



Large carbon cycle sensitivities to climate across a permafrost thaw gradient in subarctic Sweden

Kuang-Yu Chang¹, William J. Riley¹, Patrick M. Crill², Robert F. Grant³, Virginia I. Rich⁴, and Scott R. Saleska⁵

¹Climate and Ecosystem Sciences Division, Lawrence Berkeley National Laboratory, Berkeley, California, USA

²Department of Geological Sciences, Stockholm University, Stockholm, Sweden

³Department of Renewable Resources, University of Alberta, Edmonton, Alberta, Canada

⁴Department of Microbiology, The Ohio State University, Columbus, Ohio, USA

⁵Department of Ecology and Evolutionary Biology, University of Arizona, Tucson, Arizona, USA

Correspondence: Kuang-Yu Chang (ckychang@lbl.gov)

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Abstract. Permafrost peatlands store large amounts of carbon potentially vulnerable to decomposition. However, the fate of that carbon in a changing climate remains uncertain in models due to complex interactions among hydrological, biogeochemical, microbial, and plant processes. In this study, we estimated effects of climate forcing biases present in global climate reanalysis products on carbon cycle predictions at a thawing permafrost peatland in subarctic Sweden. The analysis was conducted with a comprehensive biogeochemical model (*ecosys*) across a permafrost thaw gradient encompassing intact permafrost palsa with an ice core and a shallow active layer, partly thawed bog with a deeper active layer and a variable water table, and fen with a water table close to the surface, each with distinct vegetation and microbiota. Using in situ observations to correct local cold and wet biases found in the Global Soil Wetness Project Phase 3 (GSWP3) climate reanalysis forcing, we demonstrate good model performance by comparing predicted and observed carbon dioxide (CO₂) and methane (CH₄) exchanges, thaw depth, and water table depth. The simulations driven by the bias-corrected climate suggest that the three peatland types currently accumulate carbon from the atmosphere, although the bog and fen sites can have annual positive radiative forcing impacts due to their higher CH₄ emissions. Our simulations indicate that projected precipitation increases could accelerate CH₄ emissions from the palsa area, even without further degradation of palsa permafrost. The GSWP3 cold and wet biases for this site significantly alter simulation results and lead to erroneous active layer depth (ALD)

and carbon budget estimates. Biases in simulated CO₂ and CH₄ exchanges from biased climate forcing are as large as those among the thaw stages themselves at a landscape scale across the examined permafrost thaw gradient. Future studies should thus not only focus on changes in carbon budget associated with morphological changes in thawing permafrost, but also recognize the effects of climate forcing uncertainty on carbon cycling.

1 Introduction

Confidence in future climate projections depends on the accuracy of terrestrial carbon budget estimates, which are presently very uncertain (Friedlingstein et al., 2014; Arneeth et al., 2017). In addition to the complexity in physical process representations, a major source of this uncertainty comes from challenges in quantifying climate responses induced by biogeochemical feedbacks. Increases in atmospheric carbon dioxide (CO₂) concentrations can directly stimulate carbon sequestration from plant photosynthesis (Cox et al., 2000; Friedlingstein et al., 2006) and indirectly stimulate carbon emissions (e.g., from soil warming and resulting increased respiration), although the predicted magnitudes of these exchanges strongly depend on model process representations (Zaehle et al., 2010; Grant, 2013, 2014; Ghimire et al., 2016; Chang et al., 2018).

The undecomposed carbon stored in permafrost is of critical importance for biogeochemical feedbacks to climate be-

cause it is about twice as much as currently is in the atmosphere (Hugelius et al., 2014) and is vulnerable to release to the atmosphere as permafrost thaws (Schuur et al., 2015). O'Donnell et al. (2012) suggested that permafrost thaw would result in a net loss of soil organic carbon from the entire peat column because accumulation rates at the surface were insufficient to balance deep soil organic carbon losses upon thaw. Jones et al. (2017) indicated that the loss of sporadic and discontinuous permafrost by 2100 could result in a release of up to 24 Pg of soil carbon from permafrost peatlands to the atmosphere. Lundin et al. (2016) reported that it is plausible (71 % probability) for the subarctic landscapes to serve as a net carbon source to the atmosphere while its peatland components are atmospheric carbon sinks, which highlights the importance of spatial heterogeneity in high-latitude carbon budget estimation.

