

1 Large carbon cycle sensitivities to climate across a permafrost thaw gradient in subarctic
2 Sweden

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Abstract

26
27 Permafrost peatlands store large amounts of carbon potentially vulnerable to
28 decomposition. However, the fate of that carbon in a changing climate remains uncertain
29 in models due to complex interactions among hydrological, biogeochemical, microbial,
30 and plant processes. In this study, we estimated effects of climate forcing biases present
31 in global climate reanalysis products on carbon cycle predictions at a thawing permafrost
32 peatland in subarctic Sweden. The analysis was conducted with a comprehensive
33 biogeochemical model (*ecosys*) across a permafrost thaw gradient encompassing intact
34 permafrost palsa with an ice core and a shallow active layer, partly thawed bog with a
35 deeper active layer and a variable water table, and fen with a water table close to the
36 surface, each with distinct vegetation and microbiota. Using *in situ* observations to
37 correct local cold and wet biases found in the Global Soil Wetness Project Phase 3
38 (GSWP3) climate reanalysis forcing, we demonstrate good model performance by
39 comparing predicted and observed carbon dioxide (CO₂) and methane (CH₄) exchanges,
40 thaw depth, and water table depth. The simulations driven by the bias-corrected climate
41 suggest that the three peatland types currently accumulate carbon from the atmosphere,
42 although the bog and fen sites can have annual positive radiative forcing impacts due to
43 their higher CH₄ emissions. Our simulations indicate that projected precipitation increases
44 could accelerate CH₄ emissions from the palsa area, even without further degradation of
45 palsa permafrost. The GSWP3 cold and wet biases for this site significantly alter
46 simulation results and lead to erroneous active layer depth (ALD) and carbon budget
47 estimates. Biases in simulated CO₂ and CH₄ exchanges from biased climate forcing are as
48 large as those among the thaw stages themselves at a landscape-scale across the

49 examined permafrost thaw gradient. Future studies should thus not only focus on changes
50 in carbon budget associated with morphological changes in thawing permafrost, but also
51 recognize the effects of climate forcing uncertainty on carbon cycling.

52 1. Introduction

53 Confidence in future climate projections depends on the accuracy of terrestrial
54 carbon budget estimates, which are presently very uncertain (Friedlingstein et al., 2014;
55 Arneeth et al., 2017). In addition to the complexity in physical process representations, a
56 major source of this uncertainty comes from challenges in quantifying climate responses
57 induced by biogeochemical feedbacks. Increases in atmospheric carbon dioxide (CO₂)
58 concentrations can directly stimulate carbon sequestration from plant photosynthesis
59 (Cox et al., 2000; Friedlingstein et al., 2006) and indirectly stimulate carbon emissions
60 (e.g., from soil warming and resulting increased respiration), although the predicted
61 magnitudes of these exchanges strongly depend on model process representations (Zaehle
62 et al., 2010; Grant, 2013, 2014; Ghimire et al., 2016; Chang et al, 2018).

63 The undecomposed carbon stored in permafrost is of critical importance for
64 biogeochemical feedbacks to climate because it is about twice as much as currently is in
65 the atmosphere (Hugelius et al., 2014) and is vulnerable to release to the atmosphere as
66 permafrost thaws (Schuur et al., 2015). O'Donnell et al. (2012) suggested that permafrost
67 thaw would result in a net loss of soil organic carbon from the entire peat column because
68 accumulation rates at the surface were insufficient to balance deep soil organic carbon
69 losses upon thaw. Jones et al. (2017) indicated that the loss of sporadic and discontinuous
70 permafrost by 2100 could result in a release of up to 24 Pg of soil carbon from permafrost
71 peatlands to the atmosphere. Lundin et al. (2016) reported that it is plausible (71%
72 probability) for the subarctic landscapes to serve as a net carbon source to the atmosphere
73 while its peatland components being atmospheric carbon sinks, which highlights the
74 importance of spatial heterogeneity on high latitude carbon budget estimation.

75 In addition to the overall carbon balance of the changing Arctic, the type of
76 carbon gaseous emission is important to climate feedbacks. High latitudes are predicted
77 to get wetter (IPCC, 2014), and saturated anaerobic conditions facilitate methane (CH₄)
78 production, which is a much more efficient greenhouse gas than CO₂ in terms of global
79 warming potential. Even habitats that can be net carbon sinks can produce positive
80 radiative forcing impacts on climate due to CH₄ release, as Bäckstrand et al. (2010)
81 showed for a subarctic peatland. Under projected warming and wetting trends in the
82 Arctic (Collins et al., 2013; Bintanja and Andry, 2017), carbon cycle feedbacks over the
83 permafrost region could become stronger as increased precipitation enhances surface
84 permafrost thaw and strengthens CH₄ emissions by expansion of anaerobic volume
85 (Christensen et al., 2004; Wickland et al., 2006).

86 The Stordalen Mire in northern Sweden (68.20°N, 19.03°E) is in the
87 discontinuous permafrost zone, encompassing a mosaic of thaw stages with associated
88 distinct hydrology and vegetation (Christensen et al. 2004; Malmer et al., 2005),
89 microbiota (Mondav and Woodcroft et al., 2014; Mondav et al., 2017; Woodcroft and
90 Singleton et al., 2018), and organic matter chemistry (Hodgkins et al., 2014). These
91 landscapes have been shifting over the last half-century to a more thawed state, likely due
92 to recent warming (Christensen et al. 2004). Drier hummock sites dominated by shrubs
93 have degraded to wetter sites dominated by graminoids (Malmer et al., 2005; Johansson
94 et al., 2006). The thaw-induced habitat shifts are associated with increases in landscape-
95 scale CH₄ emissions (Christensen et al. 2004; Johansson et al., 2006; Cooper et al., 2017)
96 reflective of the higher CH₄ emissions of the wetter thawed habitats (McCalley et al.,
97 2014). The higher CO₂ uptake in later thaw-stage habitats has not compensated for the

98 increase in positive radiative forcing from elevated CH₄ emissions (Bäckstrand et al.,
99 2010; Deng et al., 2014).

100 The impacts of climate sensitivity on the terrestrial carbon cycle have been
101 investigated at the global scale, and the results highlight the need to consider uncertainty
102 in climate datasets when evaluating permafrost region carbon cycle simulations
103 (Ahlström et al., 2017; Guo et al., 2017; Wu et al., 2017). Ahlström et al. (2017) showed
104 that climate forcing biases are responsible for a considerable fraction (~40%) of the
105 uncertainty range in ecosystem carbon predictions from 18 Earth System Models (ESMs)
106 reported by Anav et al. (2013). Guo et al. (2017) concluded that the differences in climate
107 forcing contribute to significant differences in simulated soil temperature, permafrost
108 area, and Active Layer Depth (ALD). Wu et al. (2017) demonstrated that differences
109 among climate forcing datasets contributes more to predictive uncertainty than
110 differences in apparent model sensitivity to climate forcing. However, notably, none of
111 these studies accessed the effects on CH₄ emissions, and their spatial resolution could not
112 represent site-level spatial heterogeneity observed in arctic tundra (Grant et al. 2017a;
113 2017b).

114 Here, we use the ecosystem model *ecosys*, which employs a comprehensive set of
115 coupled biogeochemical and hydrological processes, to estimate the effects of climate
116 forcing uncertainty and sensitivity on CO₂ and CH₄ exchanges and thaw depth
117 simulations. For the Stordalen Mire site, we estimated bias in the Global Soil Wetness
118 Project Phase 3 (GSWP3) climate reanalysis dataset using site-level long-term
119 meteorological measurements and evaluated impacts on simulated soil and plant
120 processes across the permafrost thaw gradient. This approach enables us to assess model

121 sensitivity to individual climate forcing biases, instead of the aggregated uncertainty
122 range embedded in climate datasets (e.g., variations of climate conditions represented in
123 different climate datasets) presented in previous studies. We address the following
124 questions for our study site at the Stordalen Mire: (1) What are the biases embedded in
125 the GSWP3 climate reanalysis dataset? (2) How do those biases affect model predictions
126 of thaw depth, CO₂ exchanges, and CH₄ exchanges? (3) How does climate sensitivity
127 vary across the stages of permafrost thaw? In addition to improving understanding of
128 permafrost responses to climate, we identify ecosystem carbon prediction uncertainty
129 induced by climate forcing uncertainty in general as the biases found in GSWP3 were
130 consistent with other climate reanalysis datasets during the last decade (section 3).

131

132 **2. Methods and Data**

133 **2.1 Study site description**

134 Our study sites are located at the Stordalen Mire (68.20 °N, 19.03 °E: 351 m
135 above sea level), which is about 10 km southeast of the Abisko Scientific Research
136 Station (ANS) in northern Sweden. The Stordalen Mire is in the discontinuous permafrost
137 zone along the 0 °C isotherm where permafrost at low elevations primarily presents in
138 peatlands, bordered by lakes to the northwest and southeast (Kokfelt et al., 2010). A large
139 portion of the mire consists of a slightly elevated drained area underlain by permafrost
140 characterized by a hummocky topography, and the remaining portion is largely lacking
141 permafrost with fen-like conditions (Johansson et al., 2006). The recent warming (more
142 than 1 °C) has deepened the mean ALD measured at the Stordalen Mire by around 20 cm
143 since the early 1980's, accompanied by palsa collapses and thermokarst erosion

144 (Christensen et al., 2004; Malmer et al., 2005; Johansson et al., 2006). Specifically, the
145 mean ALD has increased from 0.48 m to 0.63 m in the drier part of the mire and from
146 0.63 m to 0.86 m in the wetter part, from September 1973–1976 (Rydén and Kostov,
147 1980) to September 2003–2005 (Johansson et al., 2006).

