



The optical characteristics and sources of chromophoric dissolved organic matter (CDOM) in seasonal snow of northwestern China

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Abstract.

Chromophoric dissolved organic matter (CDOM) plays an important role in the global carbon cycle and energy budget. A field campaign was conducted across northwestern China from January to February 2012, and surface seasonal snow samples were

5 collected at 39 sites in Xinjiang and Qinghai provinces. Light-absorption measurements, fluorescence measurements and chemical analysis were conducted to investigate the optical properties and potential sources of CDOM in seasonal snow. The abundance of CDOM (the absorption coefficient at 280 nm, a_{280}) and the spectral slope from 275 to 295 nm ($S_{275-295}$) ranged from 0.15-10.57 m^{-1} and 0.0129-0.0389 nm^{-1} , respectively.

10 The highest average a_{280} ($2.30 \pm 0.52 \text{ m}^{-1}$) and lowest average $S_{275-295}$ ($0.0188 \pm 0.0015 \text{ nm}^{-1}$) in Qinghai indicated that the snow CDOM in this region had strongly terrestrial characteristic. Relatively low regional average a_{280} values were found in sites located to the north of the Tianshan Mountains and northwestern Xinjiang along the border of China ($0.93 \pm 0.68 \text{ m}^{-1}$ and $0.80 \pm 0.62 \text{ m}^{-1}$, respectively). Parallel factor

15 analysis (PARAFAC) identified three types of chromophores that were attributed to two humic-like substances (HULIS, C1 and C2) and one protein-like material (C3). C1 was mainly from soil HULIS, while the potential sources of C2 were complex and included soil, microbial activities, anthropogenic pollution and biomass burning. The good relationship between a_{280} and the intensity of C1 ($R^2 = 0.938$, $p < 0.001$) indicated that

20 the CDOM abundance in the surface snow across northwestern China was mainly controlled by terrestrial sources. In addition, the regional variations of sources for CDOM in snow were further assessed by the analysis of chemical species (e.g., soluble



ions) and air mass backward trajectories combined with satellite active fire locations.



1 Introduction

Dissolved organic matter (DOM) is widely distributed in natural aquatic ecosystems and plays a key role in the global carbon cycle (Massicotte et al., 2017). Chromophoric dissolved organic matter (CDOM), the light-absorbing constituent of DOM (Helms et al., 2008), can absorb light from ultraviolet to visible (UV-vis) wavelengths (Stedmon et al., 2000). CDOM in aquatic ecosystems originates from the microbial decomposition of plant matter and is released by organisms within the ecosystems (autochthonous), as well as is imported from surrounding terrestrial environments (allochthonous) (Yao et al., 2011; Zhang et al., 2010). Due to its light-absorbing properties, CDOM is important in biological processes (Seekell et al., 2015; Thrane et al., 2014), photochemical processes (Helms et al., 2013; Vachetalo and Wetzel, 2004) and the energy budget (Hill and Zimmerman, 2016; Pegau, 2002) in natural water bodies.

The light-absorbing properties, composition, and sources of CDOM in lakes, rivers, wetlands, coasts, estuaries and ocean have been studied all over the world (Andrew et al., 2013; Chen and Jaffe, 2014; Coble, 1996; Organelli et al., 2014; Shao et al., 2016; Stedmon et al., 2011; Wang et al., 2014). Massicotte et al. (2017) summarized the distribution of DOM concentrations, CDOM absorption and their relationship based on more than 12000 measurements across rivers, lakes and oceans globally. Their study showed that the CDOM from wetlands has the highest absorption coefficients at 350 nm ($a_{\text{CDOM}(350)}$) with a median value of 87.2 m^{-1} , while the median $a_{\text{CDOM}(350)}$ of oceans was 0.08 m^{-1} and CDOM absorption decreased from fresh water to oceans. A



strong logarithmic relationship between DOC concentrations and $a_{\text{CDOM}}(350)$ ($n = 12808$, $R^2 = 0.92$) was also found.

Compared to the aquatic ecosystems, there were far fewer studies evaluating CDOM in the cryosphere (Barker et al., 2009; Dubnick et al., 2010; Feng et al., 2016; Norman et al., 2011). However, the global glacier ecosystem is a large organic carbon pool and exports approximately $1.04 \pm 0.18 \text{ TgC yr}^{-1}$ of DOC into freshwater and marine ecosystems (Hood et al., 2015). Dubnick et al. (2010) separated the CDOM of ice samples from several glacial environments into five components, namely, four protein-like components and one humic-like component, using fluorescence spectrometry. According to the results of a field experiment conducted in Canada, biological fluorophores represented the main species of CDOM in subglacially routed meltwater (Barker et al., 2009). Feng et al. (2016) found that the CDOM in cryoconite samples in Tibetan Plateau glaciers was mainly derived from microbial activity using light absorption and fluorescence measurements. In general, CDOM in glacial environments shows high bioavailability (Hood et al., 2015).

However, to the best of our knowledge, there is no study focusing on the CDOM in seasonal snow in midlatitude regions. As the largest component of the terrestrial cryosphere (Brutel-Vuilmet et al., 2013), seasonal snow can cover up to 40% of Earth's surface during the Northern Hemisphere winter (Hall et al., 1995). For its ecological functions, snowfall is an important carbon and nutrient input for land ecosystems (Mladenov et al., 2012) and a crucial freshwater reservoir (Jones, 1999), especially for barren regions (e.g., northwestern China). In addition, snowpack is also an active field



for photochemical (Beine et al., 2011; Domine et al., 2013) and biological processes (Liu et al., 2009; Lutz et al., 2016). Unlike aquatic environments, the high surface albedo is the most obvious physical property of snow (IPCC, 2013). Once light-absorbing impurities are deposited on the snow surface, the albedo can be significantly
5 reduced, and the regional and global climate are further affected (Hadley and Kirchstetter, 2012). Several field campaigns covering the Arctic, Russia, North America and northern China have been conducted to measure insoluble light-absorbing particles (ILAPs) in seasonal snow (Doherty et al., 2010, 2014, 2015; Huang et al., 2011; Pu et al., 2017; Wang et al., 2013, 2015, 2017; Warren and Wiscombe, 1980; Ye et al., 2012;
10 Zhou et al., 2017a). However, these studies neglected CDOM, which is also an effective light absorber, whether in the atmosphere (i.e., brown carbon, BrC) or water bodies. However, Dang and Hegg (2014) suggested that the absorption coefficients of soluble chromophores in Arctic snow are lower than those of black carbon (BC) by several orders of magnitudes at visible wavelengths. Compared to the Arctic, the results might
15 be different in seasonal snow across northwestern China due to the much higher level of anthropogenic activity and dry deposition of local soil there (Pu et al., 2017).

UV-vis absorption spectral analysis is a rapid and effective method of characterizing the optical properties and sources of CDOM. The abundance of CDOM is usually described by the light absorption of DOM at certain wavelengths within the UV band,
20 for instance, 280 nm (Zhang et al., 2011). The absorption spectrum of CDOM decreases approximately exponentially with increasing wavelength (Helms et al., 2008), and the spectral slope is used to describe the wavelength dependence of absorption



(Twardowski et al., 2004). Helms et al. (2008) used the spectral slope between 275 and 295 nm ($S_{275-295}$) and the ratio of the spectral slopes (S_R) of two narrow bands (275-295 nm and 350-440 nm) to investigate the molecular weight and sources of CDOM (terrestrial or marine origin). Then, $S_{275-295}$ was used as a tracer of terrigenous DOC import to marine environments and applied to the Arctic Ocean by remote sensing (Fichot and Benner, 2012; Fichot et al., 2013). Fichot and Benner (2011) presented a method to retrieve the DOM concentration using $S_{275-295}$ for northern Gulf of Mexico marine water.

The fluorescence excitation-emission matrix (EEM) offers multiple advantages, such as high sensitivity, easy sample preparation and non-destructivity to samples (Birdwell and Valsaraj, 2010), and has been widely used to identify the source and composition (humic-like or protein-like) of CDOM in natural waterbodies (Birdwell and Engel, 2010; Coble, 1996; Zhao et al., 2016), rainwater (Zhou et al., 2017b), fog water (Birdwell and Valsaraj, 2010) and aerosols (Chen et al., 2016a, 2016b; Fu et al., 2015). However, the CDOM in nature is composed of complicated chromophores and shows overlapping signals with EEMs. To precisely extract the useful information from the very large dataset of EEMs, Stedmon et al. (2003) successfully applied parallel factor analysis (PARAFAC) to decompose the EEMs into several independent fluorescent components. Due to the great advantage of PARAFAC in interpreting the results of EEMs, this has been a “mainstream” approach in recent natural CDOM studies (Murphy et al., 2013). In addition, three fluorescence-derived indices are widely used to identify the potential sources of CDOM. Zsolnay et al. (1999) presented a



humification index (HIX) to describe the relative humification of DOM. The fluorescence index (FI) is used to identify the sources of DOM from terrestrial or microbial origins (McKnight et al., 2001), and the biological index (BIX) can be an indicator of autochthonous productivity (Huguet et al., 2009).

5 In this study, surface snow samples collected across northwestern China were subjected to UV-vis absorption, fluorescence and chemical analyses. For the first time, with the aim of presenting an understanding of CDOM in seasonal snow, the abundances, optical properties and potential sources of CDOM and their spatial distributions were investigated.

10 2 Material and methods

2.1 Sample collection

During January to February 2012, snow samples were collected at 7 sites (no. 47- 51a, 51b, 52) in Qinghai and 32 sites (no. 53-84) in Xinjiang, which are located in northwestern China. The distribution of sample sites, which are numbered
15 chronologically, is shown in Fig. 1. The sample sites were separated into five regions to investigate the spatial variations of CDOM abundance, optical properties and their potential sources. One group of sites was in Qinghai, and four groups were in Xinjiang, following the grouping scheme presented by Pu et al. (2017).