In addition to the overall carbon balance of the changing Arctic, the type of carbon gaseous emission is important to climate feedbacks. High latitudes are predicted to get wetter (IPCC, 2014), and saturated anaerobic conditions facilitate methane (CH_4) production, which is a much more efficient greenhouse gas than CO_2 in terms of global warming potential. Even habitats that can be net carbon sinks can produce positive radiative forcing impacts on climate due to CH_4 release, as Bäckstrand et al. (2010) showed for a subarctic peatland. Under projected warming and wetting trends in the Arctic (Collins et al., 2013; Bintanja and Andry, 2017), carbon cycle feedbacks over the permafrost region could become stronger as increased precipitation enhances surface permafrost thaw and strengthens CH_4 emissions through expansion of anaerobic volume (Christensen et al., 2004; Wickland et al., 2006).

The Stordalen Mire in northern Sweden (68.20° N, 19.03° E) is in the discontinuous permafrost zone, encompassing a mosaic of thaw stages with associated distinct hydrology and vegetation (Christensen et al., 2004; Malmer et al., 2005), microbiota (Mondav et al., 2014, 2017; Woodcroft et al., 2018), and organic matter chemistry (Hodgkins et al., 2014). These landscapes have been shifting over the last half-century to a more thawed state, likely due to recent warming (Christensen et al., 2004). Drier hummock sites dominated by shrubs have degraded to wetter sites dominated by graminoids (Malmer et al., 2005; Johansson et al., 2006). The thaw-induced habitat shifts are associated with increases in landscape-scale CH_4 emissions (Christensen et al., 2004; Johansson et al., 2006; Cooper et al., 2017), reflective of the higher CH_4 emissions of the wetter thawed habitats (McCalley et al., 2014). The higher CO_2 uptake in later thaw-stage habitats has not compensated for the increase in positive radiative forcing from elevated CH_4 emissions (Bäckstrand et al., 2010; Deng et al., 2014).

The impacts of climate sensitivity on the terrestrial carbon cycle have been investigated at the global scale, and the results highlight the need to consider uncertainty in climate datasets when evaluating permafrost region carbon cy-

cle simulations (Ahlström et al., 2017; Guo et al., 2017; Wu et al., 2017). Ahlström et al. (2017) showed that climate forcing biases are responsible for a considerable fraction ($\sim 40\%$) of the uncertainty range in ecosystem carbon predictions from 18 Earth System Models (ESMs) reported by Anav et al. (2013). Guo et al. (2017) concluded that the differences in climate forcing contribute to significant differences in simulated soil temperature, permafrost area, and active layer depth (ALD). Wu et al. (2017) demonstrated that differences among climate forcing datasets contribute more to predictive uncertainty than differences in apparent model sensitivity to climate forcing. However, notably, none of these studies accessed the effects on CH_4 emissions, and their spatial resolution could not represent site-level spatial heterogeneity observed in arctic tundra (Grant et al., 2017a, b).

Here, we use the ecosystem model *ecosys*, which employs a comprehensive set of coupled biogeochemical and hydrological processes to estimate the effects of climate forcing uncertainty and sensitivity on CO_2 and CH_4 exchanges and thaw depth simulations. For the Stordalen Mire site, we estimated bias in the Global Soil Wetness Project Phase 3 (GSWP3) climate reanalysis dataset using site-level long-term meteorological measurements and evaluated impacts on simulated soil and plant processes across the permafrost thaw gradient. This approach enables us to assess model sensitivity to individual climate forcing biases, instead of the aggregated uncertainty range embedded in climate datasets (e.g., variations of climate conditions represented in different climate datasets) presented in previous studies. We address the following questions for our study site at the Stordalen Mire. (1) What are the biases embedded in the GSWP3 climate reanalysis dataset? (2) How do these biases affect model predictions of thaw depth, CO_2 exchanges, and CH_4 exchanges? (3) How does climate sensitivity vary across the stages of permafrost thaw? In addition to improving understanding of permafrost responses to climate, we identify ecosystem carbon prediction uncertainty induced by climate forcing uncertainty in general as the biases found in GSWP3 were consistent with other climate reanalysis datasets during the last decade (Sect. 3).