148 Significant changes in climate over this region have been recorded during the last
149 few decades. The annual mean air temperature measured at the ANS has risen by 2.5 °C
150 from 1913 to 2006, where it exceeded the 0 °C threshold (0.6 °C in 2006) for the first
151 time over the past century (Callaghan et al., 2010). The measured annual total
152 precipitation has also increased from 306 mm y⁻¹ (years 1913 to 2009) to 336 mm y⁻¹
153 (years 1980 to 2009) (Olefeldt and Roulet, 2012), along with increased variability in
154 extreme precipitation (Callaghan et al., 2010). The measured annual maximum snow
155 depth has increased from 59 cm (years 1957 to 1971) to 70 cm (years 1986 to 2000),
156 however, the snow cover period with snow depth greater than 20 cm has decreased from
157 5.8 months (years 1957 to 1971) to 4.9 months (years 1986 to 2000) (Malmer et al.,
158 2005).

159 Inception of peat deposition at the Stordalen Mire has been dated at around 6,000
160 calendar years before present (cal. BP) (Sonesson 1972) in the southern part of the mire
161 and at around 4,700 cal. BP in the northern part (Kokfelt et al., 2010). Kokfelt et al.
162 (2010) suggested that permafrost aggregation initiated during the Little Ice Age (around
163 120–400 cal. BP) in the Stordalen Mire. At present, the Stordalen Mire can be broadly
164 classified into three peatland types: intact permafrost palsa, partly thawed bog, and fen
165 (Hodgkins et al., 2014), hereafter referred to as palsa, bog, and fen (Figure 1). The spatial
166 distribution of these peatland types in 2000 are described in Olefeldt and Roulet (2012).

167 Based on Swedish military photography, all three of the investigated peatland
168 types have existed since at least the 1930's. The palsa sites are ombrotrophic and raised
169 0.5 to 2.0 m above their surroundings, with a relatively thin peat layer (0.4 to 0.7 m,
170 Rydén et al., 1980), thinner active layer depth (less than 0.7 m in late summer), and no
171 measurable water table depth (Bäckstrand et al., 2008a; 2008b; Olefeldt and Roulet,
172 2012). The bog sites are ombrotrophic and are wetter than the palsa sites, with a thicker
173 peat layer (0.5 to ~1 m, Rydén et al., 1980), deeper ALD (greater than 0.9 m, the deepest
174 measurement depth), and water table depth fluctuating from 35 cm below the peat surface
175 to the ground surface (Bäckstrand et al., 2008a; 2008b; Olefeldt and Roulet, 2012). The
176 fen sites are minerotrophic, receiving a large amount of water from a lake to the east of
177 the mire, with water table depths near or above the ground surface (Bäckstrand et al.,
178 2008a; 2008b; Olefeldt and Roulet, 2012).

179 Differences in hydrology and permafrost conditions create high spatial
180 heterogeneity with different soil moisture, pH, and nutrient conditions that support
181 different plant communities (Bäckstrand et al., 2008a; 2008b). The palsa is dominated by
182 dwarf shrubs with some sedges, feather mosses, and lichens (Malmer et al., 2005;
183 Bäckstrand et al., 2008a; 2008b; Olefeldt and Roulet, 2012). The bog is dominated by
184 *Sphagnum* spp. mosses with a moderate abundance of sedges (Malmer et al., 2005;
185 Bäckstrand et al., 2008a; 2008b; Olefeldt and Roulet, 2012). The fen sites we studied are
186 dominated by sedges (Bäckstrand et al., 2008a; 2008b).

187

188 2.2 Field measurements

189 Continuous daily meteorological measurements have been recorded at the ANS
190 since 1913, including air temperature, precipitation, wind speed, wind direction, relative
191 humidity, and snow depth. Measurements of solar radiation, longwave radiation, and soil
192 temperature are also available at the ANS since 1982. The soil thaw depth (measured to
193 90 cm) and water table depth measurements were taken in the three peatland types 3 to 5
194 times per week from early May to mid-October during 2003 to 2007 (Bäckstrand et al.,
195 2008b).

196 CO₂ and CH₄ exchanges at the three peatland types were measured with
197 automated chambers during the thawed seasons from 2002 to 2007 (Bäckstrand et al.,
198 2008b). Chamber lids were removed when snow accumulates in winter (around
199 November), and the sampling periods for each year ranged from 60 days (28 March (day
200 87) to 27 May (day 147)) in 2002 (shortest) to 193 days (28 May (day 148) to 7
201 December (day 341)) (longest) (Bäckstrand et al., 2008b; Bäckstrand et al., 2010). Three
202 chambers were installed in the palsa, another three in the bog, and two more in the fen
203 (we term each chamber a 'subsite' in the following). Each chamber covered an area of
204 0.14 m² with a height of 25–45 cm depending on the vegetation and the depth of insertion
205 and was closed for 5 minutes every 3 hours to measure CO₂ and total hydrocarbon (THC)
206 exchanges. CH₄ exchanges were manually observed approximately 3 times per week, and
207 these measurements were used to quantify the proportion of CH₄ in the measured THC
208 (Bäckstrand et al., 2008a). The CH₄ exchanges were near zero in the palsa sites
209 (Bäckstrand et al., 2008a; Bäckstrand et al., 2008b; Bäckstrand et al., 2010), so they were
210 not used in model evaluation. We used the CO₂ and CH₄ exchanges observed at 3-hourly
211 steps when the R² values recorded in the measurements were greater than 0.8 (Tokida et

212 al., 2007), and then calculated the associated daily mean exchanges when there were 8
213 measurements per day (Table 1). The quality-controlled daily measurements only
214 covered 12.4–33.7% of the daily data points because of the lack of continuous quality-
215 controlled 3-hourly measurements. The data screening was applied to exclude unreliable
216 measurements and avoid biases from inappropriate gap filling, which is necessary for
217 model evaluations. More detailed descriptions of the CO₂ and CH₄ exchanges
218 measurements can be found in Bäckstrand et al. (2008a).

219

220 **2.3 GSWP3**

221 GSWP3 is an ongoing modeling activity that provides global gridded
222 meteorological forcing (0.5° x 0.5° resolution) and investigates changes in energy, water,
223 and carbon cycles throughout the 20th and 21st centuries. The GSWP3 dataset is based on
224 the 20th Century Reanalysis (Compo et al., 2011), using a spectral nudging dynamical
225 downscaling technique described in Yoshimura and Kanamitsu (2008). A more detailed
226 description of the GSWP can be found in Dirmeyer (2011) and van den Hurk et al.
227 (2016).

228 In this study, we extracted the meteorological conditions at the Stordalen Mire
229 from 1901 to 2010 from the GSWP3 climate reanalysis dataset. The 3-hourly products of
230 air temperature, precipitation, solar radiation, wind speed, and specific humidity were
231 interpolated to hourly intervals with cubic spline interpolation to serve as the
232 meteorological inputs used in our model.

233 The GSWP3 dataset was chosen over other existing climate reanalysis datasets for
234 its spatial and temporal resolutions. For example, the Climatic Research Unit (CRU;

235 Harris et al., 2014) dataset provided monthly meteorological forcing at 0.5° x 0.5°
236 resolution; the National Centers for Environmental Prediction (NCEP; Kalnay et al.,
237 1996; Kanamitsu et al., 2002) dataset provided 6-hourly meteorological forcing at T62
238 Gaussian grid (~1.915° x 1.895° resolution); the CRUNCEP (Viovy, 2018) dataset
239 provided 6-hourly meteorological forcing at 0.5° x 0.5° resolution; and the European
240 Centre for Medium-Range Weather Forecasts (ECMWF; Berrisford et al., 2011) dataset
241 provided 3-hourly meteorological forcing with 125 km (~1.125°) horizontal resolution.

242

243 **2.4 Model description**

244 *Ecosys* is a comprehensive biogeochemistry model that simulates ecosystem
245 responses to diverse environmental conditions with explicit representations of microbial
246 dynamics and soil carbon, nitrogen, and phosphorus biogeochemistry. The above-ground
247 processes are represented in multi-layer plant interacting canopies that are allowed to
248 change with changing environmental conditions, and the below-ground processes are
249 represented in multiple soil layers with multi-phase subsurface reactive transport. *Ecosys*
250 operates at variable time steps (down to seconds) determined by convergence criteria, and
251 it can be applied at patch scale (spatially homogenous one-dimensional) and landscape-
252 scale (spatially variable two- or three-dimensional). Detailed descriptions, including
253 inputs, outputs, governing equations, parameters, and references of the *ecosys* model can
254 be found in Grant (2013). A qualitative summary of the *ecosys* model structure is
255 provided in the supplemental material to this article.

256 The *ecosys* model has been extensively tested against eddy covariance fluxes and
257 related ecophysiological measurements with a wide range of sites and weather conditions

258 in boreal, temperate, and tropical forests (Grant et al., 2007a; Grant et al., 2007c; Grant et
259 al., 2009a; Grant et al., 2009b; Grant et al., 2009c; Grant et al., 2010), wetlands (Dimitrov
260 et al., 2011; Grant et al., 2012b; Dimitrov et al., 2014; Mezbahuddin et al., 2014),
261 grasslands (Grant and Flanagan, 2007; Grant et al., 2012a), tundra (Grant et al., 2003;
262 Grant et al., 2011b; Grant 2015; Grant et al., 2015), croplands (Grant et al., 2007b; Grant
263 et al., 2011a), and other permafrost-associated habitats (Grant and Roulet, 2002; Grant,
264 2017a; Grant et al., 2017b). All *ecosys* model structures are unchanged from those
265 described in these earlier studies.