The sample sites were chosen to be upwind and far enough away from roads, railways,
20 cities and villages to minimize the effects of local pollution. Hence, the collected samples can be representative of a wide range of areas. Pictures of some of the sample



sites are shown in Fig. 2. At each site, snow samples were collected every 5 cm from top to bottom, and if there was a melt layer or fresh snow on the top layer, such a sample was collected individually. A pair of two adjacent vertical profiles of snow were gathered (“left” and “right”) for assessing the variability of the same snowpack and to
5 enhance the accuracy of the measurements. During this campaign, 13 fresh snow samples that had fallen during the sampling time were collected. In addition, at some sites, the snow was thin and patchy and the wind was strong; hence, the samples were gathered from snow drifts. These samples were potentially influenced by the deposition of local soil (Ye et al., 2012). After being returned to the laboratory in Lanzhou
10 University, all the samples were stored in a freezer at -20°C or lower for subsequent analysis. More details on the sampling methods have been reported previously (Doherty et al., 2010; Wang et al., 2013; Ye et al., 2012).

2.2 Fluorescence measurement

The snow water samples were first filtrated using $0.22\ \mu\text{m}$ PTFE syringe filters (Jinteng,
15 Tianjin, China), and the filtrates were stored in prebaked glass vials ($450\ ^{\circ}\text{C}$ for 4 h) at $4\ ^{\circ}\text{C}$ until measurement within 24 h. The ultrapure water ($18.2\ \text{M}\Omega\text{-cm}$) filtrated by the PTFE syringe filters exhibited no clear fluorescence signal.

The EEMs of surface snow samples were determined by an Aqualog spectrofluorometer system (Horiba Scientific, NJ, USA) in a 1 cm quartz cell. The
20 scanning ranges were 240 to 600 nm in 5 nm intervals for excitation and 250 to 825 nm in 4.65 nm (8 pixel) intervals for emission, and the integrating time was 5 s. An



ultrapure water blank was subtracted to remove the water Raman scatter peaks.

The inner filter effect (IFE) of EEM was corrected using the method shown in Kothawala et al. (2013). The fluorescence intensities were calibrated by the Raman peak of the ultrapure water reference at a 350 nm excitation wavelength following the method presented by Lawaetz and Stedmon (2009). The Rayleigh scatter peaks of EEMs were addressed by the EEMscat MATLAB toolbox (version 3) (Bahram et al., 2006) using an interpolation algorithm.

The PARAFAC analysis was performed by the DOMFluor toolbox (version 1.7) in MATLAB (Stedmon and Bro, 2008). In addition, the emission wavelengths longer than 650 nm were removed to eliminate the uncertainty of measurement. According to the analysis of residual error and visual inspection, the three-component PARAFAC model was selected. The residual error decreased distinctly when the components increased from two to three (Fig. S1). The three-component model was also validated by the split-half analysis with the “S₄C₄T₂” split scheme (Murphy et al., 2013) (Fig. S2). The fluorescence intensity of the fluorescence component was expressed as F_{\max} in Raman unit (RU) (Stedmon and Markager, 2005a). In addition, the fluorescence-derived indices were calculated by Eq. (1-3) (Huguet et al., 2009; McKnight et al., 2001; Zsolnay et al., 1999):

$$FI = I_{370}^{450} / I_{370}^{499}, \quad (1)$$

$$BIX = I_{310}^{379} / I_{310}^{430}, \quad (2)$$

$$HIX = I_{255}^{434-480} / I_{255}^{300-345}, \quad (3)$$

where I is the fluorescence intensity, the subscript denotes the excitation wavelength



and the superscript denotes the corresponding emission wavelength or wavelength range. We note that the wavelengths used in the calculation were changed slightly (1 nm or less) due to different instruments.

2.3 UV-vis absorption measurement

5 The UV-vis absorption spectra of snow samples were derived from 240 to 600 nm in 5 nm intervals while the fluorescence measurements were conducted by an Aqualog spectrofluorometer system (Horiba Scientific, NJ, USA), and an ultrapure water blank was used as a reference. The absorbance of CDOM was assumed to be zero above 550 nm, and the average absorbance between 550-600 nm was subtracted from the whole
10 spectrum to correct the baseline shifts and scattering effects of the measurement. The absorbances of the samples were converted to absorption coefficients using the following equation:

$$a(\lambda) = \ln(10) \cdot A(\lambda) / L, \quad (4)$$

where A is the absorbance of the sample, λ is the wavelength, L is the path length of
15 cuvette (0.01 m), and a is the absorption coefficient (m^{-1}). The abundance of CDOM is presented by the absorption coefficients at 280 nm (a_{280}) (Zhang et al., 2010). The spectral slope between 275-295 nm ($S_{275-295}$) was determined by a linear regression between logarithmic absorbance and wavelengths since the correlation coefficients were higher than that of the exponential fit. If the difference in $S_{275-295}$ between the
20 linear and exponential regressions was higher than 10%, indicating a high uncertainty for absorption measurement, such data were removed. The absorption Ångström



exponents (AAEs, from 240 to 550 nm) of CDOM were calculated using power-law fit from Eq. (5), as follows:

$$a(\lambda) = K \cdot \lambda^{-AAE}, \quad (5)$$

where a is the absorption coefficient (m^{-1}), K is a constant and λ is the wavelength.

5 Because the light absorption within the visible wavelengths of some samples were too low to be measured by the spectrometer, 19 of 39 samples were available for the calculation of AAE.

Furthermore, the “left” samples of site 51b and 58 showed abnormal absorption and fluorescence spectra compared to the other samples, and were supposed to be
10 contaminated, and hence, these two samples were not used in the absorption and fluorescence analyses.

2.4 Soluble ions

The major soluble ions of surface snow water samples were analyzed with an ion chromatograph (Dionex, Sunnyvale, CA, USA) using an AS11 column for the anions
15 SO_4^{2-} , NO_3^- , Cl^- , and F^- and a CS12 column for the cations Na^+ , K^+ , Ca^{2+} , Mg^{2+} , and NH_4^+ . The soluble ions showed no obvious differences between filtered and unfiltered samples (Pu et al., 2017). In addition, the concentration of non-sea-salt K^+ (nss- K^+) was corrected by Mg^{2+} (a sea salt indicator) using the following formula (Tao et al., 2016):

$$20 \quad nss-K^+ = K^+ - 0.159Mg^{2+}. \quad (6)$$



2.5 Hierarchical cluster analysis

A hierarchical cluster analysis was used to classify the CDOM in snow based on the relative abundances of PARAFAC components. Euclidean distance was used to estimate the distances between samples, and the unweighted average distance was
5 chosen for the hierarchical cluster analysis by comparing the multiple correlation coefficients for the weighted average, centroid, farthest neighbor, shortest neighbor, weighted center of mass and Ward's distance. A total of four clusters were determined and labeled clusters A-D.

2.6 Active fire data and air mass backward trajectories

10 Active fire data (MCD14DL) from Moderate Resolution Imaging Spectroradiometer (MODIS) Collection 6 were used to capture potential fire source distributions. The data are available online: <http://earthdata.nasa.gov/firms>. To further analyze the potential sources of CDOM, 72-h air mass backward trajectories were conducted by the HYbrid
Single-Particle Lagrangian Integrated Trajectory (HYSPLIT,
15 <http://ready.arl.noaa.gov/HYSPLIT.php>) model (version 4). The model was run at 500 m above ground level four times a day for a period of 30 days preceding the sampling date at a given site.

3 Results and discussion

3.1 The absorption characteristics of CDOM (a_{280} , $S_{275-295}$ and AAE)

20 The distributions of a_{280} and $S_{275-295}$ are shown in Fig. 3 and the corresponding



values are summarized in Table 1. a_{280} ranged widely from 0.15 to 10.57 m^{-1} with an average of $1.69 \pm 1.80 \text{ m}^{-1}$. The highest value appeared at site 67 (10.57 m^{-1}), followed by sites 53, 79 and 47 (5.25 m^{-1} , 3.13 m^{-1} and 3.11 m^{-1} , respectively). Most of these samples were collected from the snow drifts under strong wind. These values were

5 higher than the a_{280} of snow, ice and cryoconite CDOM on the Tibetan Plateau (typically lower than 2.0 m^{-1} ; Feng et al., 2016, 2017). The lowest value was found at site 66 (0.15 m^{-1}), followed by sites 70, 82, 73 and 83 (0.21 m^{-1} , 0.23 m^{-1} , 0.30 m^{-1} and 0.31 m^{-1} , respectively), and these values were compared to the absorption of soluble light-absorbing species in snow at Alaska with typical values of 0.1-0.15 m^{-1} at 250 nm

10 (Beine et al., 2011). Some of these samples comprised freshly fallen snow and some were collected at remote sites in northern Xinjiang, which were far from pollution sources (Pu et al., 2017). The values of $S_{275-295}$ ranged from 0.0129 to 0.0389 nm^{-1} with an average of $0.0243 \pm 0.0073 \text{ nm}^{-1}$. Hansen et al. (2016) summarized the values of $S_{275-295}$ for oceanic and terrestrial systems, and the values range from 0.020-0.030

15 nm^{-1} for ocean, 0.010-0.020 nm^{-1} for coastal water, and 0.012-0.023 nm^{-1} for terrestrial river systems. The wide range of $S_{275-295}$ indicates complex sources of CDOM in seasonal snow across northwestern China. Furthermore, as shown in Fig. 4, $S_{275-295}$ decreases logarithmically with increasing a_{280} ($R^2 = 0.586$, $p < 0.001$). As presented by Helms et al. (2008), lower $S_{275-295}$ values correspond to terrestrial and higher

20 molecular weight CDOM. Therefore, the higher a_{280} values of our samples may be due to greater inputs of terrestrially derived CDOM with high molecular weights, such as HULIS (Zhou et al., 2013). Similar negative relationships between absorption



coefficients and $S_{275-295}$ was also found in estuarine (Asmala et al., 2012) and marine water (Zhou et al., 2013). The AAEs of 19 CDOM samples are also shown in Table 1. The values ranged from 4.41–8.91 with an average of 5.55 ± 1.11 . This value is comparable with the average AAE of HULIS (6.11, from 300 to 550 nm), the major CDOM species in snow (Beine et al., 2011), extracted from Alaskan snow (Voisin et al., 2012).

