent with previous studies in the Greenland ice sheet, where nitrate additions to surface ice did not alter the cryoconite community cell abundance and 16S rRNA gene-based community composition (Cameron et al., 2017).

2017). The bacterial community structure also exhibited significant temporal changes in the subsurface layer and nitrogen was the most important explaining factor (Table 1 and Fig. 5), again indicating greater environmental filtering (Kim et al., 2016). This is consistent with the finding in the Arctic that nitrogen influences surface snow bacterial community composition via regulating algae metabolism (Lutz et al., 2017). This is also consistent with the higher modelled contribution of deterministic

- 345 processes relative to stochastic processes in the subsurface layer than the surface layer (Supplementary Fig. S11). Deterministic processes could be due to environmental filtering or biotic interactions, whereas stochastic processes include dispersal limitation, community drift, and speciation (Stegen et al., 2012). The surface layer could obtain nitrogen through aeolian deposition processes (Bjorkman et al., 2013), whereas the subsurface snow could only receive limited external microbial and nutrient input through supraglacial meltwater. The latter could be particularly limited during the glacier deposition period when
- 350 the glacier surface temperature is below zero degrees (Fig. S2).

Our results suggest that both bacteria and snow physiochemical properties experience rapid changes across the nine days during the snow deposition period in the Tibetan glacier investigated here. Traditionally, supraglacial snow is recognized as a cold oligotrophic environment with a very slow metabolism rate (Quesada and Vincent, 2012; Marshall and Chalmers, 1997), but increasing evidence has suggested that bacterial community changes can occur on a short temporal scale. For example,

- 355 Hell et al. (2013) reported changes in the dominant bacterial phylum Proteobacteria across five days and active bacterial metabolism has been observed in the Greenland Ice Sheet supraglacial ice (Nicholes et al., 2019). In addition, active bacteria affiliated with Proteobacteria have been identified in the Antarctic (Lopatina et al., 2013) and Arctic (Holland et al., 2020) snow at temperatures below zero degrees, therefore supporting the present study that bacterial community changes in nine days could be possible. This indicates that supraglacial snow can harbour an active bacterial community and interact with the
- 360 nutrient transformation process.

4.2 Distinct nitrogen-transformation processes in surface and subsurface snow

Both ammonium and nitrate concentrations 1b)-Nitrogen concentration significantly increased in the surface snow, which can 365 be due to (Fig. 1). The increase in ammonium is traditionally explained by biogenic emissions due to local vegetal and animal sources (Filippa et al., 2010), while the increase in nitrate has been largely attributed to atmospheric deposition and/or microbial nitrogen fixation activity. (Björkman et al., 2014). Nitrogen deposition occurs at a rate of 282 kg N km⁻² yr⁻¹ in the region of our investigation (Lü and Tian, 2007), making it a possible source (Björkman et al., 2014). On the other hand, microbial nitrogen fixation can support supraglacial microbial communities when this equals 0.19 mg N for the 0.5 m \times 0.5 m area 370 sampled each day (assuming nitrogen deposition occurred evenly across the year). If further assuming the deposited nitrogen only affects the surface snow (i.e., the top 15 cm as defined in the present study), the daily nitrogen increase is estimated to be $0.084 \text{ mg N L}^{-1}$. This is lower than the slope of total nitrogen increase observed in the surface snow of the present study (0.21 mg N L⁻¹ day⁻¹). Thus, either the atmospheric nitrogen deposition has more than doubled, or bacterial nitrogen fixation could be an alternative sourcessource of nitrogen become limited input (Telling et al., 2011). In the current study, a potential 375 contribution from bacterial nitrogen fixation The latter is supported by the significant biosynthesis of nitrogen-containing compounds by bacteria with increased dissolved organic nitrogen reported in the Antarctic surface snow (Antony et al., 2017). The contribution of bacterial nitrogen fixation is further supported by the increase in the relative abundance of Cyanobacteria and the predicted abundance of nifH gene in surface snow (Supplementary Fig. S6 and Fig. S12). The exact nitrogen fixation rate was 58). We did not measurequantified in the rates of present study, but the results suggest that microbial nitrogen fixation 380 in our system, however, even in a study focused on cryoconite holes, which are expected to have higher rates of biological activity than snowpack, the measured rate of could be an overlooked source of nitrogen in Tibetan glacier snow, further transcriptomic and nitrogen-isotope analyses may provide further evidence on the microbial activity in nitrogen fixation (average 0.04 kg N km⁻² yr⁻¹) were orders of magnitude lower than the nitrogen input from precipitation in the same region (Telling et al., 2011). Therefore, we expect atmospheric deposition to be the dominant driver of the elevated nitrogen level 385 observed in the surface snow.

In contrast with the surface layer, nitrogen concentrations (total N, nitrate, and ammonium) significantly decreased in the subsurface <u>snow</u> with time (Fig. 1). This indicates distinct nitrogen transformation processes, where the surface was associated with nitrogen accumulation, and the subsurface was dominated by nitrogen consumption. We propose that the anaerobic <u>denitrification 1</u>). In a snow reactive nitrogen oxides (NO_v) survey in Greenland, NO_v flux was reported to exit snow in 52 out

390 of 112 measurements, and the magnitude cannot be explained by the photolysis of nitrate alone (Dibb et al., 1998). Furthermore, the short sampling period of the present study does not allow rapid photolysis to occur (Larose et al., 2013b), therefore collectively suggesting an alternative source of NO_y emission could exist. The denitrification process could contribute to nitrogen consumption, which is evidenced by the significant-increase of predicted genes associated with denitrification processes (*narG*; Supplementary Fig. S8), which encodes nitrate reductase. S12) (Telling et al., 2011; Zhang et al., 2020).

- 395 The This is consistent with the high relative abundance of denitrification-related genes being detected in the snowpack of Spitsbergen Island of Svalbard, Norway (Larose et al., 2013a). Despite the oxygen level in the subsurface snow was not measured, the occurrence of <u>anaerobic</u> denitrification reactions in subsurface snow has been reported in Arctic snowpacks, where the genes encoding all the required enzymes for denitrification were detected (Larose et al., 2013). The different nitrogen transformation processes in the surface and subsurface layers could be attributed to the substantial differences in geophysical
- 400 conditions, such as the lower light intensity (Xiang et al., 2009) and (Larose et al., 2013a). Furthermore, Poniecka et al. (2018) showed that cryoconite microorganisms can generate an anoxic zone 2 mm below the sediment surface within an hour. Thus, anaerobic conditions (Larose et al., 2013)pockets in the subsurface snow layer.at 15-30 cm deep could exist, which allows denitrification reactions to occur. Further-metagenomic and metatranscriptomic analyses targeting the genes associated with nitrogen cycling are required to further confirm the distinct nitrogen transformation processes between the surface and 405 subsurface layers.