266 **2.5 Experimental design**

267 To evaluate the effects of climate on model predictions, we conducted four sets of
268 simulations at each of the three peatland types at the Stordalen Mire from 1901 to 2010.
269 The climate data from 1901 to 2001 were used for model initialization (i.e., spinup) and
270 those from 2002 to 2010 were used for analysis. The 110 year simulations were
271 performed to ensure the simulation was equilibrated with local climate (Grant et al.
272 2017a).

273 The meteorological conditions for all the simulations were based on the hourly
274 data extracted from the GSWP3 climate reanalysis dataset (section 2.3). The monthly
275 mean bias of the GSWP3 for this location was calculated by comparing it to the air
276 temperature and precipitation measured at the ANS, for years 1913 to 2010 (section 3.1).
277 The full series of air temperature and precipitation extracted from GSWP3 were then
278 bias-corrected using the monthly mean bias calculated from 1913 to 2010; we label this
279 model scenario CTRL. Our bias correction was conceptually similar to the one used in
280 Ahlström et al. (2017), where the bias-corrected climate forcing fields were the ESM

281 outputs adjusted by the corresponding bias calculated from observations in a reference
282 period.

283 The simulation results from CTRL should represent the reliability of applying
284 *ecosys* at the Stordalen Mire because CTRL is driven by the best local climate
285 description. We first evaluated predicted thaw depth, water table depth, and CO₂ and CH₄
286 exchanges using the CTRL simulation (section 3.2 to 3.4). In the second set of
287 simulations, BIASED-COLD, the biased GSWP3 air temperature data was used, and we
288 corrected only the GSWP3 precipitation. Deviations between CTRL and BIASED-COLD
289 reflect biased air temperature's effects on responses across the thaw gradient. In the third
290 set of simulations, BIASED-WET, we bias-corrected the air temperature extracted from
291 GSWP3, which allows us to quantify the effects of biased precipitation. Finally, we used
292 the meteorological conditions directly extracted from GSWP3 to drive our fourth set of
293 simulations, BIASED-COLD&BIASED-WET, which reveals the uncertainty range of
294 subarctic peatland simulation associated with the local biases in GSWP3 climate forcing.

295 While the three peatland types share the same climate conditions, they differ in
296 soil hydrologic conditions and vegetation characteristics (section 2.1; Figure 1). The bulk
297 density and porosity profiles were set to the values reported in Rydén et al. (1980), who
298 suggested a decreasing trend of bulk density and an increasing trend of porosity from
299 *palsa* (0.12 Mgm⁻³ at surface; 92–93% within the upper 10 cm) to *bog* and *fen* (0.06
300 Mgm⁻³ at surface; 96–97% within the upper 10 cm). The peatland soil carbon-to-nitrogen
301 (CN) ratios and pH values were assigned according to Hodgkins et al. (2014), who
302 documented an increasing trend of pH from *palsa* (4.0), to *bog* (4.2), to *fen* (5.7), and a
303 decreasing trend of soil organic matter CN ratio from *bog* (46±18), to *palsa* (39±24), to

304 fen (19 ± 0.4). Common values of field capacity (0.4) and wilting point (0.15) were used
305 for the three peatland types (Deng et al., 2014). The soil property and vegetation
306 parameters used in our simulation for the three peatland types are summarized in
307 Supplemental Material Table1 and Supplemental Material Table2, respectively.

308

309 **3 Results and Discussion**

310 **3.1 GSWP3 climate comparison to observations**

311 As described in section 2.3, we extracted meteorological conditions at the
312 Stordalen Mire from the GSWP3 climate reanalysis dataset. The closest GSWP3 grid cell
313 was centered at 68.0°N and 19.0°E , which covers the Stordalen Mire and the ANS. The
314 annual mean air temperature and precipitation calculated at this GSWP3 grid cell were -
315 3.65°C and 683.88 mm y^{-1} , respectively, for years 1913 to 2010. A cold bias (-3.09°C)
316 was identified in the GSWP3 annual mean air temperature during the 1913 to 2010
317 period, although a very high correlation coefficient ($r = 0.99$) was found when compared
318 with the ANS measurements (Figure 2a). Both time series exhibit an overall warming
319 trend from the early 20th century to the present ($0.01^\circ\text{C y}^{-1}$), with an even larger warming
320 trend from 1980 to 2010 ($0.05^\circ\text{C y}^{-1}$ [ANS] and $0.04^\circ\text{C y}^{-1}$ [GSWP3]).

321 Similarly, the GSWP3 annual total precipitation data correlates well with ANS
322 measurements ($r = 0.80$) but has a wet bias of 380 mm y^{-1} between 1913 and 2010
323 (Figure 2b). An increasing trend in annual total precipitation was recorded in both time
324 series from the early 20th century to present (0.47 mm y^{-2} [ANS] and 1.07 mm y^{-2}
325 [GSWP3]), although a decreasing trend was found from 1980 to 2010 (-0.56 mm y^{-2}
326 [ANS] and -2.39 mm y^{-2} [GSWP3]).

327 The seasonal cycle of the GSWP3 monthly mean air temperature also matches
328 that measured at the ANS, with a very high correlation coefficient ($r = 0.99$; Figure 3a).
329 The underestimation bias and inter-annual variability of GSWP3 air temperature are
330 greater in winter (maximum underestimate in December, at $-4.52\text{ }^{\circ}\text{C}$ with inter-annual
331 variability of $3.53\text{ }^{\circ}\text{C}$) and smaller in summer (minimum underestimate in July, at -1.52
332 $^{\circ}\text{C}$ with inter-annual variability of $1.65\text{ }^{\circ}\text{C}$), respectively.

333 The magnitude and inter-annual variability of the GSWP3 monthly mean
334 precipitation are comparable between winter and summer, while the ANS measurements
335 exhibit stronger seasonality with lower magnitudes during winter. Despite the differences
336 found in seasonal patterns, a high correlation coefficient ($r = 0.64$) was found between the
337 monthly mean precipitation extracted from GSWP3 and the ANS measurements. The
338 overestimation of monthly mean precipitation was greatest in December (43.25 mm
339 month^{-1}) and smallest in August ($18.75\text{ mm month}^{-1}$).

340 These comparisons suggest that GSWP3 air temperature and precipitation data
341 reasonably capture measured seasonal and long-term trends over past decades, but are
342 biased cold and wet compared to observations, especially during winter. Similar cold and
343 wet biases exist in CRUNCEP and ECMWF climate reanalysis datasets during our 2003
344 to 2007 study period (Supplemental Material Figure 1). The annual mean air temperature
345 and precipitation at the Stordalen Mire for years 2003 to 2007 were $-2.49\text{ }^{\circ}\text{C}$ and 795.09
346 mm y^{-1} ; $-2.46\text{ }^{\circ}\text{C}$ and 708.60 mm y^{-1} ; and $-2.28\text{ }^{\circ}\text{C}$ and 765.67 mm y^{-1} in the GSWP3,
347 CRUNCEP, and ECMWF climate reanalysis datasets, respectively.

348

349 **3.2 Model testing**

350 3.2.1 Thaw depth

351 We first evaluated *ecosys* against observations using bias-corrected climate
352 forcing (i.e., the CTRL simulation). Predicted thaw depth agrees well with measurements
353 collected from 2003 to 2007 for all examined peatland types (Figure 4), with a correlation
354 coefficient of 0.95, 0.87, and 0.41 at the palsa, bog, and fen, respectively. Both
355 simulations and observations show that the rate of thaw depth deepening in the summer
356 varies with peatland type (i.e., relatively slow, moderate, and rapid in the palsa, bog, and
357 fen, respectively).

358 Predicted and observed maximum thaw depths (i.e., ALD) in the intact permafrost
359 palsa were between 45 and 60 cm in September. In the partly thawed bog, the simulated
360 thaw depth is slightly shallower than that observed before August. The simulated bog
361 thaw depth exceeds 90 cm by the end of August, which matches the time when measured
362 thaw depth reaches its maximum. In contrast, the thaw depth exceeds 90 cm nearly one
363 month earlier in the fen. The patterns of thawing permafrost presented here are consistent
364 with Deng et al. (2014), who simulated the same site using the DNDC model.

365

366 3.2.2 CO₂ exchanges

367 The daily Net Ecosystem Exchange (NEE) simulated in the CTRL simulation
368 reasonably captures observed seasonal dynamics from 2003 to 2007 for all the examined
369 peatland types (Figure 5). The simulations and observations generally showed net CO₂
370 uptake (with some episodic CO₂ emissions) during summer and release during winter.
371 The observations and simulations also showed large CO₂ emissions in the palsa site
372 during fall of 2004. Simulated fall CO₂ bursts in the three sites in other years could not be

373 confirmed because of a lack of observations during these periods. Similar to the patterns
374 reported in Raz-Yaseef et al. (2016), some episodic CO₂ emission pulses were simulated
375 as surface ice thaws in spring, but there were no measurements to confirm those events.
376 The correlation coefficients of the simulated and observed daily NEE ranged from 0.58 to
377 0.60, and most of the discrepancies between the simulations and observations were within
378 the ranges of NEE variability measured at different subsites (automated chambers) within
379 the same peatland type. The simulated CO₂ uptake rates in the bog were greater than the
380 observations in summer, which could be due to overestimated plant biomass or
381 overestimated CO₂ uptake rate per plant biomass. However, we currently do not have
382 data to examine the cause of this overestimation because the CO₂ flux derived from
383 automated chambers only represents the aggregated results of all controlling factors.