The detailed results of each region are discussed below. Region 1 (sites 47-52) is located in the eastern Tibetan Plateau in Qinghai province, which is typically higher than 4000 m above sea level. In this region, the snowpack was usually patchy and thin (Fig. 2a). During windy time, local soil can be blown and deposited on the snow surface, this had been observed by previous studies (Pu et al., 2017; Ye et al., 2012), and the filters for samples in this region were in yellow color due to high loading of soil. The average a_{280} was highest among all five regions ($2.30 \pm 0.52 \text{ m}^{-1}$), and the $S_{275-295}$ fell in the range of 0.0170–0.0212 nm^{-1} (mean: $0.0188 \pm 0.0015 \text{ nm}^{-1}$). The lowest regional average and slight variation of $S_{275-295}$ clearly indicates the dominant contribution of a terrestrial source (e.g., local soil) to snow CDOM (Fichot and Benner, 2012; Helms et al., 2008).

Region 2 (sites 53-59, 61, and 79) is along the Tianshan Mountains. Within this region, some sites were farmland (e.g., sites 55 and 56) (Fig. 2c), which can be influenced by local agriculture activities (e.g., crop and grass burning). The snow at some other sites (e.g., sites 53, 57 and 79) was patchy (Fig. 2b), and blowing soil have been an important source of CDOM. Thus, the a_{280} values of these sites (1.66-5.25 m^{-1}



¹) were higher than those at the other sites (0.41-0.54 m⁻¹) with high snow coverage rates, new fallen snow or long distances from human activities (high altitude) (e.g., sites 54, 58 and 59). Overall, region 2 showed a high average a_{280} (2.00±1.50 m⁻¹), and the average $S_{275-295}$ was 0.0229±0.0073 nm⁻¹.

5 Region 3 (sites 60, 62, 63 and 80-84) is located to the north of the Tianshan Mountains and close to the industrial city belt in central Xinjiang. Hence, human activities may dominate the contribution of CDOM in snow in this region. However, the a_{280} values were mostly less than 1.0 m⁻¹ except for at sites 60 and 84, and the average value was low (0.93±0.68 m⁻¹). Because samples of these sites were almost
10 new fallen snow, the deposition of pollutants to the snowpack can be quite slight. Sites 60 and 84 were both close to industrial cities (Fig. 1 in Pu et al., 2017), and the transportation and deposition of anthropogenic pollutants may be responsible for the high a_{280} (2.39 m⁻¹ and 1.65 m⁻¹, respectively). The average $S_{275-295}$ was 0.0218±0.0057 nm⁻¹ and was comparable to that in region 2.

15 Region 4 (sites 64-71) is located in northwestern Xinjiang. In this region, the snow depth increased from south to north. At sites 64-67, the snow was thin due to the low snowfall in sample year (Ye et al., 2012). The maximal and minimal a_{280} values of the entire campaign, 10.57 m⁻¹ (site 67, snow drift, Fig. 2e) and 0.15 m⁻¹ (site 66, new snow), were found in this area. Sites 68-71, which were the northmost sites during this
20 campaign, were far from industrial areas, and the snow depth and coverage rate were high, which led to low a_{280} values. Generally, the a_{280} was in the range of 0.5-2.0 m⁻¹ with an average of 0.80±0.62 m⁻¹ (excluded site 67), which was the lowest compared



to the values in the other four regions. The mean value of $S_{275-295}$ ($0.0255 \pm 0.0060 \text{ nm}^{-1}$) was higher than that in regions 2 and 3. The low $S_{275-295}$ of site 67 (0.0169 nm^{-1}) suggested a strongly terrestrial input of CDOM.

Region 5 (sites 72-78) is located in northeastern Xinjiang along the border of China. The average value of a_{280} was $1.17 \pm 0.63 \text{ m}^{-1}$, which was intermediate among five regions. However, the $S_{275-295}$ was typically higher than 0.0300 nm^{-1} with an average of $0.0324 \pm 0.0060 \text{ nm}^{-1}$. These values indicated that non-terrestrial sources were dominant for snow CDOM in this region.

3.2 The fluorescence characteristics of CDOM

3.2.1 PARAFAC components

The EEMs of snow water samples were analyzed by PARAFAC, and three separate fluorescent components were identified (Fig. 5a-c). Figure 5d-f present the corresponding excitation and emission loading spectra of each component. The excitation/emission (Ex/Em) wavelengths of each component's fluorescence peaks are summarized in Table 2. The spectral characteristics of three components in this study are similar to the fluorophores identified in aquatic environments in previous studies (Stedmon and Markager, 2005a; Stedmon et al., 2003; Zhang et al., 2010, 2011).

C1 showed a primary peak at $<240/453 \text{ nm}$ for Ex/Em, which was similar to the component 1 reported by Stedmon and Markager (2005a). This kind of fluorophore absorbs light mainly in the UVC band and shows a broad emission peak, and is usually identified as terrestrial CDOM (Stedmon et al., 2003). However, in our study, a



secondary peak was also presented (Ex/Em = 305/453 nm). The appearance of an
excitation peak at longer wavelengths may indicate that C1 is more aromatic and has
higher molecular weight (Coble et al., 1998). C1 also resembled another terrestrial
fluorophore, namely, component 4 in Stedmon and Markager (2005a) (Ex/Em = <250,
5 360/440), which has been widely found in nature fresh water environments and even
water-extracted organic matter in aerosols (Chen et al., 2016b; Mladenov et al., 2011;
Zhang et al., 2009; Zhao et al., 2016). C1 in our study showed blue-shifted excitation
wavelengths, which likely corresponded to a lower molecular weight (Coble et al.,
1998).

10 C2 had a primary (secondary) peak at <240(300)/393 nm (Ex/Em), namely, peak M,
that was first measured in the oceanic system by Coble (1996). Hence, this component
is usually identified as marine HULIS. However, the marine sources of CDOM can be
nearly ignored in this study because the sample sites were several thousand kilometers
away from the ocean. Stedmon et al. (2003) found a similar fluorophore (component 4
15 therein) that showed terrestrial source in a terrestrially dominated estuary region. The
subsequent studies suggested that the C2-like component is also linked to microbial
activity and phytoplankton degradation in natural aquatic systems (Yamashita et al.,
2008; Zhang et al., 2009) or DOM in wastewater from anthropogenic sources (Stedmon
and Markager, 2005a).

20 C3 is a typical fluorophore that is categorized as tyrosine-like CDOM and that
exhibits Ex/Em pairs of <240(270)/315 nm. C3 reflects autochthonously labile CDOM
produced by biological processes (Stedmon et al., 2003) and has been commonly



reported in previous studies of natural water bodies and the water extraction of aerosols (Chen et al., 2016b; Murphy et al., 2008; Stedmon and Markager, 2005b).

To analyze the mutual relationships between three components, linear correlation analyses were conducted (Fig. 6). The F_{\max} values of C3 at site 67 were much higher than those of any other sample (shown as red markers in Fig. S3), which can strongly influence the results of the correlation analysis. When excluding the data of site 67, the R^2 between $F_{\max}(C1)$ and $F_{\max}(C3)$ fell from 0.316 to 0.082, and the linear relationship became nonsignificant (Fig. S3). Therefore, we chose to use the dataset that excludes site 67 in the analysis, and the results are shown below. $F_{\max}(C1)$ and $F_{\max}(C2)$ exhibited a significant and positive correlation ($R^2 = 0.332$, $p < 0.001$); however, this value was much lower than that in previous studies of natural water, for instance, $R^2 = 0.63$ for inland lakes (Zhao et al., 2016) and $R^2 = 0.88$ for inland rivers (Zhang et al., 2011). This result indicated that two humic-like components might only share partially common sources. The relationships between the two humic-like components and the protein-like component are quite different. Not surprisingly, $F_{\max}(C1)$ and $F_{\max}(C3)$ showed no correlation ($R^2 = 0.082$, $p > 0.05$), while a significant linear relationship ($R^2 = 0.364$, $p < 0.001$) was found between $F_{\max}(C2)$ and $F_{\max}(C3)$, which implied a potential microbial source for C2, consistent with the finding of Yamashita et al. (2008).

3.2.2 Regional variation in PARAFAC components

Figure 7 shows the relative intensities of the three fluorescence components at each site, and Table 3 exhibits the regional average of the relative intensity and F_{\max} for each



component. Overall, C2 was the most intense fluorophore and accounted for 42% on average of the total fluorescence intensity of all samples, followed by C3 (38% on average) and C1 (20% on average). Compared to glacial snow and ice samples, which were dominated by protein-like substances (Dubnick et al., 2010; Feng et al., 2016), the
5 seasonal snow samples in this study showed fewer microbial characteristics.

In Qinghai (region 1), the most obvious feature was that C1 accounted for approximately 35% of the total fluorescence intensity. This value was significantly higher than the other regions. As mentioned in Sect. 3.1, the snowpack of these sites was strongly influenced by the local soil, hence clearly implying that C1 was mainly
10 from the soil HULIS. This finding confirmed the constantly terrestrial source of the C1-like fluorophore, regardless of whether the sample was collected from natural water bodies, aerosol water extraction or snow. The relative contribution of C3 (the protein-like component) in region 1 was much lower than that in other regions. This result was mainly due to the high $F_{\max}(C1)$ in region 1 since the regional variation of $F_{\max}(C3)$ was
15 slight (Table 3).