4.2 Nitrogen drives the overall bacterial community structure changes in snow

Bacterial richness and diversity exhibited distinct temporal patterns. Alpha diversity indices remained consistent throughout the nine day period in the surface snow layer, while they reduced significantly in the subsurface snow layer. This indicates 410 that the microbiome in the subsurface layer may be subjected to greater environmental filtering than in the surface layer (Xiang et al., 2009), and, combined with our chemical data, indicates a relationship with nitrogen. This is consistent with the strong and positive correlative relationships between nitrogen and bacterial diversity in the subsurface layer, which was not observed in the surface layer (Fig. 3). Nitrogen is an essential nutrient for microbial growth and plays important roles-in controlling microbial diversity and ecosystem productivity (Vitousek et al., 2002; Xia et al., 2008; Sun et al., 2014). The positive 415 correlations between nitrogen availability and alpha diversity indices in the subsurface layer suggest that nitrogen limitation is an important determinant of bacterial diversity in glacier snowpack as burial removes the influence of surface deposition. Similar conclusions have been reported in subglacial pore waters and sediments (Ren et al., 2019). In comparison, the surface layer is not subjected to nitrogen limitation, which is consistent with previous studies in supraglacial ecosystems (Cameron et al., 2017; Holland et al., 2020). The bacterial community also exhibited significant temporal changes in the subsurface layer 420 that were most closely correlated with nitrogen concentration (Table 1). Consistent with the alpha diversity indices, the bacterial community in the subsurface layer exhibited stronger temporal changes (Fig. 2), again indicating greater environmental filtering. This was also consistent with the higher modelled contribution from deterministic processes relative to stochastic processes in the subsurface layer relative to the surface layer (Supplementary Fig S7). Owing to the accumulation of the surface snow, the external nutrient inputs to the subsurface layer are limited, as is the deposition of microorganisms 425 dispersed via aeolian processes. Together, these explain the stronger signal from environment filtering due to nitrogen limitation (Stegen et al., 2012). Despite the microbiome in the surface layer being subjected to less nutrient limitation and

environmental filtering, the PCoA2 of the surface layer was strongly correlated with time (Fig. 5b). This suggests that while the richness and diversity did not exhibit any significant temporal changes, the bacterial community structure in the surface layer still changed, potentially r unmeasured environmental factors (such as UV radiation) (Maccario et al., 2014).

430 **4.3** <u>Subsurface snow</u> <u>exhibits greater complexity in Nitrogen transformation processes mediate</u> biotic interactions in snow

Biotic interactions can explain a substantial proportion of the community structure variations (Hacquard et al., 2015; Dang and Lovell, 2016). Our results indicated that the subsurface community network was more complex as evidenced by the higher average connectivity and a shorter path length (GD) than the surface community network (Table 2). This is likely due to the

435 enhanced environmental filtering, as has been observed in other systems subjected to environmental stresses_(Ji et al., 2019; Wang et al., 2018). A higher ratio of positive-to-total interactions but lower modularity was identified in the subsurface snow network (Table 2). In general, higher positive interactions indicate increased microbial cooperation (Ju et al., 2014; Scheffer et al., 2012), whereas reduction in modularity indicates microbial niche-homogenization (Ji et al., 2019). The enhanced biotic associations and cooperation in the subsurface layer may be attributed to the dominanceoccurrence of denitrification processes,

- 440 as denitrification is a multi-step process that involves multiple microbial bacterial cohorts to complete (Henry et al., 2005; Madsen, 2011; Yuan et al., 2021). The enhanced collaboration and deterministic succession had been reported in microbialbacterial community associated with the anoxic decomposition of microcystis biomass (Wu et al., 20192020), and cross-feeding leads to enhanced positive interactions among the different members of the community (Borchert et al., 2021).
- The path lengths of the subsurface network were lower than that of the surface layer (Table 2). The shorter path length has 445 been proposed to be associated with a highertransfer efficiencies higher transfer efficiency of information and materials across the microorganisms in the network (Du et al., 2020), which are required for complex biological processes that require extensive microbialbacterial collaboration, such as denitrification (Yuan et al., 2021). Thus, the short path length is consistent with the dominance of denitrification processes in the subsurface layer. Previous studies have proposed microbial interactions as biotic drivers that impact microbial diversity (Calcagno et al., 2017; Hunt and Ward, 2015). Thus, those microorganisms who are not
- adapted to the subsurface environment would be excluded from the environment, which provides an alternative explanation for the reduction in diversity (Scheffer et al., 2012; Ziegler et al., 2018; Bergk Pinto et al., 2019).

5 Conclusion

475

- 455 Our results demonstrated the key rolesdynamics of nitrogen in shaping the and bacterial community in Tibetan glacialsupraglacial snow- over nine days. The surface and subsurface snow isare associated with the accumulation and consumption of nitrogen, respectively. Due to dryatmospheric nitrogen deposition and microbialbacterial nitrogen fixation activities, nitrogen limitation is unlikely to occur in the surface snow, thus additional nitrogen deposition due to global climate change is unlikely to cause a substantial impact on the bacterial community in surface snow. In summary, our results provide
- 460 a new perspective on the bacterial communities in snowpacksubstantially impact the bacterial community in surface snow. In contrast, nitrogen consumption was inferred in the subsurface snow. Nitrogen is traditionally recognized to be released from supraglacial environments due to photolysis, whereas the present study hints that the bacterial denitrification process could be an alternative route. Therefore, the increased nitrogen deposition due to anthropogenic activities may enhance the denitrification process in the subsurface snow. The enhanced nitrogen emission could reduce the impact of increased nitrogen
- 465 deposition on downstream glacier-fed rivers, but may feedback global warming positively. In summary, our results provide a new perspective on the dynamics of nutrients and bacterial community in supraglacial snow of the Tibetan Plateau, and further studies based on metagenome and metatranscriptome can enhance the understanding of microbialbacterial functions-in snow and predict their future changes.
- 470 *Data availability*. Sequence data generated in the present study have been deposited to the National Center for Biotechnology Information (NCBI) Sequence Read Archive under the ID PRJNA649151.

Author contributions. YL and MJ conceived the study and developed the idea. YC performed DNA extraction. YC and FW performed the environmental characterization measure. YC conducted the data statistical analysis. YC and KS wrote the first draft of the paper, and MJ, TV, and YL revised the paper substantially. All authors read and approved the final paper.

Competing interests. The authors declare that they have no conflict of interest.

Acknowledgements. We thank Paudel Adhikari, Namita and Zhihao Zhang for their valuable input related to writing or providing maps of the sampling sites.