384 As described in section 2.2, simulated CO₂ exchanges were evaluated for 3-hourly
385 and daily time steps when quality-controlled measurements were available (R² values and
386 relative root mean squared errors (RRMSEs) shown in Table 2). Simulated NEE is in
387 reasonable agreement with the 3-hourly NEE measurements with RRMSEs ranging
388 from 8.4 to 19.1%. Model comparisons with observations were generally poorer at daily
389 time steps, although the calculated RRMSEs were comparable to those reported in Deng
390 et al. (2014). We suspect these differences resulted from uncertainty in determining an
391 accurate observed daily NEE that is representative of the entire peatland type. This may
392 be due to (1) limited daily data points (less than 14% across the study period, Table 1)
393 due to lack of continuous quality-controlled 3-hourly measurements, and (2) the large
394 variability of daily NEE ranges measured at different subsites within the same peatland
395 type (Figure 5). Our results thus indicate that NEE is affected by thaw stage (Bäckstrand

396 et al., 2010; Deng et al., 2014) and fine scale spatial heterogeneity of the system. More
397 detailed measurements with higher spatial and temporal resolutions within the same
398 peatland type would be necessary to characterize the effects of this type of heterogeneity.

399

400 3.2.3 Water table depth and CH₄ exchanges

401 Simulated water table depth generally captures observed seasonal patterns
402 measured in the bog and fen sites from 2003 to 2007 (Figure 6a, c). During summer, the
403 predicted bog water table depth fluctuates around the ground surface (-7 to -1 cm), and
404 the predicted water table depth is at or above the ground surface in the fen. Water table
405 depths simulated by *ecosys* are generally higher than measured in the bog, where
406 measured water table depths are often below the ground surface with greater seasonal
407 variability. Simulated fen water table depths have better overall fit to observations, being
408 higher (~5 cm) than measurements in 2003 and 2004, close to measurements in 2005 and
409 2006, and slightly deeper (~2 cm) than measurements in 2007. These differences in
410 modeled and observed water table depth could be driven by the limitations of our one-
411 dimensional column simulation that could not resolve topographic effects and thus hinder
412 the variations of water table depth, which is a particular issue in simulating the dynamic
413 water table of the bog. For example, no excessive water could be transported to the
414 neighboring grids to deepen local water table depth under our current model
415 configuration. A multi-dimensional simulation that includes realistic topographic effects
416 could help improve the representation of water table dynamics, and estimates of the
417 measurement uncertainty would help facilitate the assessment of simulation bias.

418 Simulated and measured daily CH₄ exchanges correlate reasonably well in the
419 bog ($r = 0.49$) and well in the fen ($r = 0.65$) across the study period (Figure 6b, d). Both
420 the simulations and observations have stronger CH₄ emissions during summer with peak
421 emissions in late summer. Some episodic CH₄ emission pulses (Mastepanov et al., 2008)
422 were simulated during shoulder seasons, and the simulated amount of post-growing
423 season CH₄ emissions agrees well with those measured in 2007.

424 Most of the discrepancies between simulated and observed CH₄ emissions were
425 within the variability of measurements across subsites within the same peatland type. The
426 3-hourly and daily RRMSEs ranged from 11.1 to 22.3% (Table 2) and the daily RRMSEs
427 were comparable to results presented in Deng et al. (2014). Our results show that model
428 evaluation of CH₄ emissions with finer temporal resolution observations is not necessarily
429 superior to evaluation with coarser temporal resolution, as compared to the NEE
430 counterpart, which could be related to comparatively lesser CH₄ emission variability
431 measured across subsites within the same peatland type (Figure 6b, d).

432

433 3.3 Variability across the permafrost thaw gradient

434 Thaw rate and ALD increase along the thaw gradient (i.e., palsa to bog to fen),
435 and landscape variations are generally greater than simulated inter-annual variability
436 (Figure 7a). Maximum carbon uptake also increases along the thaw gradient, and
437 variations across the landscape are comparable with simulated intra-seasonal and inter-
438 annual variabilities (Figure 7b). The simulated mean seasonal cumulative NEE were
439 calculated based on the seasonality identified in Bäckstrand et al. (2010) to help facilitate
440 the inter-comparison of carbon budgets estimated at the Stordalen Mire, and to better

441 capture the actual seasonality recorded at the study site. The results show that the
442 magnitude of mean growing season CO₂ uptake is highest in the fen and lowest in the
443 palsa (Table 3). The same rank applies to the magnitude of mean CO₂ emissions over the
444 non-growing season, although differences across the thaw gradient are smaller.

445 CH₄ emission rates increase significantly along the thaw gradient, and the palsa
446 site emissions are negligible (Figure 7c). Mean cumulative CH₄ emissions simulated in
447 the fen are much higher than those in the bog, and most CH₄ emissions occur during the
448 growing season (Table 3). The higher CH₄ emissions in the fen can be attributed to its
449 faster seasonal thaw rate (Figure 7a) and a water table depth close to the surface (Figure
450 6c). Seasonal cumulative NEE and CH₄ emissions from observations could not be
451 accessed due to the lack of continuous quality controlled carbon flux measurements
452 during our study period (Table 1).

453

454 **3.4 Climate sensitivity of permafrost thaw**

455 **3.4.1 Thaw responses to climate**

456 Our results indicate that the ALD currently simulated in the bog and fen is around
457 108 cm and 130 cm, respectively. However, the maximum depth of our thaw depth
458 measurements is 90 cm, which makes it difficult to evaluate our model performance on
459 ALD simulation. Our results highlight the need to acquire measurements at deeper depth
460 to resolve whether there is no permafrost currently remaining in the bog and fen, or there
461 is a talik with permafrost developed deeper than the simulated ALDs. Such information
462 could be important in predicting microbial activity and thermokarst in permafrost

463 peatlands (Schuur et al., 2015), but it may not significantly alter the effects of climate
464 forcing uncertainty discussed in our study.

465 For each of the four sets of simulations with different climate forcing (section
466 2.5), simulated mean ALD from 2003 to 2007 is always greatest in the fen and lowest in
467 the palsa (Figure 8). This consistent trend along the thaw gradient indicates that ALDs
468 are largely regulated by their distinct ecological and hydrological conditions, because all
469 three sites had the same climate forcing in each set of simulations (i.e., CTRL, BIASED-
470 COLD, BIASED-WET, and BIASED-COLD&BIASED-WET). Therefore, the palsa,
471 bog, and fen have different resilience against the changes in climate forcing, and this type
472 of ecosystem resilience plays an important role in determining ALD under changes in
473 climate conditions.

474 Effects of climate on simulated ALD are similar across peatland types (Figure 8).
475 With increased precipitation (BIASED-WET vs. CTRL), simulated ALD generally
476 becomes deeper with greater inter-annual variability because the increased snowpack
477 depth keeps the soil warmer with lower soil ice content during winter. This effect is less
478 prominent in the comparison between experiments BIASED-COLD and BIASED-
479 COLD&BIASED-WET, because the cold biases in these two experiments (section 3.1)
480 constrain thaw depth development. For example, summertime soil heating in some of the
481 simulation years was not strong enough to thaw the soil ice between 20-40 cm completely
482 in the BIASED-COLD&BIASED-WET run, resulting in shallower ALDs simulated in
483 the palsa and fen even with the snowpack warming effect. The simulated ALD also
484 becomes deeper with higher air temperature (CTRL vs. BIASED-COLD; BIASED-WET
485 vs. BIASED-COLD&BIASED-WET) at all the examined peatland types. This response

486 is more evident in the comparison between experiments BIASED-WET and BIASED-
487 COLD&BIASED-WET, probably driven by their wet biases (section 3.1) that facilitate
488 thaw depth deepening (via increased thermal conductivity and advective heat transport;
489 Grant et al. 2017a). Similar dependencies between ALD and climate were shown in
490 Åkerman and Johansson (2008) and Johansson et al. (2013), based on multi-year
491 measurements and snow manipulation experiments.

492 Therefore, the combined cold and wet biases in the GSWP3 climate reanalysis
493 dataset could counteract their individual effects on simulated ALD development at the
494 Stordalen Mire. Our results indicate a 28.6%, 0.7%, and 11.7% underestimation of ALD
495 simulated in the palsa, bog, and fen, respectively, when applying the GSWP3 climate
496 reanalysis data over this region without proper bias correction (BIASED-
497 COLD&BIASED-WET vs. CTRL). Our sensitivity analysis suggests that projected
498 warming and wetting trends (Collins et al., 2013) could significantly increase ALD in the
499 Arctic, since increases in precipitation and air temperature can both contribute to ALD
500 deepening.

501

502 **3.4.2 Carbon budget responses to climate**

503 Simulations with the four climate forcing datasets (section 2.5) indicate annual
504 mean (from 2003 to 2007) CO₂ sinks and CH₄ sources, except the weak CO₂ emissions
505 simulated in the fen in experiment BIASED-COLD&BIASED-WET due to reduced
506 sedge productivity driven by increased temperature and oxygen stresses (Figure 9a,b).
507 Our results also indicate that differences in annual CO₂ and CH₄ exchanges across the
508 four climate forcing datasets for a single peatland type are as large as those across

509 peatland types for a single climate forcing dataset (Figure 9a,b). These large CO₂ and
510 CH₄ exchanges climate sensitivities demonstrate that the peatland's dynamical responses
511 to climate have stronger effects on the carbon cycle than on ALDs (Figure 8).