In Xinjiang, the relative intensity of C1 varied by region, while C2 and C3 accounted for roughly equal contributions. In region 2, the relative contribution of C1 was high (25% on average), which implied a substantial amount of CDOM originating from local soil. In region 3, where most of the samples were new fallen snow (7 of 8 sites), the
20 CDOM showed much fewer terrestrial characteristics as demonstrated by the lowest relative intensity of C1 (9% on average). The great difference between relative contributions of C1 and C2 in this region indicated that the sources of these two humic-



like components were different. In regions 4 and 5, the contributions of C1 (both were approximately 17%) were nearly double that in region 3. As the snow samples collected in northern Xinjiang were more aged than those in region 3, the deposition of soil was more significant, which led to a higher C1 fraction.

5 At site 54 and 82, the fractional intensity of C3 exceeded 70%. This value was approximately twofold higher than the average of all samples. This result can be explained by two possible reasons, (1) lower inputs of C1 and C2 were present at these sites, and (2) much greater microbial activities were available in the snowpack at sites 54 and 82. We found algal communities near the sample sites (Fig. S4), providing
10 evidence for the latter reason.

At site 67, the fluorescence intensities of C1, C2 and C3 were all highest among all samples (0.30 RU, 0.39 RU and 0.38 RU, respectively), especially for C3. The average intensity of C3 for samples excluding site 67 was 0.10 RU, with a low standard deviation of 0.02 RU, and this value was approximately one-fourth of that at site 67.
15 Therefore, rather than owing to microbial activity alone, the extremely high $F_{\max}(C3)$ of site 67 may be due to other sources, for instance, some organic compounds released from diesel combustion may show similar spectra (Mladenov et al., 2011).

To assess the similarities and differences between samples from different regions, a hierarchical cluster analysis based on the relative intensities of three fluorescence
20 components was conducted (Fig. 8). The snow samples were separated into four clusters (clusters A-D) (Fig. S5). Samples classified into clusters A and B were dominant. The high relative contribution of C1, which was 34% on average, was the most remarkable



feature of cluster A and led to a low relative abundance of C3 (26% on average) (Table S1). All samples in region 1 and most samples in region 2 were assigned to cluster A. For cluster B, the contribution of C1 was low (13% on average), and the abundance of C3 (47% on average) was slightly higher than that of C2 (40% on average). For sites in
5 northern Xinjiang (regions 4 and 5), most samples were classified into cluster B. The samples assigned to cluster C, including those of sites 60, 62, 69, 72, 76 and 84, showed dominant contribution of C2 (57% on average). Half of these samples were found in region 3, and the others were dispersed in regions 4 and 5. Pu et al. (2017) analyzed the concentrations and sources of ILAPs in snow in the same field campaign. They found
10 that the dominant source of ILAPs at these sites was industrial pollution. This result indicates that the humic-like fluorophore C2 may partly originate from anthropogenic pollution. Cluster D contained only two samples from sites 54 and 82. The difference between cluster D and others was an extremely high relative contribution of protein-like component C3 (73% on average), which indicated the high bioavailability of snow
15 CDOM there.

3.2.3 Fluorescence-derived indices

The values of the three established fluorescence-derived indices for surface snow samples are shown in Fig. 9. The regional average values are shown in Table 4. The HIX of samples in this study fell into the range of 0.16-3.20 with an average of $1.21 \pm$
20 0.78. The highest average HIX appeared in region 1 (2.21 ± 0.42), which was highly consistent with the low $S_{275-295}$ and high relative intensity of C1 and further



demonstrated the large terrestrial contribution in this region. The lowest average HIX was found in region 3 (0.62 ± 0.37 on average), which suggests that the CDOM was fresh. This finding is easily explained as the snow samples in this region were nearly from new fallen snow. Compared to the HIX of other types of samples (Table 5), the HIX of snow across northwestern China was higher than that of spring water (Birdwell and Engel, 2010); comparable to cryoconite in glaciers from the Tibetan Plateau (Feng et al., 2016), inland lakes (Zhang et al., 2010), cave water (Birdwell and Engel, 2010) and north Pacific Ocean water (Helms et al., 2013); and was lower than estuarine water (Huguet et al., 2009), fog water (Birdwell and Valsaraj, 2010), groundwater (Huang et al., 2015), water extraction of alpine aerosol (Xie et al., 2016) and urban aerosol (Mladenov et al., 2011).

According to McKnight et al. (2001) and Huguet et al. (2009), the values of BIX > 1.0 or FI > 1.9 indicate microbially derived DOM. The values of BIX and FI for the snow samples were typically below 1.0 and 1.9, respectively, implying unremarkably autochthonous characteristics. The regional distributions of BIX and FI corresponded with that of HIX. The samples with highest average BIX and FI were in region 3 (0.93 ± 0.25 and 1.60 ± 0.15 , respectively), and the samples in region 1 exhibited the lowest average values of BIX and FI (0.49 ± 0.05 and 1.29 ± 0.05 , respectively). The BIX and FI of different types of samples changed little, and the only exception was the FI of cryoconite in glaciers from the Tibetan Plateau (3.24, Feng et al., 2016), which was approximately twice as high as those of the other samples.



3.3 Correlation between light absorption and fluorescence properties.

The relationships between the a_{280} and F_{\max} of each component and the fluorescence-derived indices are shown in Fig. 10. Similar to Fig. 6, we used the dataset excluding site 67 in the analysis, and the comparisons with the results using the entire dataset are shown in Fig. S6. The $F_{\max}(C1)$ exhibited a strong and positive linear relationship with a_{280} ($R^2 = 0.938$, $p < 0.001$), while the correlation coefficients between a_{280} and $F_{\max}(C2)$ ($R^2 = 0.464$, $p < 0.001$), $F_{\max}(C3)$ ($R^2 = 0.152$, $p < 0.05$) were much lower. These results suggested that the variation of CDOM in northwestern China seasonal snow was mainly controlled by terrestrially derived HULIS. This finding is different from those results found for inland lakes and rivers (Zhang et al., 2010, 2011; Zhao et al., 2016), in which the correlations between CDOM absorption and C1- and C2-like fluorophores were comparable and both strong.

The relationships between a_{280} and the three fluorescence-derived indices are also shown (Fig. 10d-f). The results showed considerable similarity to that between a_{280} and F_{\max} . a_{280} and HIX exhibited a significant and positive correlation ($R^2 = 0.828$, $p < 0.001$). This result suggested that the CDOM abundance in northwestern China seasonal snow increases with the degree of humification. Negative log-log correlations were found between a_{280} and BIX ($R^2 = 0.269$, $p < 0.001$), FI ($R^2 = 0.547$, $p < 0.001$), which indicates that a higher CDOM abundance corresponds to higher terrestrial inputs.

The AAE of CDOM also showed a significant correlation with the fluorescence-derived indices (Fig. S7). HIX was negatively and linearly correlated with AAE ($R^2 = 0.396$, $p < 0.005$), indicating that the AAE can decrease with the increasing humification



degree of snow CDOM, which can be characterized by the higher aromaticity (Bayer et al., 2000). This result was consistent with a study of BrC in aerosol from the Los Angeles Basin (Zhang et al., 2013). They suggested that compared to water-soluble BrC, the lower AAE of methanol-extracted BrC can be explained by the higher aromaticity of insoluble organic components. In contrast, positive correlations were found between AAE and BIX ($R^2 = 0.814$, $p < 0.001$) and FI ($R^2 = 0.463$, $p < 0.005$), which implied that the microbial-derived CDOM has a higher AAE than terrestrial CDOM.

3.4 Source distributions of CDOM

As discussed in Sect. 3.2.1, C2 shared partially common sources with C1 and C3, which represented soil and microbial origins, respectively. Furthermore, the $F_{\max}(C2)$ and $F_{\max}(C3)$ varied slightly among regions, while the variation of $F_{\max}(C1)$ was dramatic (Table 3). This phenomenon clearly reflected that the HULIS fluorophore C2 may originate from sources besides soil and microbial activity. To further analyze the potential sources of PARAFAC components, the correlation coefficients between three major ions and F_{\max} of fluorescence components were calculated. The results are shown in Table 6. $F_{\max}(C2)$ showed a significant and positive correlation with three ions ($p < 0.001$). The secondary ions SO_4^{2-} and NO_3^- are commonly considered as the markers of anthropogenic emissions from the burning of fossil fuel, such as oil and coal (Doherty et al., 2014; Oh et al., 2011; Pu et al., 2017), and $nss-K^+$ is widely used as a tracer of biomass burning (Tao et al., 2016). Therefore, C2 may also originate from anthropogenic pollution and biomass burning; in other words, there were two additional



potential sources of snow CDOM in this study. Since the contribution of microbial-derived C_3 to a_{280} was low (Fig. 10c), three major sources of snow CDOM were identified, namely, soil, biomass burning and anthropogenic pollution. The regional variations of CDOM sources are discussed below using analyses of chemical species
5 and air mass backward trajectories. In addition, the sources of CDOM in snow are also compared with those of particulate light absorption of ILAPs.