480 Financial support. This work was supported by the National Natural Science Foundation of China (grant number 91851207), the Second Tibetan Plateau Scientific Expedition and Research (STEP) program (grant number 2019QZKK0503), the National Key Research and Development Program of China (grant number 2019YFC1509103), and the Strategic Priority Research Program (A) of the Chinese Academy of Sciences (grant number XDA20050101).

References

- 485 <u>Amato, P., Hennebelle, R., Magand, O., Sancelme, M., Delort, A.-M., Barbante, C., Boutron, C., and Ferrari, C.: Bacterial characterization of the snow cover at Spitzberg, Svalbard, FEMS microbiology Ecology, 59, 255-264, 2007.</u> <u>Antony, R., Willoughby, A. S., Grannas, A. M., Catanzano, V., Sleighter, R. L., Thamban, M., Hatcher, P. G., and Nair, S.: Molecular Insights on Dissolved Organic Matter Transformation by Supraglacial Microbial Communities, Environmental Science & Technology, 51, 4328-4337, 10.1021/acs.est.6b05780, 2017.</u>
- Bergk Pinto, B., Maccario, L., Dommergue, A., Vogel, T. M., and Larose, C.: Do organic substrates drive microbial community interactions in Arctic snow?, Frontiers in microbiology, 10, 2492-2492, 10.3389/fmicb.2019.02492, 2019.
 Björkman, M. P., Vega, C. P., Kühnel, R., Spataro, F., Ianniello, A., Esposito, G., Kaiser, J., Marca, A., Hodson, A., Isaksson, E., and Roberts, T. J.: Nitrate postdeposition processes in Svalbard surface snow, Journal of Geophysical Research: Atmospheres, 119, 12,953-912,976, 10.1002/2013jd021234, 2014.
- 495 Bolyen, E., Rideout, J. R., Dillon, M. R., Bokulich, N. A., Abnet, C., Al-Ghalith, G. A., Alexander, H., Alm, E. J., Arumugam, M., and Asnicar, F.: QIIME 2: Reproducible, interactive, scalable, and extensible microbiome data science, PeerJ Preprints2167-9843, 2018.

Borchert, E., Hammerschmidt, K., Hentschel, U., and Deines, P.: Enhancing Microbial Pollutant Degradation by Integrating Eco-Evolutionary Principles with Environmental Biotechnology, Trends in microbiology, 10.1016/j.tim.2021.03.002, 2021.

Brooks, P. D., Williams, M. W., and Schmidt, S. K.: Inorganic nitrogen and microbial biomass dynamics before and during spring snowmelt, Biogeochemistry, 43, 1-15, 10.1023/A:1005947511910, 1998.
 Calcagno, V., Jarne, P., Loreau, M., Mouquet, N., and David, P.: Diversity spurs diversification in ecological communities, Nature Communications, 8, 15810, 10.1038/ncomms15810, 2017.

Callahan, B. J., McMurdie, P. J., and Holmes, S. P.: Exact sequence variants should replace operational taxonomic units in marker-gene data analysis, The ISME journal, 11, 2639, 2017.

Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., and Holmes, S. P.: DADA2: high-resolution sample inference from Illumina amplicon data, Nature methods, 13, 581, 2016.

Cameron, K. A., Stibal, M., Chrismas, N., Box, J., and Jacobsen, C. S.: Nitrate addition has minimal short-term impacts on Greenland ice sheet supraglacial prokaryotes, Environmental Microbiology Reports, 9, 144-150, https://doi.org/10.1111/1758-2220.12510.2017

```
510 2229.12510, 2017.
```

505

Caporaso, J. G., Lauber, C. L., Walters, W. A., Berg-Lyons, D., Huntley, J., Fierer, N., Owens, S. M., Betley, J., Fraser, L., and Bauer, M.: Ultra-high-throughput microbial community analysis on the Illumina HiSeq and MiSeq platforms, The ISME journal, 6, 1621, 2012.

Carey, C. J., Hart, S. C., Aciego, S. M., Riebe, C. S., Blakowski, M. A., and Aronson, E. L.: Microbial community structure

515 of subalpine snow in the Sierra Nevada, California, Arctic, Antarctic, and Alpine Research, 48, 685-701, 10.1657/AAAR0015-062, 2016. Carpenter, E. J., Lin, S., and Capone, D. G.: Bacterial activity in South Pole snow, Appl. Environ. Microbiol., 66, 4514-4517, 2000.

Chen, D., Xu, B., Yao, T., Guo, Z., Cui, P., Chen, F., Zhang, R., Zhang, X., Zhang, Y., Fan, J., Hou, Z., and Zhang, T.:

520 Assessment of past, present and future environmental changes on the Tibetan Plateau, Chinese Science Bulletin, 60, 3025-3035, 2015.

Dang, H. and Lovell, C. R.: Microbial surface colonization and biofilm development in marine environments, Microbiology and Molecular Biology Reviews, 80, 91, 10.1128/MMBR.00037-15, 2016.

Deng, Y., Jiang, Y.-H., Yang, Y., He, Z., Luo, F., and Zhou, J.: Molecular ecological network analyses, BMC bioinformatics, 13, 113, 2012.

Dibb, J. E., Talbot, R. W., Munger, J. W., Jacob, D. J., and Fan, S.-M.: Air-snow exchange of HNO3and NOyat Summit, Greenland, Journal of Geophysical Research: Atmospheres, 103, 3475-3486, 10.1029/97jd03132, 1998.

525

530

Du, S., Ya, T., Zhang, M., Zhu, M., Li, N., Liu, S., and Wang, X.: Distinct microbial communities and their networks in an anammox coupled with sulfur autotrophic/mixotrophic denitrification system, Environmental Pollution, 262, 114190, https://doi.org/10.1016/j.envpol.2020.114190, 2020.

Filippa, G., Freppaz, M., Williams, M. W., and Zanini, E.: Major element chemistry in inner alpine snowpacks (Aosta Valley Region, NW Italy), Cold Regions Science and Technology, 64, 158-166, https://doi.org/10.1016/j.coldregions.2010.07.005, 2010.

Friedman, J. and Gore, J.: Ecological systems biology: The dynamics of interacting populations, Current Opinion in Systems

- Biology, 1, 114-121, https://doi.org/10.1016/j.coisb.2016.12.001, 2017.
 Hacquard, S., Garrido-Oter, R., González, A., Spaepen, S., Ackermann, G., Lebeis, S., McHardy, A. C., Dangl, J. L., Knight, R., and Ley, R.: Microbiota and host nutrition across plant and animal kingdoms, Cell host & microbe, 17, 603-616, 2015.
 Hall, M. H. P. and Fagre, D. B.: Modeled climate-induced glacier change in Glacier National Park, 1850-2100, Bioscience, 53, 131-140, 10.1641/0006-3568(2003)053[0131:Mcigci]2.0.Co;2, 2003.
- 540 Hell, K., Edwards, A., Zarsky, J., Podmirseg, S. M., Girdwood, S., Pachebat, J. A., Insam, H., and Sattler, B.: The dynamic bacterial communities of a melting High Arctic glacier snowpack, The ISME journal, 7, 1814, 2013.