2 Methods and data

2.1 Study site description

Our study sites are located at the Stordalen Mire (68.20° N, 19.03° E; 351 m above sea level), which is about 10 km southeast of the Abisko Scientific Research Station (ANS) in northern Sweden. The Stordalen Mire is in the discontinuous permafrost zone along the 0°C isotherm, where permafrost at low elevations is primarily present in peatlands, bordered by lakes to the northwest and southeast (Kokfelt et al., 2010). A large portion of the mire consists of a slightly

elevated drained area underlain by permafrost characterized by a hummocky topography, and the remaining portion is largely lacking permafrost with fen-like conditions (Johansson et al., 2006). The recent warming (more than 1 °C) has deepened the mean ALD measured at the Stordalen Mire by around 20 cm since the early 1980s, accompanied by palsa collapses and thermokarst erosion (Christensen et al., 2004; Malmer et al., 2005; Johansson et al., 2006). Specifically, the mean ALD increased from 0.48 to 0.63 m in the drier part of the mire and from 0.63 to 0.86 m in the wetter part, from September 1973–1976 (Rydén and Kostov, 1980) to September 2003–2005 (Johansson et al., 2006).

Significant changes in climate over this region have been recorded during the last few decades. The annual mean air temperature measured at the ANS rose by 2.5 °C from 1913 to 2006, when it exceeded the 0 °C threshold (0.6 °C in 2006) for the first time over the past century (Callaghan et al., 2010). The measured annual total precipitation also increased from 306 mm yr⁻¹ (years 1913 to 2009) to 336 mm yr⁻¹ (years 1980 to 2009) (Olefeldt and Roulet, 2012), along with increased variability in extreme precipitation (Callaghan et al., 2010). The measured annual maximum snow depth increased from 59 cm (years 1957 to 1971) to 70 cm (years 1986 to 2000); however, the snow cover period with snow depth greater than 20 cm decreased from 5.8 months (years 1957 to 1971) to 4.9 months (years 1986 to 2000) (Malmer et al., 2005).

Inception of peat deposition at the Stordalen Mire has been dated at around 6000 calendar years before present (cal BP) (Sonesson, 1972) in the southern part of the mire and at around 4700 cal BP in the northern part (Kokfelt et al., 2010). Kokfelt et al. (2010) suggested that permafrost aggregation initiated during the Little Ice Age (around 120–400 cal BP) in the Stordalen Mire. At present, the Stordalen Mire can be broadly classified into three peatland types: intact permafrost palsa, partly thawed bog, and fen (Hodgkins et al., 2014), hereafter referred to as palsa, bog, and fen (Fig. 1). The spatial distribution of these peatland types in 2000 is described in Olefeldt and Roulet (2012).

Based on Swedish military photography, all three of the investigated peatland types have existed since at least the 1930s. The palsa sites are ombrotrophic and raised 0.5 to 2.0 m above their surroundings, with a relatively thin peat layer (0.4 to 0.7 m; Rydén et al., 1980), thinner active layer depth (less than 0.7 m in late summer), and no measurable water table depth (Bäckstrand et al., 2008a, b; Olefeldt and Roulet, 2012). The bog sites are ombrotrophic and are wetter than the palsa sites, with a thicker peat layer (0.5 to ~ 1 m; Rydén et al., 1980), deeper ALD (greater than 0.9 m, the deepest measurement depth), and water table depth fluctuating from 35 cm below the peat surface to the ground surface (Bäckstrand et al., 2008a, b; Olefeldt and Roulet, 2012). The fen sites are minerotrophic, receiving a large amount of water from a lake to the east of the mire, with water table depths

near or above the ground surface (Bäckstrand et al., 2008a, b; Olefeldt and Roulet, 2012).

Differences in hydrology and permafrost conditions create high spatial heterogeneity with different soil moisture, pH, and nutrient conditions that support different plant communities (Bäckstrand et al., 2008a, b). The palsa is dominated by dwarf shrubs with some sedges, feather mosses, and lichens (Malmer et al., 2005; Bäckstrand et al., 2008a, b; Olefeldt and Roulet, 2012). The bog is dominated by *Sphagnum* spp. mosses with a moderate abundance of sedges (Malmer et al., 2005; Bäckstrand et al., 2008a, b; Olefeldt and Roulet, 2012). The fen sites we studied are dominated by sedges (Bäckstrand et al., 2008a, b).