512 With bias-corrected precipitation, increased air temperature (CTRL vs. BIASED-
513 COLD) leads to stronger CO₂ uptake and greater CH₄ emissions at all the examined
514 peatland types (Figure 9a,b), mainly because enhanced sedge growth facilitates carbon
515 cycling under a warmer environment (results not shown). This air temperature sensitivity
516 affects CO₂ and CH₄ exchanges within the same peatland type without significantly
517 changing ALD (Figure 8). For both experiments, CO₂ uptake and CH₄ emissions are
518 greatest in the fen and lowest in the palsa, consistent with the measurements reported in
519 Bäckstrand et al. (2010) for the same period. Based on the Coupled Model
520 Intercomparison Project, phase 5 (CMIP5) ESM simulations, arctic annual mean surface
521 air temperature is projected to increase by 8.5±2.1 °C over the 21st century (Bintanja and
522 Andry, 2017). This projected air temperature increase is more than double the air
523 temperature difference between site-observed and GSWP3 temperatures, which could
524 significantly enhance CH₄ emissions regardless of palsa degradation into bog and fen.

525 On the other hand, wet biases (BIASED-WET and BIASED-COLD&BIASED-
526 WET) increase CH₄ emissions in the palsa; wetter and colder conditions result in as much
527 CH₄ release as the current fen, while wetter conditions alone drive palsa emissions
528 comparable to the current bog (Figure 9b). The large precipitation sensitivity found in
529 palsa CH₄ emissions could have strong effects on palsa carbon cycling because arctic
530 precipitation is projected to increase by 50 – 60% towards the end of the 21st century
531 (based on CMIP5 estimates; Bintanja and Andry, 2017). The comparison between

532 experiments BIASED-WET and BIASED-COLD&BIASED-WET shows that in the
533 tundra, increased air temperature strengthens CO₂ uptake and weakens CH₄ emissions.
534 This shift is primarily driven in the model by increased shrub and moss productivity
535 under the warmer environment, which facilitate CO₂ uptake while drying out the soil and
536 reducing CH₄ emissions (results not shown). In the bog and fen sites, increased air
537 temperature under wet bias strengthens both the simulated CO₂ uptake and CH₄
538 emissions (BIASED-WET vs. BIASED-COLD&BIASED-WET), due to enhanced sedge
539 growth under the warmer environment that facilitates carbon cycling in the experiment
540 BIASED-WET. The low CH₄ emissions in bog and fen simulated in experiment
541 BIASED-COLD&BIASED-WET are driven by increased temperature and oxygen
542 stresses that greatly reduce heterotrophic respiration (CH₄ production) and sedge cover
543 (aerenchyma transport).

544 We assessed the integrated effects of the changes in CO₂ and CH₄ exchanges
545 identified in the full suite of simulations in terms of the Net Carbon Balance (NCB) and
546 net emissions of greenhouse gases expressed as CO₂ equivalents (Net Greenhouse Gas
547 Balance; NGGB). NCB was defined as the sum of the annual total CO₂ and CH₄
548 exchanges. NGGB was defined in a similar fashion as the NCB, but considers the greater
549 radiative forcing potential of CH₄ than CO₂ (28 times over a 100-year horizon, Myhre et
550 al., 2013) when calculating the annual total. The calculated NCB values are mostly
551 negative because the stronger CO₂ uptake dominates the weaker CH₄ emissions (Figure
552 9c). The results suggest that all the examined peatland types serve as net carbon sinks
553 under current climate (CTRL), consistent with the estimates reported in Deng et al.
554 (2014) and Lundin et al. (2016). We find a 24, 36, and 38 g C m⁻² y⁻¹ underestimation of

555 NCB simulated in the palsa, bog, and fen sites, respectively, due to the cold and wet
556 biases in the GSWP3 climate reanalysis dataset (BIASED-COLD&BIASED-WET vs.
557 CTRL). NGGB is affected more strongly by CH₄ emissions (Figure 9d) due to its larger
558 radiative forcing potential. NGGB values are positive over the bog and fen, suggesting
559 that these sites have positive radiative forcing impacts despite being net carbon sinks.
560 NGGB simulated in the palsa is generally negative (i.e., a net sink from the atmosphere)
561 due to lower CH₄ emissions, except for the simulation conducted without any climate bias
562 correction (correcting only air temperature increased CH₄ emissions but not enough to
563 compensate for the significantly higher CO₂ sink). Our results indicate that the simulated
564 NGGB would be biased by 298, -66, and -252 g CO₂-eq m⁻² y⁻¹ in the palsa, bog, and fen,
565 respectively, without proper bias correction for the GSWP3 climate reanalysis dataset
566 (BIASED-COLD&BIASED-WET vs. CTRL). Using the GSWP3 products directly thus
567 effectively eliminates the positive radiative forcing from the expanding bog and fen,
568 while creating a potentially dramatically inaccurate positive radiative forcing from the
569 shrinking palsa.

570

571 **3.4.3 Climate sensitivity versus landscape heterogeneity**

572 Climate sensitivity and landscape heterogeneity are defined here as variability
573 across the four climate forcing datasets for a single peatland type, and variability across
574 three peatland types with bias-corrected climate (CTRL), respectively. We estimated
575 carbon cycle variability associated with climate sensitivity and landscape heterogeneity to
576 quantify the corresponding uncertainty in our annual carbon cycle assessments from 2003
577 to 2007. Our results indicate that differences in simulated annual mean CO₂ exchanges

578 and NCB from climate sensitivity are greater than those from landscape heterogeneity
579 (Figure 9a,c); i.e., annual CO₂ uptake strength is more sensitive to climate forcing
580 uncertainty than to peatland type representation. In terms of the simulated annual mean
581 CH₄ emissions and NGGB, our results indicate that variability from climate sensitivity is
582 comparable to those from landscape heterogeneity (Figure 9b,d). Therefore, bias-
583 corrected climate and realistic peatland characterization are both necessary to reduce the
584 uncertainty in representing carbon cycling dynamics and their radiative forcing effects.

585 In addition to their effects on carbon cycle predictions, changes in climate
586 conditions also affect permafrost degradation and thus induce changes in areal cover of
587 peatland types. Malmer et al. (2005) showed that there were -0.95, 0.24, and 0.62 ha areal
588 cover changes (-10.3%, 4.0%, and 46.3% percentage changes) from 1970 to 2000 in
589 tundra, bog, and fen, respectively, at the Stordalen Mire. By applying the annual mean
590 CO₂ and CH₄ exchanges simulated with bias-corrected climate from 2003 to 2007, the
591 areal cover changes from 1970 to 2000 alone would lead to -44 kg C y⁻¹, 76 kg C y⁻¹, and
592 2076 kg CO₂-eq y⁻¹ changes in annual mean CO₂ exchanges, CH₄ exchanges, and NGGB,
593 respectively, at the Stordalen Mire. The changes in landscape-scale carbon cycle
594 dynamics indicate that the radiative warming impact of increased CH₄ emissions is large
595 enough to offset the radiative cooling impact of increased CO₂ uptake at the Stordalen
596 Mire, consistent with the estimates reported in Deng et al. (2014). The areal cover
597 changes across peatland types could persist or accelerate under the projected warming
598 and wetting trends in the Arctic (Collins et al., 2013; Bintanja and Andry, 2017), which
599 could stimulate CH₄ emissions and produce a stronger radiative warming impact.
600

601 **4. Conclusions**

602 We evaluated the climate bias in a widely used atmospheric reanalysis product
603 (GSWP3) at our northern Sweden Stordalen Mire site. We then applied a comprehensive
604 biogeochemistry model, *ecosys*, to estimate the effects of these biases on active layer
605 development and carbon cycling across a thaw gradient at the site. Our results show that
606 *ecosys* reasonably represented measured hydrological, thermal, and biogeochemical cycle
607 processes in the intact permafrost palsa, partly thawed bog, and fen. We found that the
608 cold and wet biases in the GSWP3 climate reanalysis dataset significantly alter model
609 simulations, leading to biases in simulated Active Layer Depths, Net Carbon Balance,
610 and Net Greenhouse Gas Balance by up to 28.6%, $38 \text{ g C m}^{-2} \text{ y}^{-1}$, and $298 \text{ g CO}_2\text{-eq m}^{-2}$
611 y^{-1} , respectively. The Net Carbon Balance simulated with bias-corrected climate suggests
612 that all the examined peatland types are currently net carbon sinks from the atmosphere,
613 although the bog and fen sites can have positive radiative forcing impacts due to their
614 higher CH_4 emissions.

615 Our results indicate that the annual means of ALD, CO_2 uptake, and CH_4
616 emissions generally increase along the permafrost thaw gradient at the Stordalen Mire
617 under current climate, consistent with previous studies in this region. Our analysis
618 suggests that palsa, bog, and fen differ strongly in their carbon cycling dynamics and
619 have different responses to climate forcing biases. Differences in simulated CO_2 and CH_4
620 exchanges driven by uncertainty from climate forcing are as large as those from
621 landscape heterogeneity across the examined permafrost thaw gradient. Model
622 simulations demonstrate that the palsa site exhibits the strongest sensitivity to biases in
623 air temperature and precipitation. The wet bias in GSWP3 could erroneously increase

624 predicted CH₄ emissions from the palsa site to a magnitude comparable to emissions
625 currently measured in **the** bog and fen sites. These results also show that increased
626 precipitation projected for high latitude regions could strongly accelerate CH₄ emissions
627 from the palsa area, even without degradation of palsa into bog and fen. Future studies
628 should thus recognize the effects of climate forcing uncertainty on carbon cycling, in
629 addition to tracking changes in carbon budgets associated with areal changes in
630 permafrost degradation.

631

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641

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969

970 Table 1. Temporal coverage of quality-controlled CO₂ and CH₄ exchanges measured by
 971 automated chambers at the three peatland types in the Stordalen Mire during the years
 972 2002 to 2007.