Henry, S., Baudoin, E., Lopez-Gutierrez, J. C., Martin-Laurent, F., Brauman, A., and Philippot, L.: Quantification of denitrifying bacteria in soils by nirK gene targeted real-time PCR, J. Microbiol. Methods, 61, 289-290, 10.1016/j.mimet.2004.12.008, 2005.

545 Hodson, A., Anesio, A. M., Tranter, M., Fountain, A., Osborn, M., Priscu, J., Laybourn-Parry, J., and Sattler, B.: Glacial ecosystems, Ecological Monographs, 78, 41-67, 2008.

Hodson, A. J., Mumford, P. N., Kohler, J., and Wynn, P. M.: The High Arctic glacial ecosystem: new insights from nutrient budgets, Biogeochemistry, 72, 233-256, 10.1007/s10533-004-0362-0, 2005.

Holland, A. T., Bergk Pinto, B., Layton, R., Williamson, C. J., Anesio, A. M., Vogel, T. M., Larose, C., and Tranter, M.: Over

550 Winter Microbial Processes in a Svalbard Snow Pack: An Experimental Approach, Frontiers in Microbiology, 11, 10.3389/fmicb.2020.01029, 2020.

Hunt, D. E. and Ward, C. S.: A network-based approach to disturbance transmission through microbial interactions, Front Microbiol, 6, 1182, 10.3389/fmicb.2015.01182, 2015.

Ji, M., Kong, W., Yue, L., Wang, J., Deng, Y., and Zhu, L.: Salinity reduces bacterial diversity, but increases network complexity in Tibetan Plateau lakes, FEMS Microbiology Ecology, 95, 10.1093/femsec/fiz190, 2019.

Ju, F., Xia, Y., Guo, F., Wang, Z., and Zhang, T.: Taxonomic relatedness shapes bacterial assembly in activated sludge of globally distributed wastewater treatment plants, Environmental microbiology, 16, 2421-2432, 2014. Khan, N., Maezato, Y., McClure, R. S., Brislawn, C. J., Mobberley, J. M., Isern, N., Chrisler, W. B., Markillie, L. M., Barney,

B. M., Song, H.-S., Nelson, W. C., and Bernstein, H. C.: Phenotypic responses to interspecies competition and commensalism
in a naturally-derived microbial co-culture, Scientific Reports, 8, 297, 10.1038/s41598-017-18630-1, 2018.

Kim, M., Jung, J. Y., Laffly, D., Kwon, H. Y., and Lee, Y. K.: Shifts in bacterial community structure during succession in a glacier foreland of the High Arctic, FEMS Microbiology Ecology, 93, 10.1093/femsec/fiw213, 2016. Kohler, T. J., Vinsova, P., Falteisek, L., Zarsky, J. D., Yde, J. C., Hatton, J. E., Hawkings, J. R., Lamarche-Gagnon, G., Hood,

<u>E., Cameron, K. A., and Stibal, M.: Patterns in Microbial Assemblages Exported From the Meltwater of Arctic and Sub-Arctic</u>
 Glaciers, Front Microbiol, 11, 669, 10.3389/fmicb.2020.00669, 2020.

Krug, L., Erlacher, A., Markut, K., Berg, G., and Cernava, T.: The microbiome of alpine snow algae shows a specific interkingdom connectivity and algae-bacteria interactions with supportive capacities, The ISME Journal, 14, 2197-2210, 10.1038/s41396-020-0677-4, 2020.

Larose, C., Dommergue, A., and Vogel, T. M.: Microbial nitrogen cycling in Arctic snowpacks, Environmental Research 570 Letters, 8, 035004, 10.1088/1748-9326/8/3/035004, 20132013a.

Larose, C., Berger, S., Ferrari, C., Navarro, E., Dommergue, A., Schneider, D., and Vogel, T. M.: Microbial sequences retrieved from environmental samples from seasonal Arctic snow and meltwater from Svalbard, Norway, Extremophiles, 14, 205-212, 2010.

Larose, C., Prestat, E., Cecillon, S., Berger, S., Malandain, C., Lyon, D., Ferrari, C., Schneider, D., Dommergue, A., and Vogel,

575 <u>T. M.: Interactions between Snow Chemistry, Mercury Inputs and Microbial Population Dynamics in an Arctic Snowpack,</u> PLOS ONE, 8, e79972, 10.1371/journal.pone.0079972, 2013b

Lazzaro, A., Wismer, A., Schneebeli, M., Erny, I., and Zeyer, J.: Microbial abundance and community structure in a melting alpine snowpack, Extremophiles, 19, 631-642, 2015.

Liu, K., Liu, Y., Hu, A., Wang, F., Zhang, Z., Yan, Q., Ji, M., and Vick-Majors, T. J.: Fate of glacier surface snow-originating
 bacteria in the glacier-fed hydrologic continuums, Environ Microbiol, n/a, 10.1111/1462-2920.15788, 2021.

Liu, Y., Yao, T., Kang, S., Jiao, N., Zeng, Y., Shi, Y., Luo, T., Jing, Z., and Huang, S.: Seasonal variation of snow microbial community structure in the East Rongbuk glacier, Mt. Everest, Chinese Science Bulletin, 51, 1476-1486, 10.1007/s11434-006-1476-7, 2006.

Lopatina, A., Krylenkov, V., and Severinov, K.: Activity and bacterial diversity of snow around Russian Antarctic stations, Research in microbiology, 164, 949-958, 2013.

585

600

Lü, C. and Tian, H.: Spatial and temporal patterns of nitrogen deposition in China: Synthesis of observational data, Journal of Geophysical Research: Atmospheres, 112, https://doi.org/10.1029/2006JD007990, 2007.

Maccario, L., Vogel, T. M., and Larose, C.: Potential drivers of microbial community structure and function in Arctic spring snow, Frontiers in microbiology, 5, 413, 2014.

590 Lutz, S., Anesio, A. M., Edwards, A., and Benning, L. G.: Linking microbial diversity and functionality of arctic glacial surface habitats, Environ Microbiol, 19, 551-565, 10.1111/1462-2920.13494, 2017.

Maccario, L., Carpenter, S. D., Deming, J. W., Vogel, T. M., and Larose, C.: Sources and selection of snow-specific microbial communities in a Greenlandic sea ice snow cover, Scientific Reports, 9, 2290, 10.1038/s41598-019-38744-y, 2019.