In Qinghai (region 1), the backward trajectories to a typical site (site 47, Fig. 11a) were mostly from the west, and very few fire locations were encountered. Combined with the low ratio of $(SO_4^{2-}+NO_3^-)/nss-K^+$ (Fig. 12), the CDOM produced by biomass
10 burning and anthropogenic pollution is negligible in region 1. As mentioned in Sect. 3.1, the snowpack in Qinghai was strongly influenced by local soil, and hence, the soil HULIS was clearly the primary source of CDOM. In region 2, the contribution of soil to CDOM was also significant, which was confirmed by the high relative intensity of C_1 , as discussed in Sect. 3.2.2. For the results of the backward trajectories, along the
15 paths of the air masses to site 55 in region 2 (Fig. 11b), the local fire spots due to the agricultural activities could also be important contributors to snow CDOM. Additionally, the value of $(SO_4^{2-}+NO_3^-)/nss-K^+$ was also low in region 2, which indicated an insignificant role of anthropogenic pollution. Therefore, in region 2, a mixed source of soil and biomass burning is reasonable. In region 3, the extremely high
20 ratio of $(SO_4^{2-}+NO_3^-)$ to $nss-K^+$ implied a strong contribution of anthropogenic pollution to snow CDOM. Additionally, the mass ratio of Cl^- and Na^+ (2.48, Fig. S8) was also significantly higher than that in sea water (1.18, Hara et al., 2004), which



indicated that Cl^- might originate from other sources in addition to sea salt, such as coal combustion (Wang H. L. et al., 2008; Wang Y. et al., 2006), while the values in other regions were comparable to 1.18. These results, again, confirmed that the CDOM from anthropogenic pollution was dominant in region 3 but inapparent in other regions. The backward trajectories also showed consistent results (Fig. 11c). Most of the trajectories to site 84 came from the northwest and passed through the cities with heavy industry (e.g., Karamay and Shihezi). Therefore, air pollutants can be transported to the sample area and deposited in the surface snow. In region 4, the ratios of $(\text{SO}_4^{2-} + \text{NO}_3^-)$ to nss-K^+ were much lower than those in region 3, and the air masses originated from Siberia (Fig. 11d and e), where significant crop and grass burning take place (Hegg et al., 2010), strongly influenced this region. In region 5, the value of $(\text{SO}_4^{2-} + \text{NO}_3^-)/\text{nss-K}^+$ was comparable with that in region 4, which suggested CDOM from biomass burning rather than pollution. Although the backward trajectories were more dispersive (Fig. 11f), the amount of air mass from the Siberia was also respectable which can explain this finding. Furthermore, the low mass ratios of Cl^- and Na^+ in region 4 and 5 also implied the slight influence of anthropogenic pollution. Overall, biomass burning was the dominant source of CDOM in snow in both regions 4 and 5.

Pu et al. (2017) identified the potential sources of particulate light absorption of ILAPs (denoted as $C_{\text{BC}}^{\text{max}}$) in snow during the same field campaign, using a positive matrix factorization (PMF) model. The comparison of $C_{\text{BC}}^{\text{max}}$ and a_{280} among regions is shown in Fig. S9. In regions 1 and 5, there were no correlation between $C_{\text{BC}}^{\text{max}}$ and a_{280} , which indicated entirely different sources of CDOM and particulate absorption of



ILAPs. As reported by Pu et al. (2017), the dominant sources of C_{BC}^{max} were biomass burning and industrial pollution in regions 1 and 5, respectively, while those of CDOM were soil and biomass burning in our study. Much higher correlation coefficients were found in regions 2-4 ($R^2 = 0.72, 0.95$ and 0.81 , respectively), especially for region 3, 5 which implied consistency for sources of CDOM and particulate absorption of ILAPs. The relative weaker relationship presented in region 2 may have resulted from the mixed sources of CDOM (soil and biomass burning); however, biomass burning was the primary source for particulate absorption of ILAPs.

3.5 Comparing the light absorption by CDOM and BC

10 Figure 13 shows the relative contributions of CDOM and BC to light absorption. As mentioned above, light absorption within visible wavelengths was available for 19 samples. The BC concentrations in surface snow were obtained from Pu et al. (2017), and the mass absorption coefficient (MAC) at 550 nm and the AAE of BC used in the calculation were $6.3 \text{ m}^2 \text{ g}^{-1}$ and 1.1, respectively (Pu et al., 2017).

15 The light absorption of CDOM was 0.02-1.17 times that for BC at 400 nm with an average of 0.34 ± 0.34 . At 500 nm, this value decreased quickly to 0.10 ± 0.11 on average and ranged from 0.005-0.40. The CDOM absorption at 400 nm was comparable to or even slightly higher than the absorption by BC at sites 50, 52 and 79. This finding is quite different from the results for Alaskan snow. Dang and Hegg (2014) converted the
20 CDOM absorption in snow at Barrow, Alaska, reported by Beine et al. (2011), into equivalent BC mixing ratios of 0.14 ng g^{-1} at 400 nm and 0.07 ng g^{-1} at 550 nm. As



presented by Doherty et al. (2013), the mixing ratio of BC in Barrow snow ranged from 10-30 ng g⁻¹. Hence, the absorption of CDOM in Alaskan snow can be safely ignored, but this does not appear reasonable for some sites across northwestern China.

Previous studies on impurities in seasonal snow have focused on insoluble particles (e.g., BC, insoluble OC and dust) (Doherty et al., 2010, 2014; Pu et al., 2017; Wang et al., 2013). The above discussion indicated that in some specific areas of northwestern China, the absorption of CDOM in snow was remarkable. What is the common feature of such sites? The average $S_{275-295}$ (0.0187 ± 0.0022) of these 19 sites was lowest compared to the five regional averages, and the average BIX (0.60 ± 0.20) and FI (1.31 ± 0.09) were lower than those of region 2, in which the influence of local soil was obvious; in addition, the average values of HIX (1.87 ± 0.57) and relative contribution of C1 ($30 \pm 8\%$) were higher than those of region 2. These results indicated that the CDOM of these sites was undoubtedly of terrestrial origin (e.g., wind-blown soil). Hence, we suggest that the absorption by CDOM in the snowpack, which is heavily affected by soil, cannot be ignored.

4 Conclusions

Surface snow samples was collected across northwestern China from January to February 2012. The a_{280} and $S_{275-295}$ of snow water samples ranged from 0.15-10.57 m⁻¹ and 0.0129-0.0389 nm⁻¹, respectively. The average value of a_{280} (1.69 ± 1.80 m⁻¹) was approximately 10 times higher than that of snow in Alaska (Beine et al., 2011). Samples in Qinghai (region 1) exhibited the highest average a_{280} (2.30 ± 0.52 m⁻¹) and



lowest average $S_{275-295}$ ($0.0188 \pm 0.0015 \text{ nm}^{-1}$) resulting from the strong influence of local soil. Low average a_{280} appeared in central Xinjiang (region 3, $0.93 \pm 0.68 \text{ m}^{-1}$), where the samples were nearly collected from new fallen snow, and northwestern Xinjiang (region 4, $0.80 \pm 0.62 \text{ m}^{-1}$ when excluded site 67) which was far from industrial areas. In the Tianshan Mountains (region 2) and northeastern Xinjiang (region 5), the average a_{280} values were $2.00 \pm 1.50 \text{ m}^{-1}$ and $1.16 \pm 0.63 \text{ m}^{-1}$, respectively. For all sites in Qinghai and some of the sites in Xinjiang (19 of 39 sites), the light absorption of CDOM cannot be neglected and even was remarkable at some sites compared to that of BC (0.34 ± 0.34 times relative to BC at 400 nm on average) due to the high contribution of soil to CDOM in snow. Hence, we suggest that the CDOM absorption at visible wavelengths at such sites should be taken into consideration in future studies.

Based on PARAFAC, two humic-like fluorophores (C1 and C2) and one protein-like fluorophore (C3) were identified. C2 showed partially common sources with terrestrially derived C1 ($R^2 = 0.332$, $p < 0.001$). The positive relationship between $F_{\max}(\text{C2})$ and $F_{\max}(\text{C3})$ ($R^2 = 0.364$, $p < 0.001$) indicated a potential microbial source of C2. In region 1, the relative intensity of C1 (35%) was much higher than that of the other regions, and combined with the highest HIX and lowest BIX and FI, indicated a clearly terrestrial origin of C1. In contrast, the contribution of C1 to total fluorescence was lowest in region 3 (9%), and the values of fluorescence-derived indices also showed the consistent results. The high correlations between a_{280} and $F_{\max}(\text{C1})$ ($R^2 = 0.938$, $p < 0.001$), HIX ($R^2 = 0.828$, $p < 0.001$) indicated that the CDOM abundance in surface snow across northwestern China was mainly controlled by the terrestrial sources. A



hierarchical cluster analysis was used to classify samples into four clusters (A-D) on the basis of the relative intensity of three fluorescent components. All samples in region 1 and most samples in region 2 were assigned to cluster A (a high contribution of C1). The number of samples assigned to cluster B (roughly equal contributions of C2 and C3) and cluster C (a dominant contribution of C2) were nearly even in region 3. For regions 4 and 5, most samples were classified into cluster B. Only two samples were assigned to cluster D due to the dominant contribution of C3.

According to the correlation analysis between three major ions and the F_{\max} of PARAFAC components, in addition to soil and microbial activity, C2 exhibited potential sources of anthropogenic pollution and biomass burning. Furthermore, the spatial variation of CDOM sources was assessed by using regional variations of $(\text{SO}_4^{2-} + \text{NO}_3^-) / \text{nss-K}^+$, $\text{Cl}^- / \text{Na}^+$ and air mass backward trajectory analysis. In regions 1 and 3-5, the dominant sources were soil, anthropogenic pollution, biomass burning and biomass burning, respectively, while there may be a soil-biomass-burning mixed source in region 2.

This study investigated the light absorption, fluorescence properties and potential sources of CDOM in seasonal snow across northwestern China. Future studies should focus on the molecular characteristics and the relationships between the optical properties of CDOM in snow. Understanding the structures, chemistry, and sources of snow CDOM and its effects on the carbon cycle is important.

Data availability. All datasets and codes used to produce this study can be obtained by contacting Xin Wang (wxin@lzu.edu.cn). The elevation data used in this study are



available at <http://rda.ucar.edu/datasets/ds759.3/#!access>.

Competing interests. The authors declare that they have no conflict of interest.