Sites	Number of data points	CO ₂		Number of data points	CH ₄	
		3 Hourly coverage (%)	Daily coverage (%)		3 Hourly coverage (%)	Daily coverage (%)
Palsa	12752	65.8	12.4	N/A	N/A	N/A
Bog	12821	68.5	12.7	6660	96.2	25.0
Fen	8989	63.8	13.7	4923	90.5	33.7

973 **Note:** The temporal coverage represents the percentage of data points that passed our
 974 quality-controlled threshold at the corresponding time steps. See main text for threshold
 975 details.

976 Table 2. Evaluation of the 3 hourly and daily CO₂ and CH₄ exchanges simulated at the
 977 palsa, bog, and fen sites.

Sites	C component	3-Hourly		Daily	
		R ²	RRMSE (%)	R ²	RRMSE (%)
Palsa	CO ₂	0.48	13.4	0.36	18.3
Bog	CO ₂	0.63	19.1	0.44	35.8
	CH ₄	0.31	16.3	0.47	22.3
Fen	CO ₂	0.64	8.4	0.43	25.5
	CH ₄	0.44	11.1	0.54	16.9

978 Note: RRMSE is relative root mean squared error.

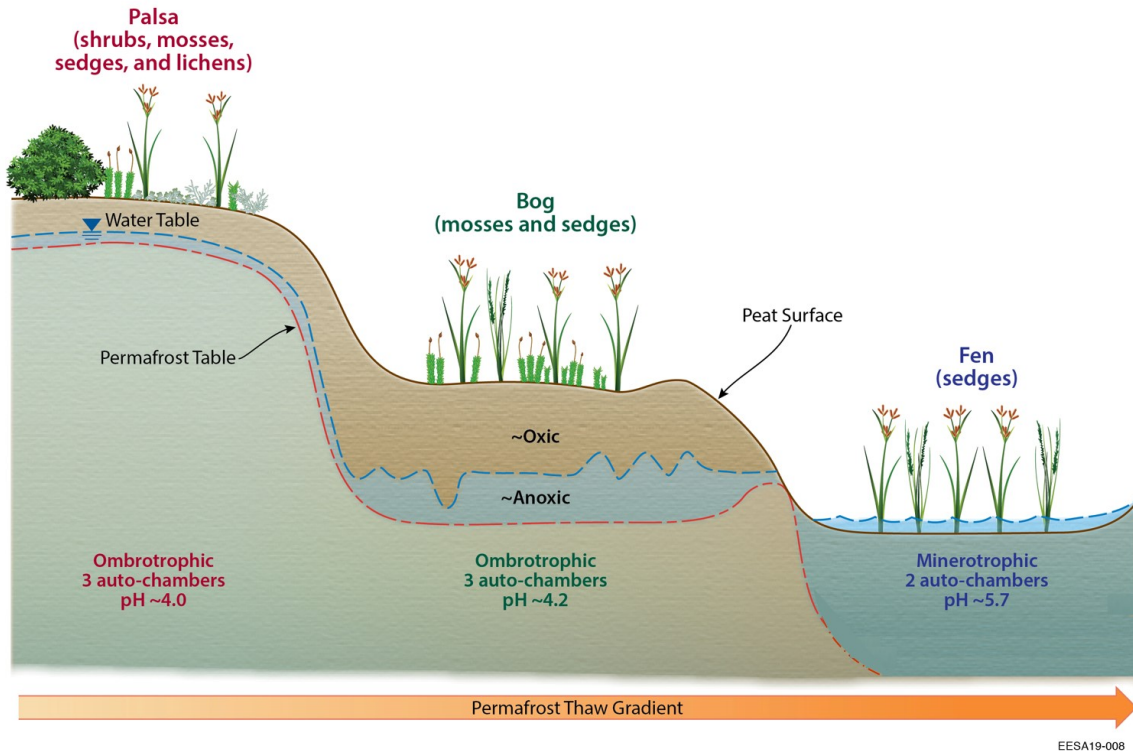
979

980 Table 3. Means and standard deviations of cumulative CO₂ and CH₄ exchanges simulated
 981 in the palsa, bog, and fen during the period 2003 to 2007.

Sites	C flux component	Growing season; Days 119–288		Non-growing season; Days 1–118 and 289–365	
		Mean	Standard deviation	Mean	Standard deviation
Palsa	CO ₂	-72.70	19.10	38.89	4.09
	CH ₄	0.04	0.02	0.01	0.002
Bog	CO ₂	-79.59	21.46	42.89	2.16
	CH ₄	3.52	0.45	0.42	0.11
Fen	CO ₂	-88.65	7.26	44.41	6.13
	CH ₄	10.86	3.95	0.78	0.18

982 Note: All gas exchanges are in units of g C m⁻².

983

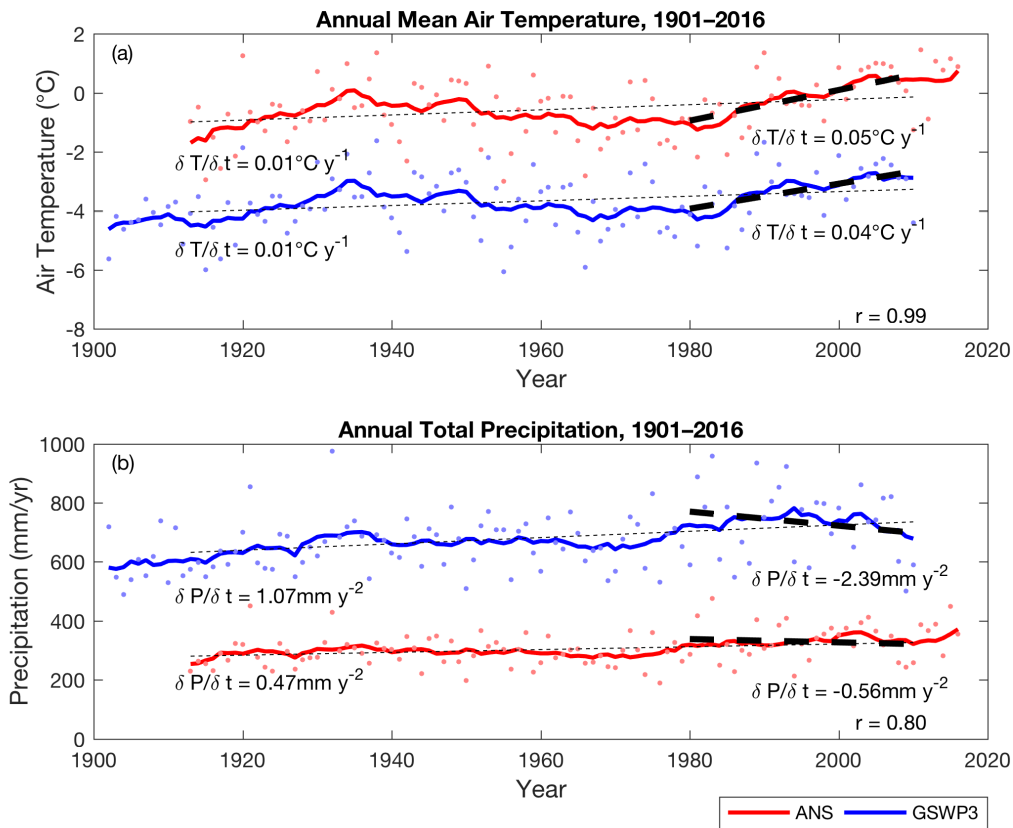


984

985 Figure 1. Schematic diagram of the sampling sites at Stordalen Mire, adapted from

986 Johansson et al. (2006).

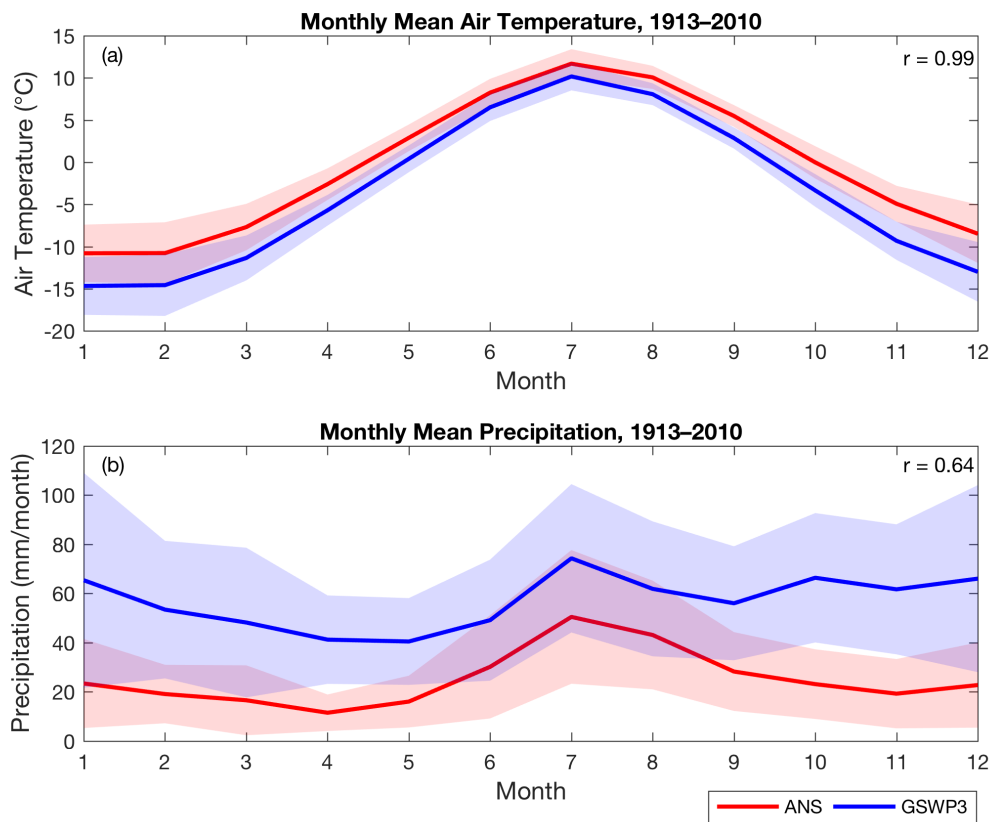
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988

989 Figure 2. Time series of air temperature (a) and precipitation (b) measured at ANS (red;
 990 years 1913–2016) and extracted from GSWP3 (blue; years 1901–2010). Dots are the
 991 annual means and solid lines are the decadal moving averages of the corresponding
 992 annual means. Thin and thick dashed lines are the trends for years 1913–2010, and years
 993 1980–2010, respectively. The inset r values are the correlation coefficients calculated
 994 between the two time series.