Madsen, E. L.: Microorganisms and their roles in fundamental biogeochemical cycles, Current Opinion in Biotechnology, 22, 456-464, https://doi.org/10.1016/j.copbio.2011.01.008, 2011.

Marshall, W. A. and Chalmers, M. O.: Airborne dispersal of antarctic terrestrial algae and cyanobacteria, Ecography, 20, 585-594, DOI 10.1111/j.1600-0587.1997.tb00427.x, 1997

McDonald, D., Clemente, J. C., Kuczynski, J., Rideout, J. R., Stombaugh, J., Wendel, D., Wilke, A., Huse, S., Hufnagle, J., and Meyer, F.: The Biological Observation Matrix (BIOM) format or: how I learned to stop worrying and love the ome-ome, GigaScience, 1, 7, 2012.

Michaud, L., Lo Giudice, A., Mysara, M., Monsieurs, P., Raffa, C., Leys, N., Amalfitano, S., and Van Houdt, R.: Snow Surface Microbiome on the High Antarctic Plateau (DOME C), PLOS ONE, 9, e104505, 10.1371/journal.pone.0104505, 2014.
Miteva, V.: Bacteria in snow and glacier ice, in: Psychrophiles: From biodiversity to biotechnology, Springer, 31-50, 2008.
Møller, A. K., Søborg, D. A., Abu Al-Soud, W., Sørensen, S. J., and Kroer, N.: Bacterial community structure in High-Arctic

605 snow and freshwater as revealed by pyrosequencing of 16S rRNA genes and cultivation, Polar Research, 32, 17390, 10.3402/polar.v32i0.17390, 2013.

Nicholes, M. J., Williamson, C. J., Tranter, M., Holland, A., Poniecka, E., Yallop, M. L., Black, Bloom, G., and Anesio, A.: Bacterial Dynamics in Supraglacial Habitats of the Greenland Ice Sheet, Frontiers in microbiology, 10, 1366-1366, 10.3389/fmicb.2019.01366, 2019.

 Ning, D., Deng, Y., Tiedje, J. M., and Zhou, J.: A general framework for quantitatively assessing ecological stochasticity, Proceedings of the National Academy of Sciences, 116, 16892-16898, 10.1073/pnas.1904623116, 2019.
 Novic, M., Lecnik, B., Hudnik, V., and Pihlar, B.: Carbonate interferences by ion chromatographic determination of anions in mineral waters, Journal of chromatography. A, 764, 249–256, 10.1016/s0021-9673(96)00905-3, 1997. Oksanen, J., Blanchet, F. G., Kindt, R., Legendre, P., O'hara, R., Simpson, G. L., Solymos, P., Stevens, M. H. H., and Wagner, H.: vegan: Community

615 Ecology Package. R package version 1.17-2, http://cran. r-project. org>. Acesso em, 23, 2010, 2010.
Poniecka, E. A., Bagshaw, E. A., Tranter, M., Sass, H., Williamson, C. J., Anesio, A. M., Team, B., and Bloom: Rapid
development of anoxic niches in supraglacial ecosystems, Arctic, Antarctic, and Alpine Research, 50, S100015, 10.1080/15230430.2017.1420859, 2018.

Qiu, J.: Tibetan glaciers shrinking rapidly, Nature News, 15, 2012.

620 Quesada, A. and Vincent, W. F.: Cyanobacteria in the Cryosphere: Snow, Ice and Extreme Cold, in: Ecology of Cyanobacteria II: Their Diversity in Space and Time, edited by: Whitton, B. A., Springer Netherlands, Dordrecht, 387-399, 10.1007/978-94-007-3855-3 14, 2012.

Rauscher, S. A., Seth, A., Liebmann, B., Qian, J. H., and Camargo, S. J.: Regional climate model - Simulated timing and character of seasonal rains in South America, Mon. Weather Rev., 135, 2642-2657, 10.1175/mwr3424.1, 2007.

625 Ren, Z., Martyniuk, N., Oleksy, I. A., Swain, A., and Hotaling, S.: Ecological Stoichiometry of the Mountain Cryosphere, Frontiers in Ecology and Evolution, 7, 10.3389/fevo.2019.00360, 2019.

Rice, E. W., Baird, R. B., Eaton, A. D., and Clesceri, L. S.: Standard methods for the examination of water and wastewater, American Public Health Association Washington, DC2012.

Scheffer, M., Carpenter, S. R., Lenton, T. M., Bascompte, J., Brock, W., Dakos, V., van de Koppel, J., van de Leemput, I. A., 630 Levin, S. A., van Nes, E. H., Pascual, M., and Vandermeer, J.: Anticipating Critical Transitions, Science, 338, 344-348.

10.1126/science.1225244, 2012. Segawa, T., Miyamoto, K., Ushida, K., Agata, K., Okada, N., and Kohshima, S.: Seasonal change in bacterial flora and biomass

in mountain snow from the Tateyama Mountains, Japan, analyzed by 16S rRNA gene sequencing and real-time PCR, Applied and Environmental Microbiology, 71, 123-130, 2005.

- Stegen, J. C., Lin, X., Konopka, A. E., and Fredrickson, J. K.: Stochastic and deterministic assembly processes in subsurface microbial communities, The ISME journal, 6, 1653, 2012.
 Sun, Y., Shen, Y.-x., Liang, P., Zhou, J., Yang, Y., and Huang, X.: Linkages between microbial functional potential and wastewater constituents in large-scale membrane bioreactors for municipal wastewater treatment, Water Research, 56, 162-171, https://doi.org/10.1016/j.watres.2014.03.003, 2014.
- Telling, J., Anesio, A. M., Tranter, M., Irvine-Fynn, T., Hodson, A., Butler, C., and Wadham, J.: Nitrogen fixation on Arctic glaciers, Svalbard, Journal of Geophysical Research-Biogeosciences, 116, Artn G03039, 10.1029/2010jg001632, 2011.
 Telling, J., Stibal, M., Anesio, A. M., Tranter, M., Nias, I., Cook, J., Bellas, C., Lis, G., Wadham, J. L., Sole, A., Nienow, P., and Hodson, A.: Microbial nitrogen cycling on the Greenland Ice Sheet, Biogeosciences, 9, 2431-2442, 10.5194/bg-9-2431-2012, 2012.
- 645 Vitousek, P. M., Hättenschwiler, S., Olander, L., and Allison, S.: Nitrogen and nature, AMBIO: A Journal of the Human Environment, 31, 97-102, 2002.