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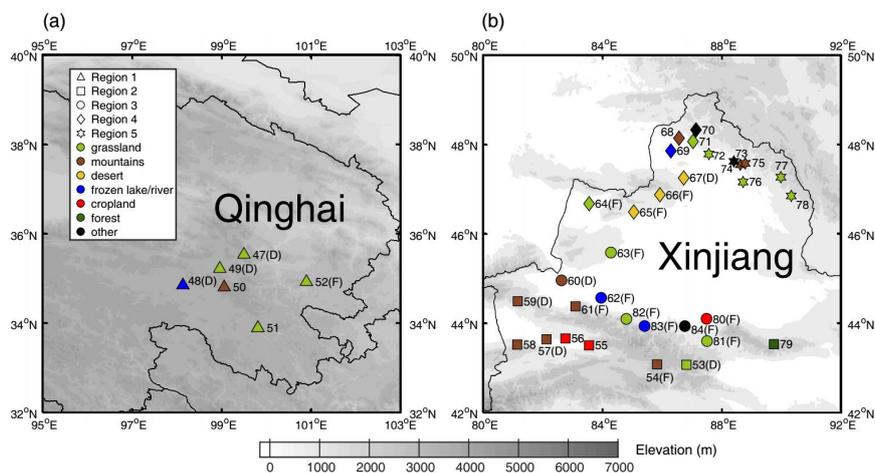


Figure 1. Sample site distribution, site numbers and regional groupings in (a) Qinghai and (b) Xinjiang. Sample areas are divided into five regions indicated by different symbol shapes, and the land cover types of sample sites are represented in different colors, as shown in the legend. The “D” indicates that the sample was collected from a snow drift, and the “F” indicates that the surface sample was fresh snow. The elevation is shown in the contour plot.

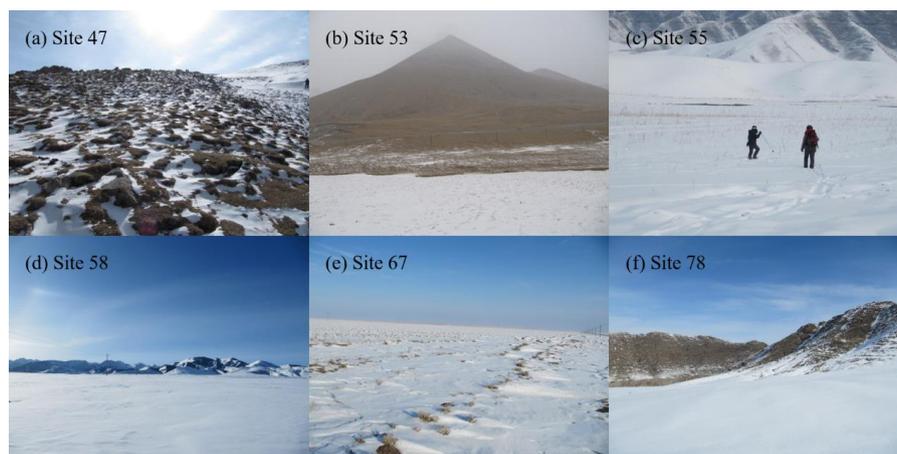


Figure 2. Pictures of typical sample sites.

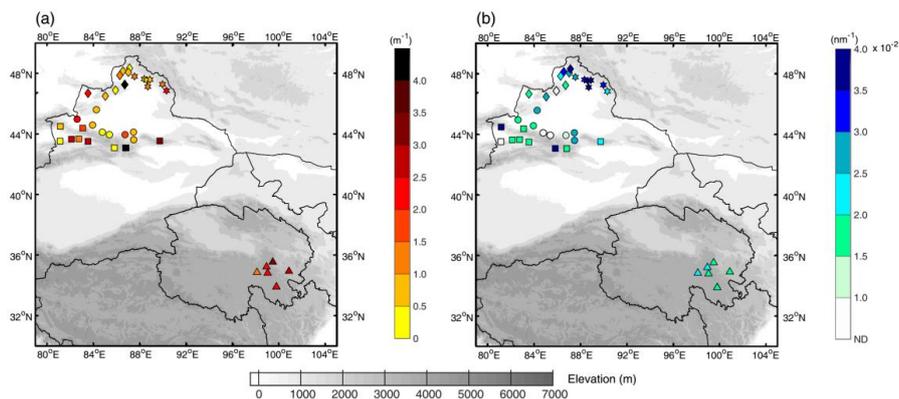


Figure 3. (a) a_{280} and (b) $S_{275-295}$ of surface snow at each site. The five regions are indicated by different symbols (same as Fig. 1).

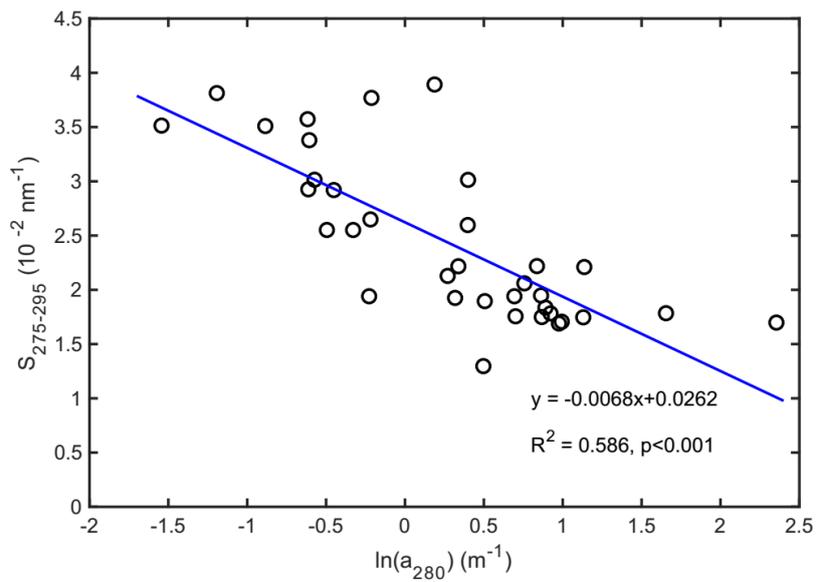


Figure 4. Relationship between log-transformed a_{280} and $S_{275-295}$

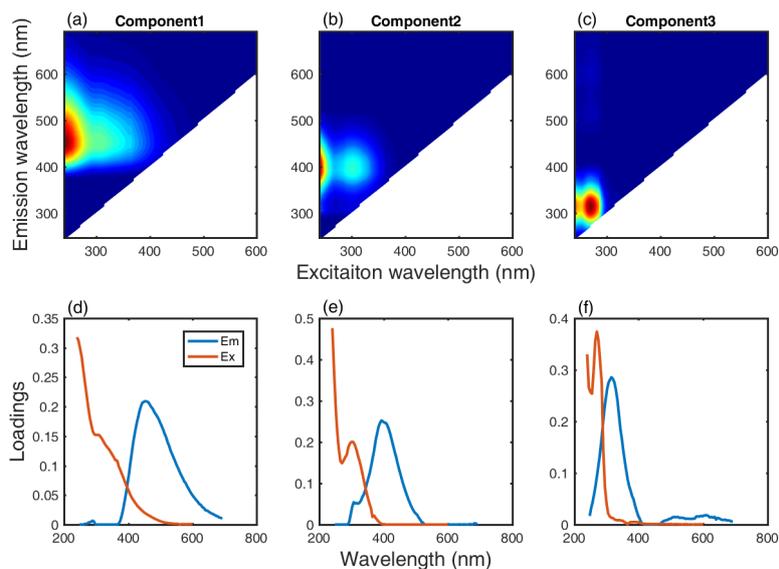


Figure 5. (a-c) The EEM components identified by the PARAFAC model and (d-f) the corresponding excitation and emission loadings of each component; the orange line indicates the excitation (Ex) loading, and the blue line indicates the emission (Em)

5 loading.

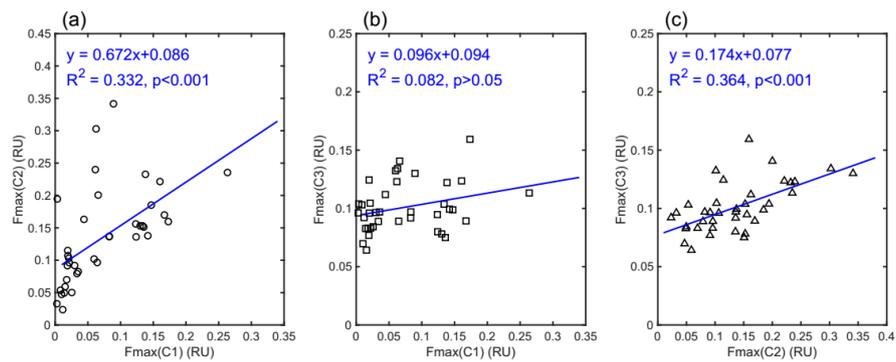


Figure 6. The linear relationships between the intensities of **(a)** C1 and C2, **(b)** C1 and C3, **(c)** C2 and C3.

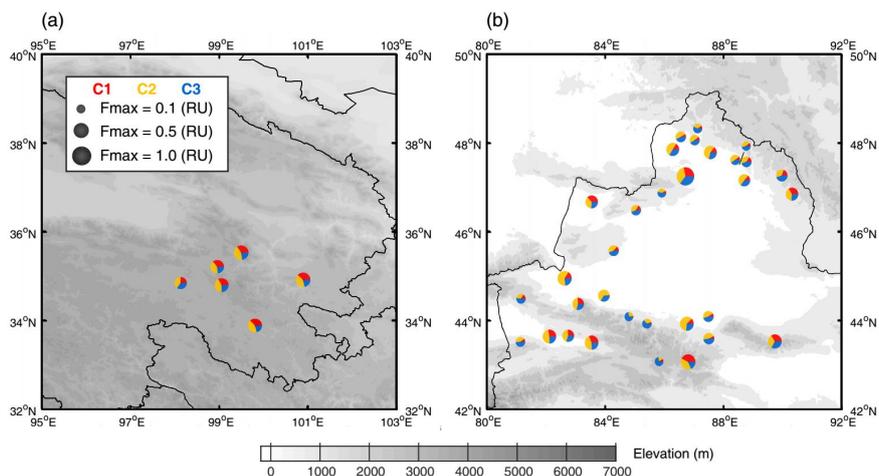


Figure 7. Distribution of relative intensities for three fluorescence components in **(a)** Qinghai and **(b)** Xinjiang. The red, yellow and blue in each pie represent C1, C2 and C3, respectively, and the size of each pie shows the total fluorescence intensity as a sum of the three components.