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996

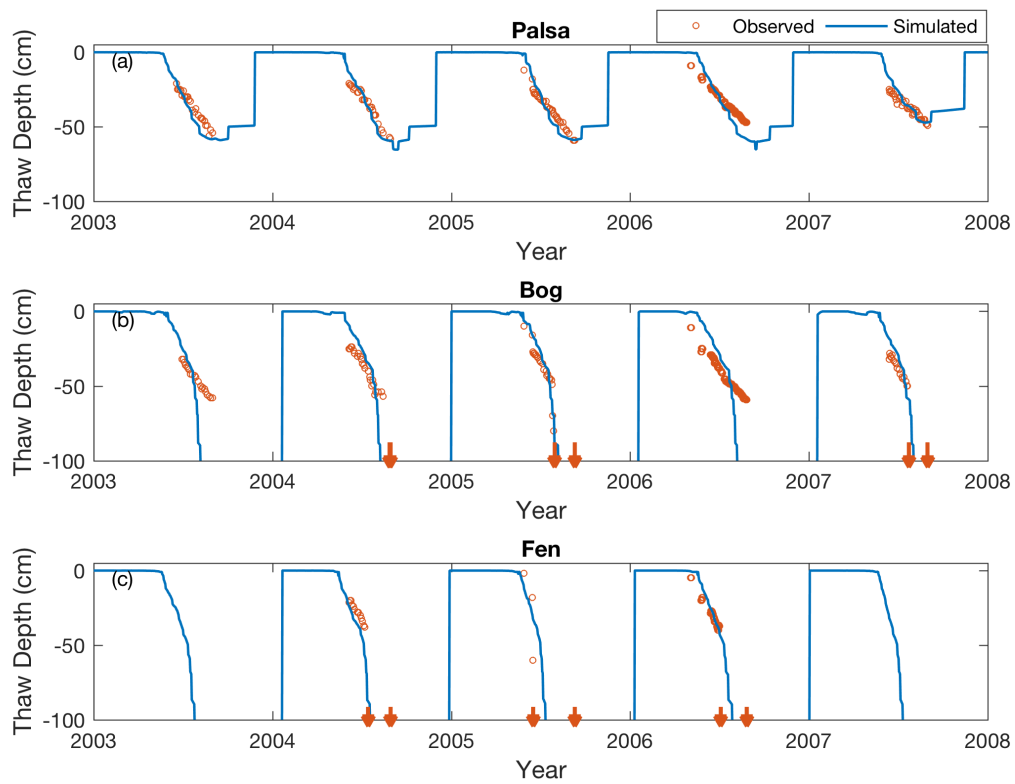
997 Figure 3. Monthly mean air temperature (a) and precipitation (b) measured at ANS (red)

998 and extracted from GSWP3 (blue). The shaded area is the inter-annual variability for the

999 corresponding dataset, represented by the standard deviations calculated at each month.

1000 The inset r values are the correlation coefficients calculated between the two time series.

1001



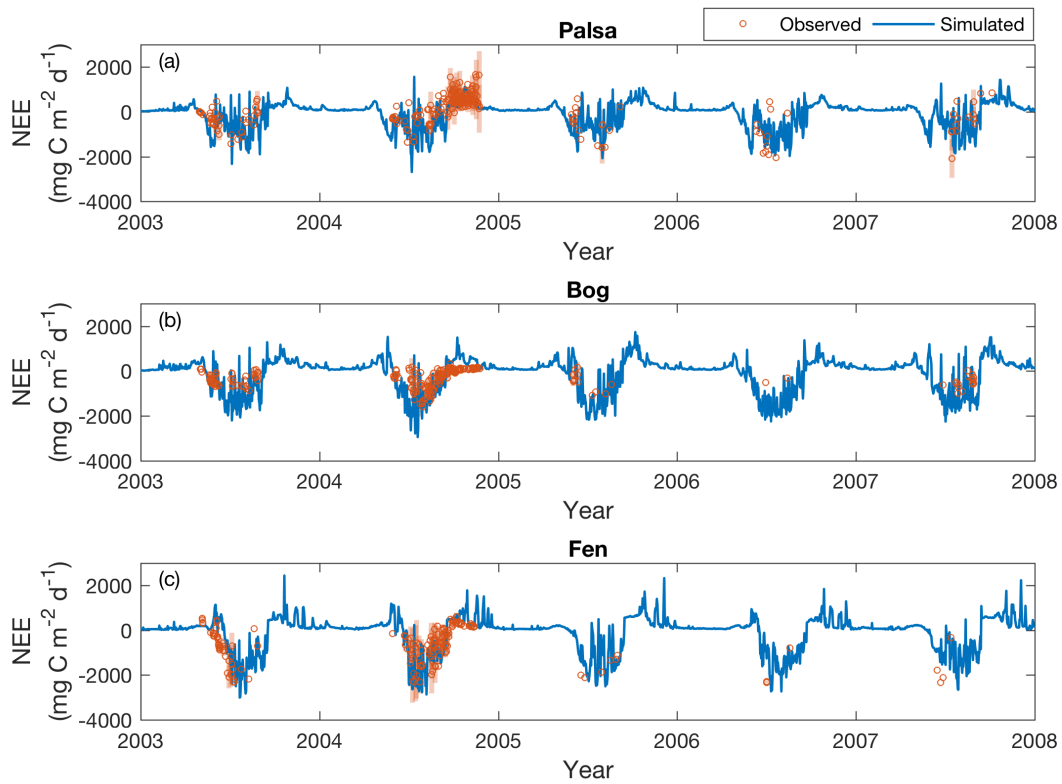
1002

1003 Figure 4. Simulated (solid lines) and measured (open circles) seasonal dynamics of thaw

1004 depth at the palsa (a), bog (b), and fen (c) sites from 2003 to 2007. Downward arrows

1005 indicate the time when measured thaw depth exceeds 90 cm for a measurement year.

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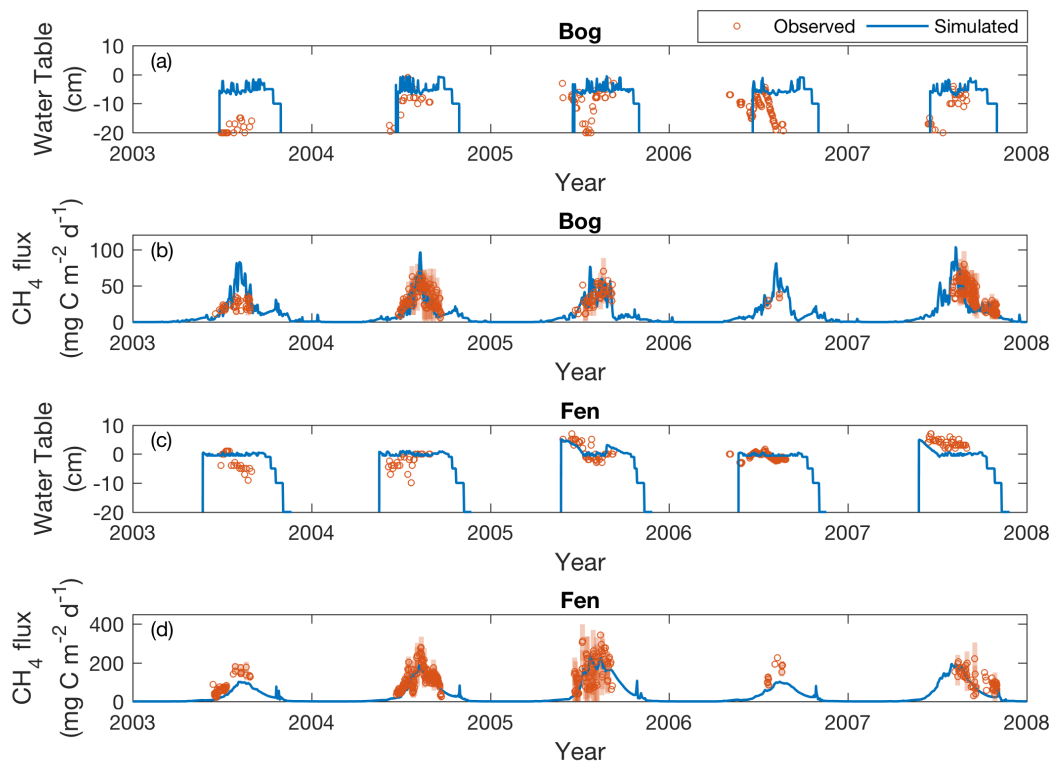
1008 Figure 5. Simulated (solid lines) and measured (open circles) daily CO₂ exchanges (NEE)

1009 at the palsa (a), bog (b), and fen (c) sites, from 2003 to 2007. Shaded bars are the

1010 standard deviations of daily NEE measured across subsites under each peatland type.