Wadham, J. L., Hawkings, J. R., Tarasov, L., Gregoire, L. J., Spencer, R. G. M., Gutjahr, M., Ridgwell, A., and Kohfeld, K.
E.: Ice sheets matter for the global carbon cycle, Nature Communications, 10, 10.1038/s41467-019-11394-4, 2019.
Wang, S., Wang, X., Han, X., and Deng, Y.: Higher precipitation strengthens the microbial interactions in semi-arid grassland

soils, Global Ecology and Biogeography, 27, 570-580, https://doi.org/10.1111/geb.12718, 2018.
 Wemheuer, F., Taylor, J. A., Daniel, R., Johnston, E., Meinicke, P., Thomas, T., and Wemheuer, B.: Tax4Fun2: prediction of habitat-specific functional profiles and functional redundancy based on 16S rRNA gene sequences, Environmental

Microbiome, 15, 1-12, 2020. Wu, Y. F., Xing, P., Liu, SH., Adams, J. M., Shi, Y., Li, Y., Song, X., Zhao, X., Chu, H., and Zhang, G.-L.: Depth-Dependent

655 Patterns of Bacterial Communities and Assembly Processes in a Typical Red Soil Critical Zone, Geomicrobiology Journal, 37, 201-212, 10.1080/01490451.2019.1688432, 2020-J., and Wu, Q. L. L.: Enhanced Microbial Interactions and Deterministic Successions During Anoxic Decomposition of Microcystis Biomass in Lake Sediment, Frontiers in Microbiology, 10, 14, 10.3389/fmicb.2019.02474, 2019.

Xia, S., Li, J., and Wang, R.: Nitrogen removal performance and microbial community structure dynamics response to carbon nitrogen ratio in a compact suspended carrier biofilm reactor, Ecological Engineering, 32, 256-262, 2008.

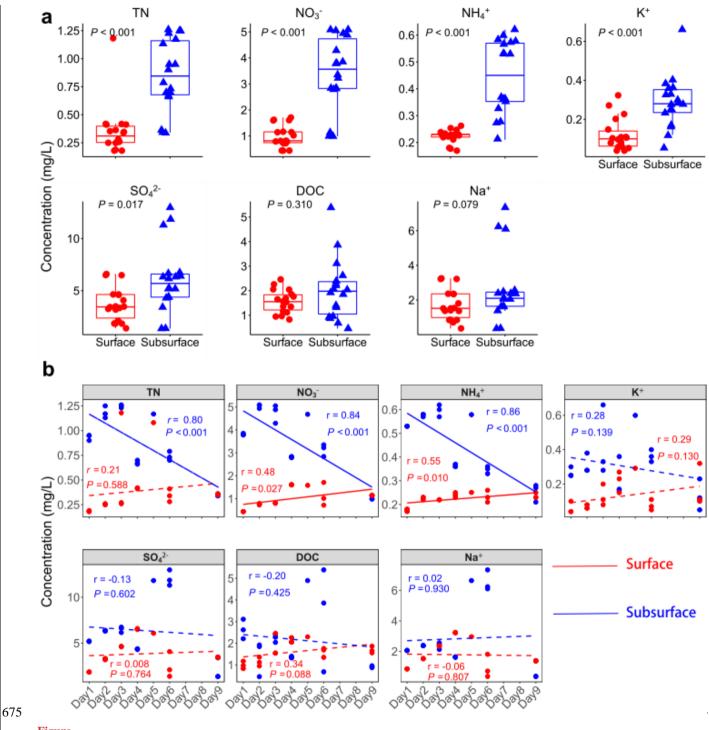
Xiang, S.-R., Shang, T.-C., Chen, Y., and Yao, T.-D.: Deposition and postdeposition mechanisms as possible drivers of microbial population variability in glacier ice, FEMS microbiology ecology, 70, 165-176, 10.1111/j.1574-6941.2009.00759.x, 2009.

Yao, T., Pu, J., Lu, A., Wang, Y., and Yu, W.: Recent Glacial Retreat and Its Impact on Hydrological Processes on the Tibetan
 Plateau, China, and Surrounding Regions, Arctic, Antarctic, and Alpine Research, 39, 642-650, 10.1657/1523-0430(07-

510)[YAO]2.0.CO;2, 2007.

Yuan, H., Huang, S., Yuan, J., You, Y., and Zhang, Y.: Characteristics of microbial denitrification under different aeration intensities: Performance, mechanism, and co-occurrence network, Science of The Total Environment, 754, 141965, https://doi.org/10.1016/j.scitotenv.2020.141965, 2021.

Zhang, J., Shu, X., Zhang, Y., Tan, X., and Zhang, Q.: The responses of epilithic algal community structure and function to light and nutrients and their linkages in subtropical rivers, Hydrobiologia, 847, 841-855, 10.1007/s10750-019-04146-4, 2020. Ziegler, M., Eguíluz, V. M., Duarte, C. M., and Voolstra, C. R.: Rare symbionts may contribute to the resilience of coral–algal assemblages, The ISME Journal, 12, 161-172, 10.1038/ismej.2017.151, 2018.



Figure

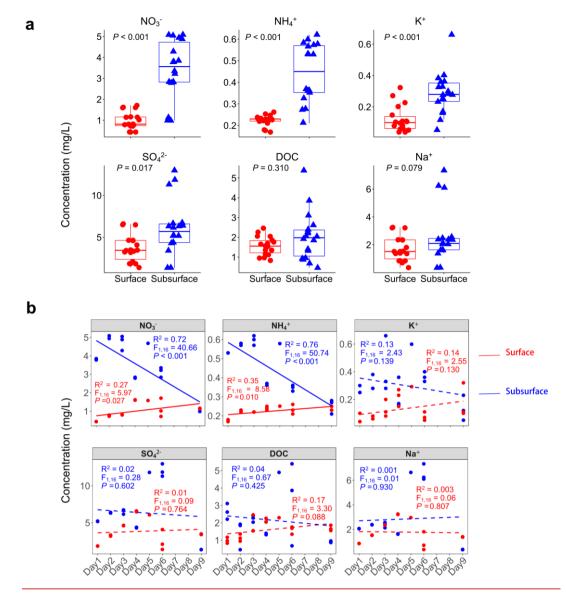


Fig. 1: The pattern of environmental factors changes in the surface and subsurface snow layers.

(a) Environmental factor comparisons in the surface and subsurface snow layers. Each dot represents an individual sample. Significantly higher concentrations of TN, NO₃⁻, NH₄⁺, K⁺, and SO₄²⁻ were observed in the subsurface layer based on Wilcoxon rank-sum test.
(b) Temporal changes of environmental factors in the surface and subsurface layers. The solid and dashed lines indicate significant and non-significant temporal changes, respectively. The concentration of NO₃⁻ and NH₄⁺ in the surface layer significantly increased with time while the concentration of TN, NO₃⁻, and NH₄⁺, in the subsurface layer, significantly increased with time while the concentration of TN, NO₃⁻, and NH₄⁺, in the subsurface layer.

685 decreased with time. Regression analysisSignificance is based on Pearson correlationlinear regression.