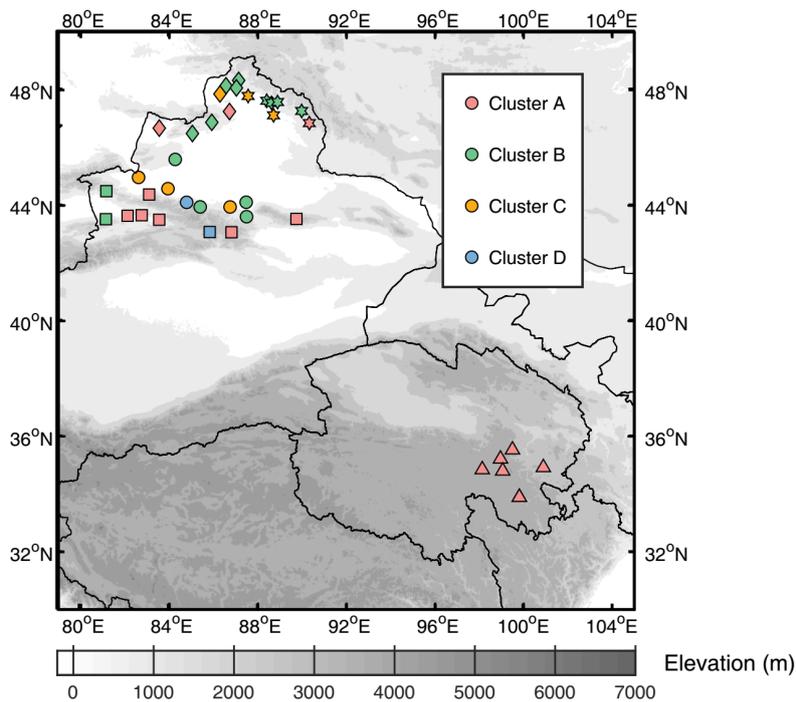


Figure 8. Hierarchical cluster analysis based on the relative intensities of the three PARAFAC components.

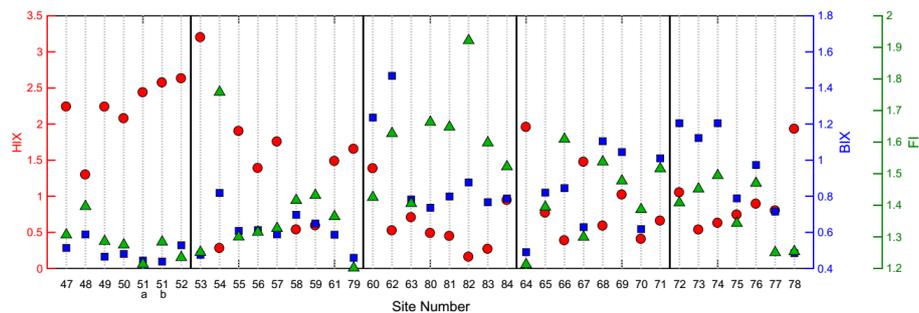


Figure 9. HIX (red y-axis and circles), BIX (blue y-axis and squares) and FI (green y-axis and triangles) of surface snow samples. The sample sites in regions 1-5 are separated by black solid lines.

5

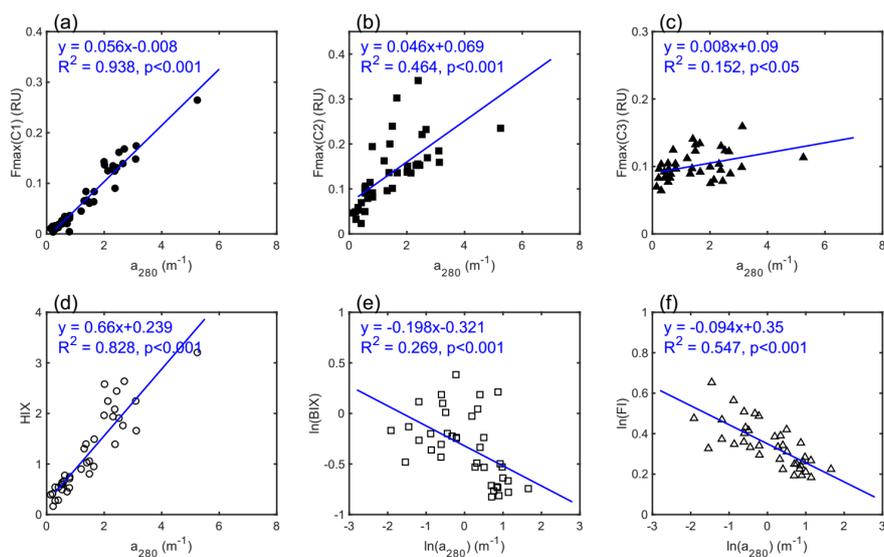


Figure 10. The relationships between a_{280} and the intensities of (a) C1, (b) C2, and (c)

C3 and the fluorescence-derived indices (d) HIX, (e) BIX, and (f) FI.

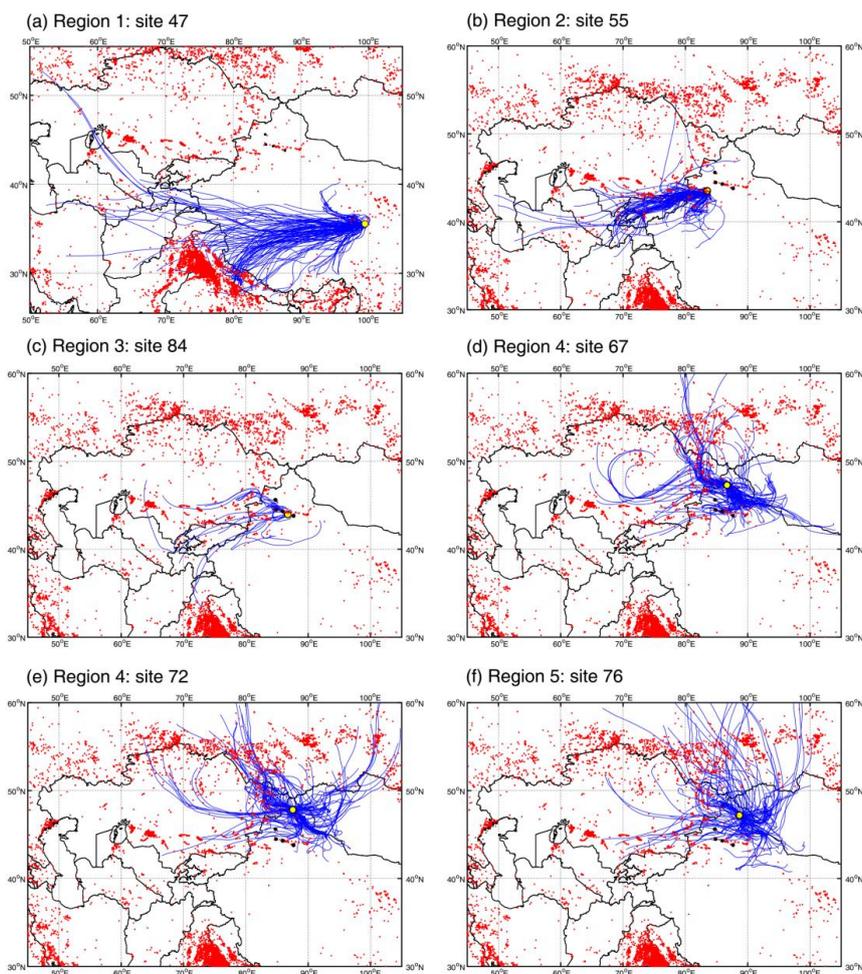


Figure 11. 72-h air mass backward trajectories (blue lines) at 500 m above ground level with the initial positions at representative sites (shown as yellow dots) in each region. Trajectories were calculated four times per day for a period of 30 days preceding the sampling date at a given site by HYSPLIT (version 4, NOAA) except for panel (c). Since the snow was fresh at site 84, the trajectories were derived for 5 days preceding the sampling date. The black dots represent the typical industrial cities in Xinjiang, namely, Karamay, Kuytun, Shihezi and Urumqi from west to east. The red dots are MODIS active fire locations.

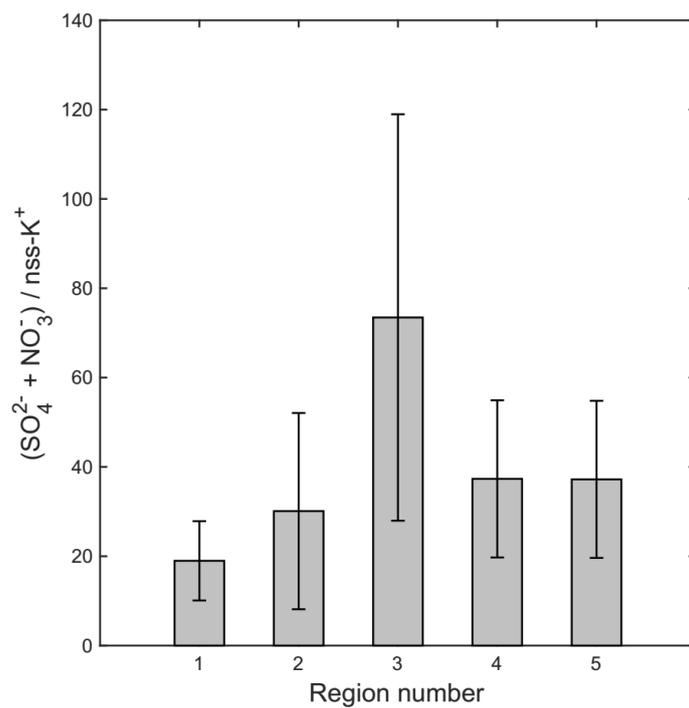


Figure 12. The regional averages of the ratio of (SO₄²⁻+NO₃⁻) and nss-K⁺.