1011 Positive and negative values indicate effluxes from and influxes to the site, respectively.

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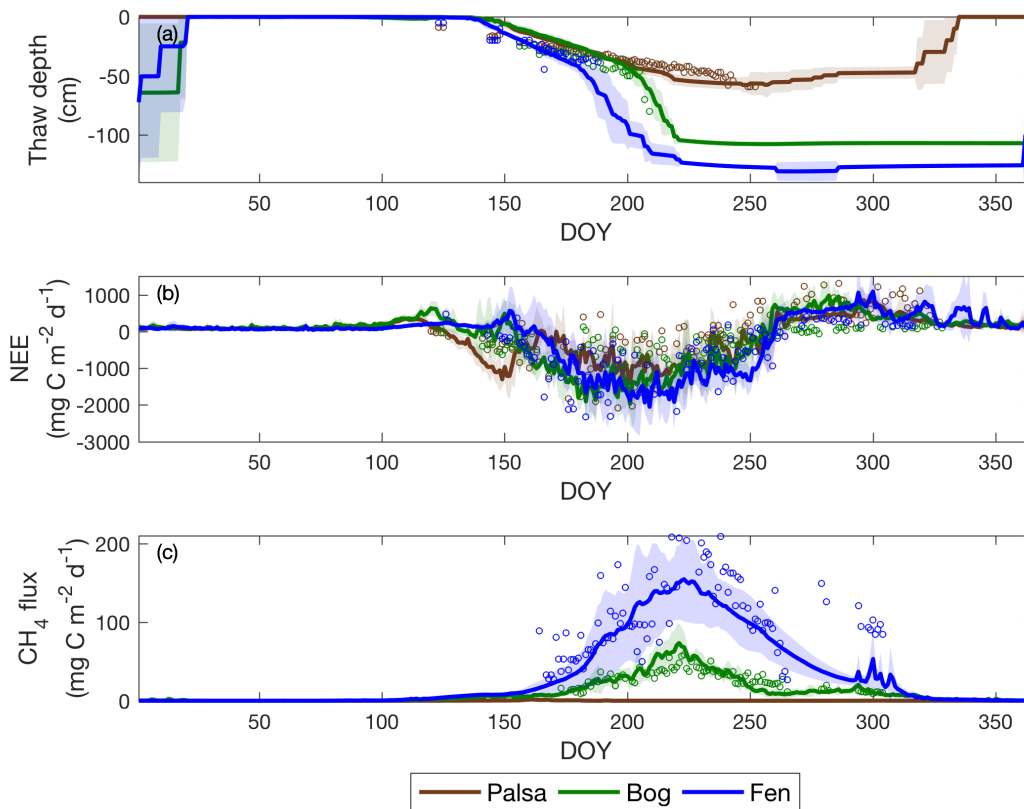
1014 Figure 6. Simulated (solid lines) and measured (open circles) water table depths

1015 CH₄ emissions at the bog and fen from 2003 to 2007. Shaded bars are the standard

1016 deviations of the daily CH₄ emissions measured across the subsites under each peatland

1017 type.

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1020 Figure 7. Daily thaw depth (a), daily NEE (b), and daily CH₄ (c) exchanges for the three

1021 sites from 2003 to 2007. Solid lines and open circles are the simulated and measured

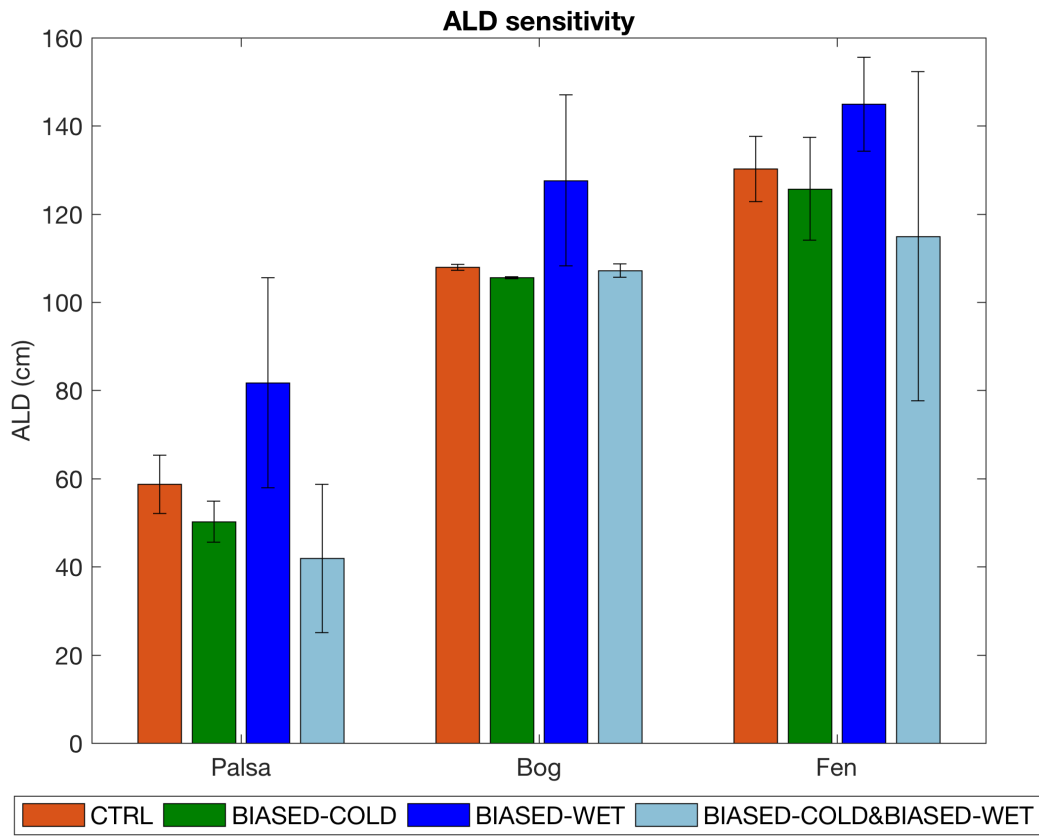
1022 inter-annual means for each day of year, respectively. The shaded area is the simulated

1023 inter-annual variability for the corresponding dataset, represented by the standard

1024 deviations calculated at each day of year. Positive and negative carbon flux values

1025 indicate effluxes from and influxes to the site, respectively.

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Figure 8. Simulated ALD at the palsa, bog, and fen for four sets of climate forcing

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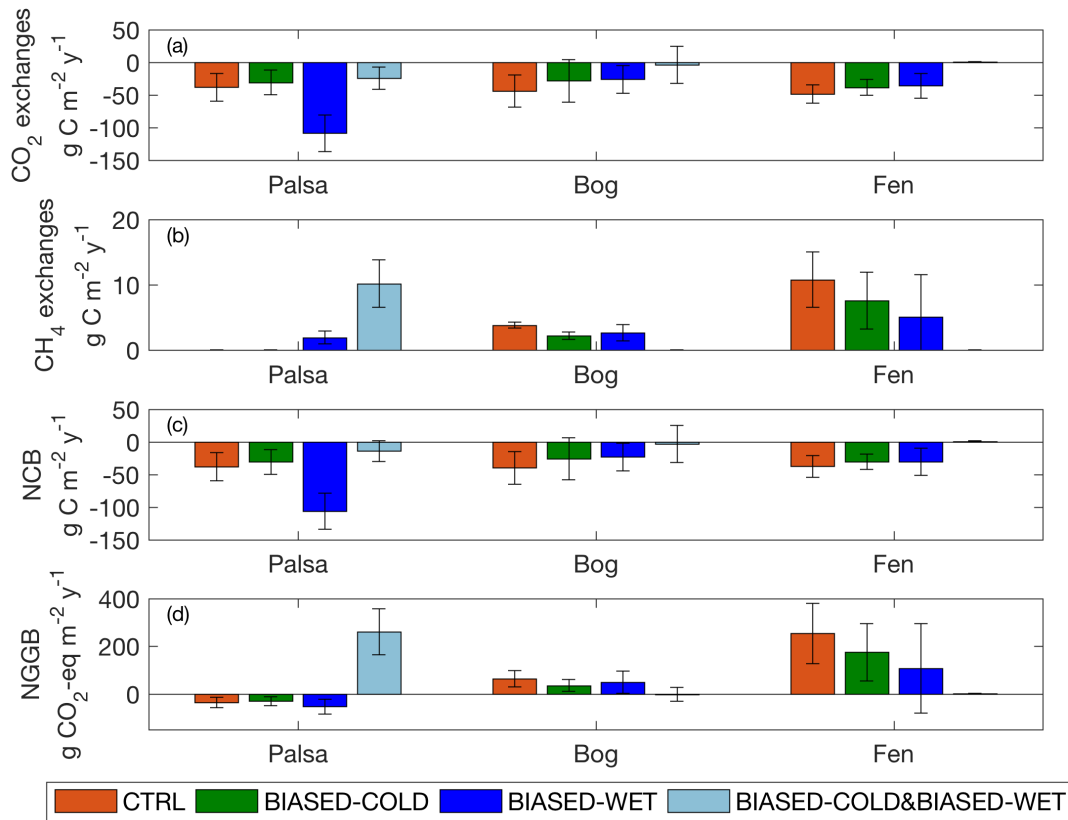
(Section 2.5). Bars and error bars are means and standard deviations calculated from

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2003 to 2007, respectively.

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1034 Figure 9. Annual CO₂ exchanges (a), CH₄ exchanges (b), Net Carbon Balance (c), and
 1035 Net Greenhouse Gas Balance (d) simulated at the palsa, bog, and fen, under each set of
 1036 simulations. Bars and error bars are the means and standard deviations calculated from
 1037 2003 to 2007, respectively. Positive and negative values indicate effluxes from and
 1038 influxes to the site, respectively.

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