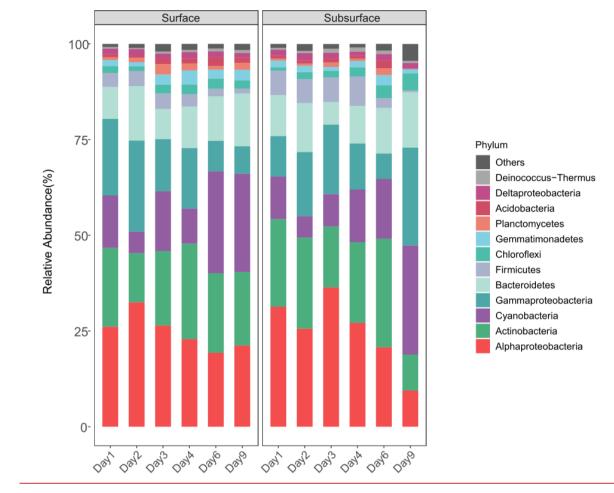


Fig. 2 Taxonomic composition of bacterial community in snow. Only dominant phyla are presented (relative abundance > 1%). The snow community are dominated by Alphaproteobacteria, Actinobacteria, Cyanobacteria, Gammaproteobacteria, Bacteroidetes, Firmicutes, Chloroflexi, Gemmatimonadetes, Planctomycetes, Acidobacteria, Deltaproteobacteria, and Deinococcus-Thermus.

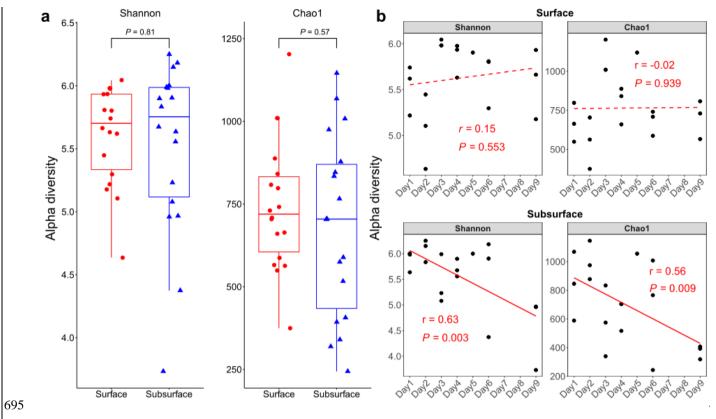


Figure 2:

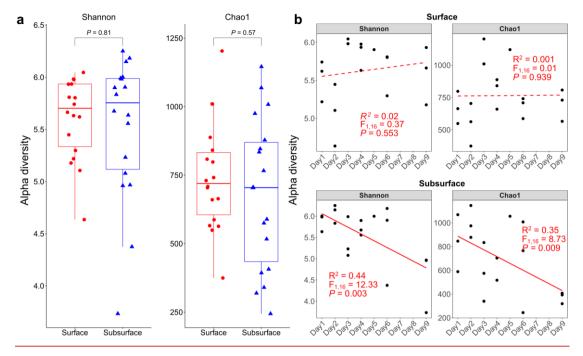
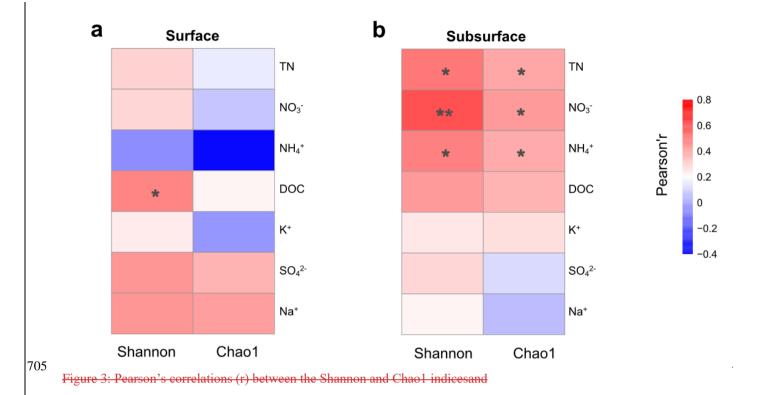


Fig. 3 Bacterial alpha diversity in snow layers.

700

(a) Bacterial alpha diversity comparison between the surface and subsurface layers. Each dot represents an individual sample. For both Shannon and Chao1 indices, no significant difference was observed between the surface and subsurface snow layers. Comparison is based on Wilcoxon rank-sum test.

_(b) Temporal changes of the alpha diversity indices in the surface and subsurface snow layers. For the surface layer, no significant correlation was observed, while both Shannon and Chao1 showed a significantly reduction with time in the subsurface layer. **Regression analysis**<u>Significance</u> is based on <u>Pearson correlationlinear regression</u>.



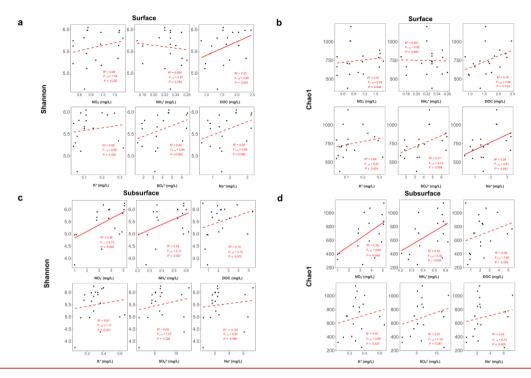
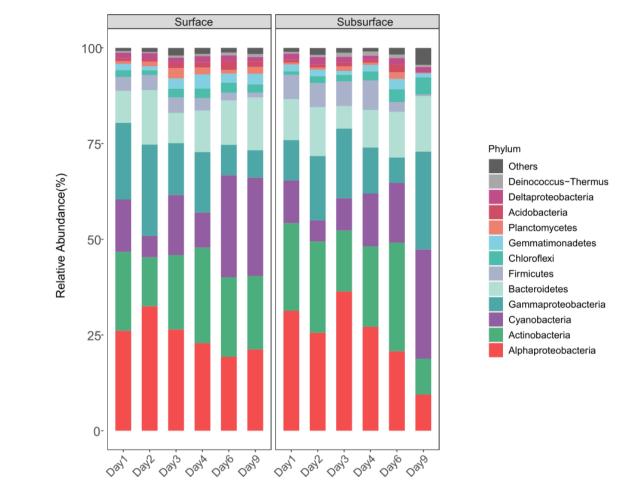
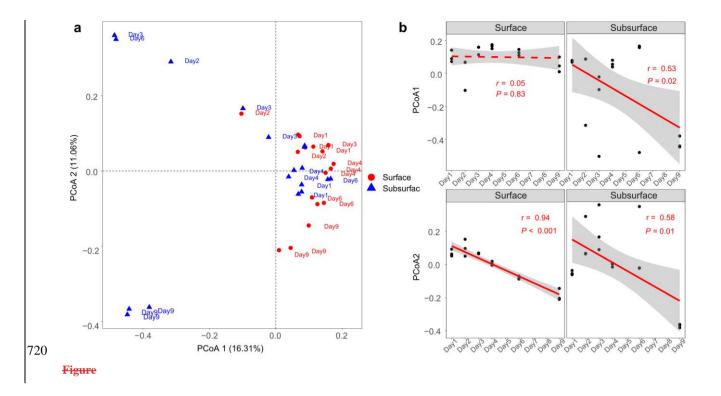