2.2 Field measurements

Continuous daily meteorological measurements have been recorded at the ANS since 1913, including air temperature, precipitation, wind speed, wind direction, relative humidity, and snow depth. Measurements of solar radiation, longwave radiation, and soil temperature have also been available at the ANS since 1982. The soil thaw depth (measured to 90 cm) and water table depth measurements were taken in the three peatland types three to five times per week from early May to mid-October during 2003 to 2007 (Bäckstrand et al., 2008b).

CO₂ and CH₄ exchanges in the three peatland types were measured with automated chambers during the thawed seasons from 2002 to 2007 (Bäckstrand et al., 2008b). Chamber lids were removed when snow accumulates in winter (around November), and the sampling periods for each year ranged from 60 days (28 March (day 87) to 27 May (day 147)) in 2002 (shortest) to 193 days (28 May (day 148) to 7 December (day 341)) (longest) (Bäckstrand et al., 2008b, 2010). Three chambers were installed in the palsa, another three in the bog, and two more in the fen (we term each chamber a “subsite” in the following). Each chamber covered an area of 0.14 m² with a height of 25–45 cm depending on the vegetation and the depth of insertion and was closed for 5 min every 3 h to measure CO₂ and total hydrocarbon (THC) exchanges. CH₄ exchanges were manually observed approximately three times per week, and these measurements were used to quantify the proportion of CH₄ in the measured THC (Bäckstrand et al., 2008a). The CH₄ exchanges were near zero in the palsa sites (Bäckstrand et al., 2008a, b, 2010), so they were not used in model evaluation. We used the CO₂ and CH₄ exchanges observed at 3-hourly steps when the *R*² values recorded in the measurements were greater than 0.8 (Tokida et al., 2007) and then calculated the associated daily mean exchanges when there were eight measurements per day (Table 1). The quality-controlled daily measurements only covered 12.4 %–33.7 % of the daily data points because of the lack of continuous quality-controlled 3-hourly measurements. The data screening was applied to exclude unreliable measurements and avoid biases from inappropriate gap filling, which is necessary for model evaluations. More de-

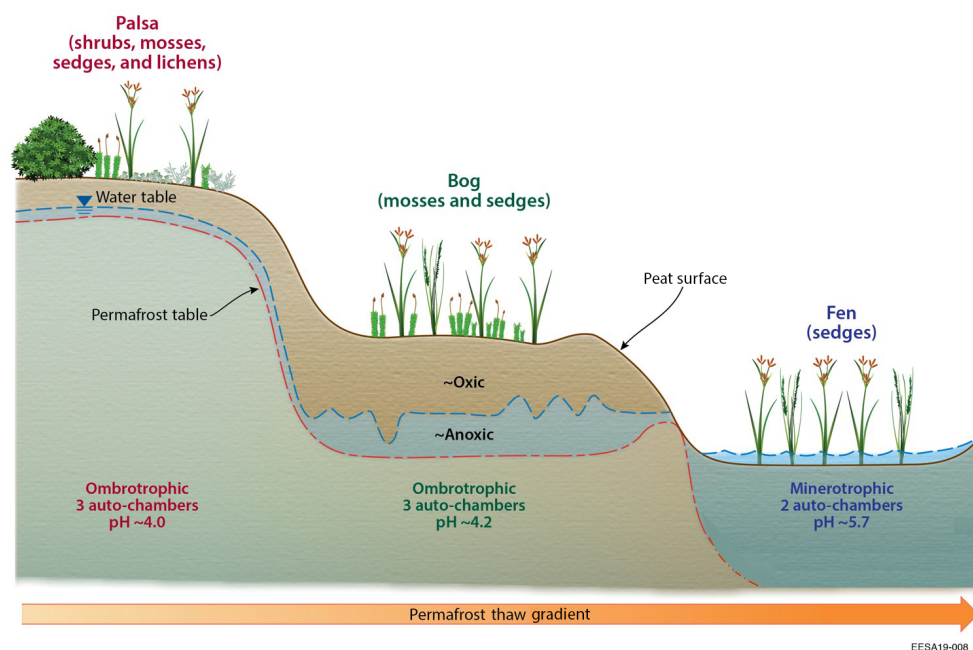


Figure 1. Schematic diagram of the sampling sites at Stordalen Mire, adapted from Johansson et al. (2006).

tailed descriptions of the CO_2 and CH_4 exchange measurements can be found in Bäckstrand et al. (2008a).