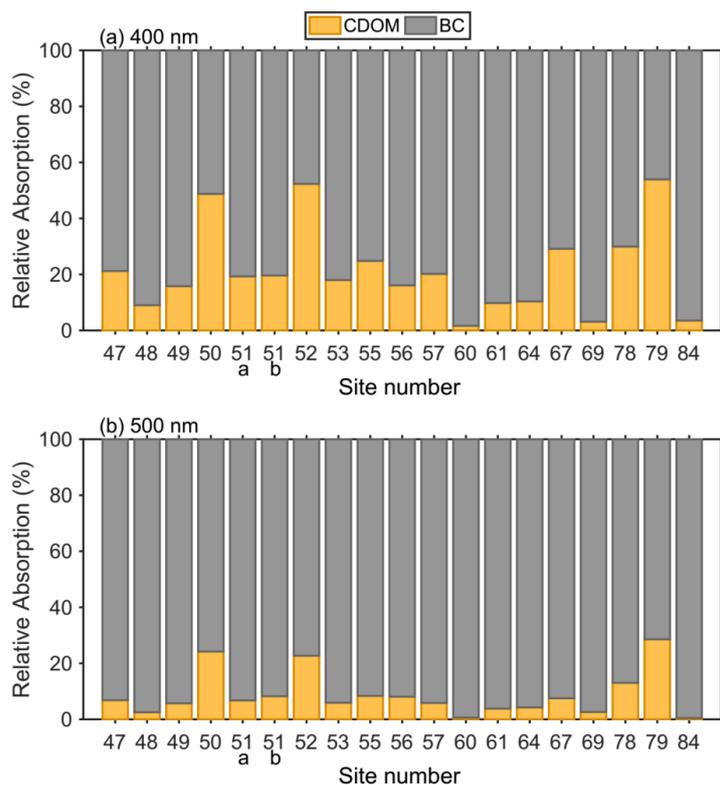


Figure 13. The relative absorption contributions of CDOM (yellow bar) and BC (gray bar) at **(a)** 400 nm and **(b)** 500 nm



Table 1. Statistics on absorption and fluorescence parameters for surface snow at each site. Note: N. A. for no data.

Site	Lat. (N)	Lon. (E)	a_{280} (m^{-1})	$S_{275-295}$ (nm^{-1})	AAE	HIX	BIX	FI
47	35.54	99.49	3.11	0.0174	4.66	2.24	0.51	1.31
48	34.85	98.13	1.32	0.0212	5.12	1.30	0.59	1.40
49	35.22	98.95	2.14	0.0206	5.20	2.24	0.47	1.29
50	34.80	99.05	2.38	0.0194	4.91	2.08	0.48	1.27
51a	33.89	99.80	2.44	0.0183	4.87	2.44	0.44	1.21
51b	33.89	99.80	2.02	0.0175	4.91	2.57	0.44	1.28
52	34.92	100.89	2.71	0.0170	4.63	2.63	0.53	1.23
53	43.07	86.81	5.25	0.0178	4.53	3.20	0.48	1.25
54	43.08	85.82	0.41	0.0350	N.A.	0.28	0.82	1.76
55	43.51	83.54	2.52	0.0178	5.13	1.90	0.61	1.30
56	43.66	82.75	1.38	0.0192	5.77	1.39	0.61	1.31
57	43.64	82.11	2.66	0.0168	5.31	1.75	0.59	1.33
58	43.52	81.13	0.42	N.A.	N.A.	0.54	0.70	1.42
59	44.49	81.15	0.54	0.0357	N.A.	0.59	0.65	1.43
60	44.96	82.63	2.39	0.0174	8.91	1.38	1.24	1.43
61	44.38	83.09	1.66	0.0189	6.06	1.49	0.59	1.36
62	44.57	83.96	0.80	0.0194	N.A.	0.52	1.47	1.63
63	45.58	84.29	0.81	0.0264	N.A.	0.71	0.78	1.41
64	46.68	83.54	2.01	0.0194	5.54	1.96	0.49	1.21
65	46.49	85.04	0.64	0.0291	N.A.	0.77	0.82	1.39
66	46.88	85.92	0.15	N.A.	N.A.	0.39	0.84	1.61
67	47.26	86.71	10.57	0.0169	4.41	1.47	0.63	1.30
68	48.15	86.56	0.57	0.0301	N.A.	0.59	1.10	1.54
69	47.86	86.29	1.41	0.0221	7.70	1.02	1.04	1.48
70	48.33	87.13	0.21	0.0351	N.A.	0.41	0.62	1.39
71	48.07	87.03	0.61	0.0255	N.A.	0.66	1.01	1.52
72	47.79	87.56	1.49	0.0259	N.A.	1.05	1.20	1.41
73	47.55	88.61	0.30	0.0381	N.A.	0.54	1.12	1.45
74	47.63	88.40	0.55	0.0337	N.A.	0.63	1.20	1.49
75	47.58	88.78	0.81	0.0376	N.A.	0.74	0.79	1.34
76	47.17	88.70	1.21	0.0389	N.A.	0.89	0.97	1.47
77	47.27	89.97	1.50	0.0301	N.A.	0.80	0.71	1.25
78	46.85	90.32	2.32	0.0221	5.52	1.93	0.48	1.25
79	43.53	89.74	3.13	0.0221	5.52	1.65	0.46	1.20
80	44.10	87.49	0.54	0.0292	N.A.	0.49	0.74	1.66
81	43.60	87.51	0.72	0.0255	N.A.	0.45	0.80	1.65
82	44.09	84.80	0.23	N.A.	N.A.	0.16	0.88	1.92
83	43.93	85.41	0.31	N.A.	N.A.	0.27	0.77	1.60
84	43.93	86.76	1.65	0.0129	6.66	0.95	0.79	1.52



Table 2. Description of the three PARAFAC components. The secondary peaks are shown in brackets.

Component number	Excitation maximal wavelength (nm)	Emission maximal wavelength (nm)	Descriptions	References
C1	<240 (305)	453	Terrestrial humic-like substances	Stedmon and Markager, 2005a; Stedmon et al., 2003
C2	<240 (300)	393	Microbial, anthropogenic or terrestrial humic-like substances	Murphy et al., 2011; Zhang et al., 2010
C3	<240 (270)	315	Tyrosine-like fluorophore	Yu et al., 2015



Table 3. Regional average of relative intensity (in percent) and F_{\max} of each fluorescence component.

Regions	C1 (%)	C2 (%)	C3 (%)	$F_{\max}(C1)$ (RU)	$F_{\max}(C2)$ (RU)	$F_{\max}(C3)$ (RU)
1	35±4	41±1	25±5	0.13±0.03	0.15±0.03	0.09±0.01
2	25±11	38±8	37±16	0.11±0.08	0.14±0.08	0.11±0.02
3	9±5	47±13	45±16	0.03±0.03	0.15±0.11	0.11±0.01
4	17±10	41±6	42±10	0.08±0.09	0.14±0.11	0.13±0.10
5	17±8	44±7	38±8	0.05±0.04	0.13±0.06	0.10±0.02
Total	20±11	42±9	38±14	0.08±0.07	0.14±0.08	0.11±0.05



Table 4. The regional average of fluorescence-derived indices.

Regions	HIX	BIX	FI
1	2.21±0.42	0.49±0.05	1.29±0.05
2	1.42±0.84	0.61±0.10	1.37±0.15
3	0.62±0.37	0.93±0.25	1.60±0.15
4	0.91±0.52	0.82±0.21	1.43±0.12
5	0.94±0.43	0.92±0.25	1.38±0.09
Total	1.21±0.78	0.76±0.26	1.42±0.16



Table 5. Summary of fluorescence-derived indices from natural water and water extraction of aerosol reported by other studies with average values for some studies are shown in brackets.

Study area	Sample type	HIX	BIX	FI	References
Tibetan Plateau	Cryoconite in glaciers	1.11-1.37 (1.27)	0.65-0.93 (0.80)	3.12-3.44 (3.24)	Feng et al., 2016
Yungui Plateau, China	Inland lakes	0.23-6.00 (1.57)	0.60-1.54 (0.93)	1.14-1.80 (1.37)	Zhang et al., 2010
Frasassi Caves, Italy	Cave water	1.79-3.28 (2.32)	0.80-1.12 (0.95)	~1.8	Birdwell and Engel, 2010
Springs in USA	Spring water	0.36-1.21 (0.76)	0.64-1.13 (0.87)	1.92-2.28 (2.09)	Birdwell and Engel, 2010
Gironde Estuary, France	Estuary	~10-17	0.6-0.8	1.14-1.22	Huguet et al., 2009
North Pacific Ocean	Ocean water	0.92-1.80 (1.49)	0.88-1.38 (1.0)	1.54-1.77 (1.66)	Helms et al., 2013
Tai Mountain, China	Fog water	3.23-6.79 (4.8)	0.64-1.02 (0.87)	1.42-1.83 (1.63)	Birdwell and Valsaraj, 2010
Jiangnan Plain, China	Ground water	2.71-7.49	0.88-0.97	-	Huang et al., 2015
Colorado, USA	Aerosol in alpine sites	0.72-4.75 (2.42)	0.54-0.75 (0.65)	1.18-1.57 (1.4)	Xie et al., 2016
Granada, Spain	Urban aerosol	2.79-4.89	-	1.48-1.61	Mladenov et al., 2011



Table 6. Pearson’s correlation coefficients (r) of major ions and F_{\max} for fluorescence components when excluding data from site 67; the results for the entire dataset are shown in parentheses. Note: * denotes $p < 0.001$.

	SO_4^{2-}	NO_3^-	nss- K^+
$F_{\max}(\text{C1})$	0.01 (0.14)	-0.10 (-0.04)	0.23 (0.48)
$F_{\max}(\text{C2})$	0.70* (0.72)	0.60* (0.57)	0.63* (0.73)
$F_{\max}(\text{C3})$	0.44 (0.42)	0.34 (0.23)	0.29 (0.68)



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