Fig. 4 The influence of environmental factors on bacterial diversity. Correlations of Shannon (a, c) and Chao1 (b, d) diversity indices with environmental factors in the surface (a) and subsurface (b) snow layers. In the surface layer, Shannon diversity index was significantly correlated with DOC. In the subsurface layer, Shannon diversity index was significantly correlated with TN, NO_3 -, Each dot represents an individual sample. The solid and NH_4^+ . Statistical significance dashed lines indicate significant and nonsignificant changes respectively. Significance is indicated by * P < 0.05, ** P < 0.01 based on linear regression.



715 Figure 4: Taxonomic composition of bacterial community in snow. Only dominant phyla are presented (relative abundance > 1%). The snow community are dominated by Alphaproteobacteria, Actinobacteria, Cyanobacteria, Gammaproteobacteria, Bacteroidetes, Firmicutes, Chloroflexi, Gemmatimonadetes, Planetomycetes, Acidobacteria, Deltaproteobacteria, and Deinococcus Thermus.



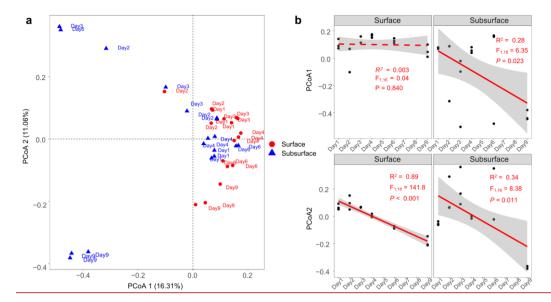
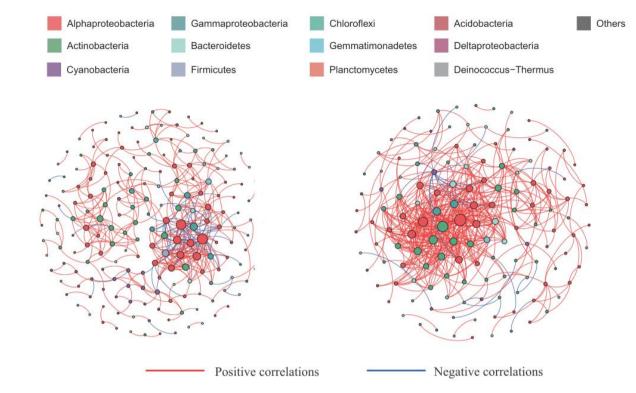


Fig. 5: Principal coordinate analysis (PCoA) of microbial communities in the surface and subsurface snow.

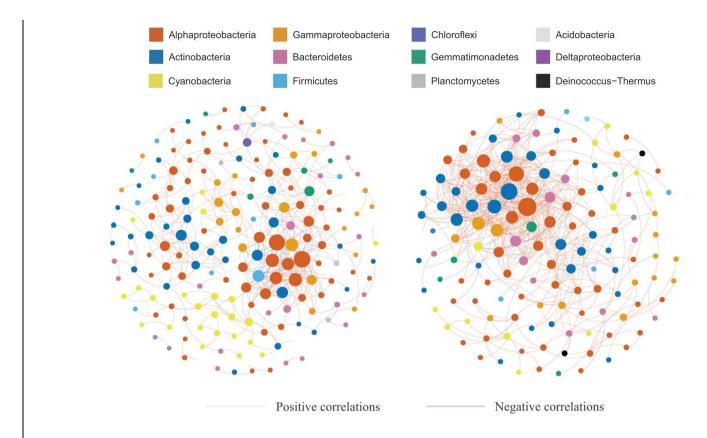
(a) Bray-Curtis distance-based PCoA ordination plot. The microbial community structures of the surface and subsurface snows are significantly different (PERMANOVA, P < 0.001).

(b) Pairwise regression analysis between PCoA scores and sampling time. The solid and dashed lines indicate significant and insignificant changes (based on <u>Pearson correlations]inear regression</u>), respectively. The PCoA1 scores for the bacterial community in <u>the</u> surface layer exhibit no significant correlation with time, while the PCoA2 scores significantly correlated with time. The PCoA1 and PCoA2 are both significantly correlated with time in the subsurface layer.

730



Figure



735 Fig. 6: Bacterial Co-occurrence networks for the surface and subsurface layers communities. Each node represents a bacterial amplicon sequence variant (ASV). The red solid lines represent positive correlations, and the blue solid lines represent negative correlations. Nodes are colored by taxonomy at the phylum level. The subsurface community networks are more complex with a higher positive-to-total correlation ratio.

Table 1. Results of Mantel test showing the relationships between bacterial community composition and environmental factors in the surface and subsurface snow layers. Significant correlations are in **bold**.

	Environmental factor	Surface		Subsurface	
		R	Р	R	Р
	TN	-0.11	0.81	0.36	0.001
	NO ₃ -	0.09	0.21	0.38	0.005
	$\mathbf{NH_{4}^{+}}$	0.01	0.36	0.25	0.01
	DOC	0.08	0.22	0.00	0.49
	Na^+	0.02	0.40	0.16	0.14
	SO_4^{2-}	0.00	0.44	0.25	0.09
	\mathbf{K}^+	0. <u>00</u>	0.56	0.11	0.24
	\mathbf{K}^+	0. <u>00</u>	0.56	0.11	

	Empirical N	letwork
	Surface	Subsurface
No. of node	197	140
No. of edges	436	523
Number of edges per node	2.21	3.73
PosotivePositive links	363	500
Negative links	73	22
Ratio of positive-to-total interactions	83%	95%
Modularity	0.65	0.40
No. of modules	23	12
Average connectivity	4.41	7.36
Average clustering coefficient (avgCC)	0.31	0.39
Average path distance (GD)	5.51	4.72
Average degree (avgK)	4.43	7.57
Graph density	0.02	0.06
Transitivity (Trans)	0.45	0.49
Connectedness (Con)	0.71	0.86

775 Table 2. Topological properties of the empirical networks for the surface and subsurface bacterial communities.