2.3 GSWP3

GSWP3 is an ongoing modeling activity that provides global gridded meteorological forcing ($0.5^\circ \times 0.5^\circ$ resolution) and investigates changes in energy, water, and carbon cycles throughout the 20th and 21st centuries. The GSWP3 dataset is based on the 20th Century Reanalysis (Compo et al., 2011), using a spectral nudging dynamical downscaling technique described in Yoshimura and Kanamitsu (2008). A more detailed description of the GSWP can be found in Dirmeyer (2011) and van den Hurk et al. (2016).

In this study, we extracted the meteorological conditions at the Stordalen Mire from 1901 to 2010 from the GSWP3 climate reanalysis dataset. The 3-hourly products of air temperature, precipitation, solar radiation, wind speed, and specific humidity were interpolated to hourly intervals with cubic spline interpolation to serve as the meteorological inputs used in our model.

The GSWP3 dataset was chosen over other existing climate reanalysis datasets for its spatial and temporal resolutions. For example, the Climatic Research Unit (CRU; Harris et al., 2014) dataset provided monthly meteorological forcing at $0.5^\circ \times 0.5^\circ$ resolution; the National Centers for Environmental Prediction (NCEP; Kalnay et al., 1996; Kanamitsu et al., 2002) dataset provided 6-hourly meteorological forcing for a T62 Gaussian grid ($\sim 1.915^\circ \times 1.895^\circ$ resolution); the CRUNCEP (Viovy, 2018) dataset provided 6-hourly meteorological forcing at $0.5^\circ \times 0.5^\circ$ resolution; and

the European Centre for Medium-Range Weather Forecasts (ECMWF; Berrisford et al., 2011) dataset provided 3-hourly meteorological forcing with 125 km ($\sim 1.125^\circ$) horizontal resolution.

2.4 Model description

The *ecosys* model is a comprehensive biogeochemistry model that simulates ecosystem responses to diverse environmental conditions with explicit representations of microbial dynamics and soil carbon, nitrogen, and phosphorus biogeochemistry. The above-ground processes are represented in multilayer plant interacting canopies that are allowed to change with changing environmental conditions, and the below-ground processes are represented in multiple soil layers with multiphase subsurface reactive transport. The *ecosys* model operates at variable time steps (down to seconds) determined by convergence criteria, and it can be applied at patch scale (spatially homogenous one-dimensional) and landscape scale (spatially variable two- or three-dimensional). Detailed descriptions, including inputs, outputs, governing equations, parameters, and references of the *ecosys* model can be found in Grant (2013). A qualitative summary of the *ecosys* model structure is provided in the Supplement.

The *ecosys* model has been extensively tested against eddy covariance fluxes and related ecophysiological measurements with a wide range of sites and weather conditions in boreal, temperate, and tropical forests (Grant et al., 2007a, c, 2009a, b, c, 2010), wetlands (Dimitrov et al., 2011, 2014; Grant et al., 2012b; Mezbahuddin et al., 2014), grasslands

Table 1. Temporal coverage of quality-controlled CO₂ and CH₄ exchanges measured by automated chambers in the three peatland types in the Stordalen Mire during the years 2002 to 2007.

Sites	CO ₂			CH ₄		
	Number of data points	3-hourly coverage (%)	Daily coverage (%)	Number of data points	3-hourly coverage (%)	Daily coverage (%)
Palsa	12 752	65.8	12.4	n/a	n/a	n/a
Bog	12 821	68.5	12.7	6660	96.2	25.0
Fen	8989	63.8	13.7	4923	90.5	33.7

Note that the temporal coverage represents the percentage of data points that passed our quality-controlled threshold at the corresponding time steps. See main text for threshold details. n/a: not applicable.

(Grant and Flanagan, 2007; Grant et al., 2012a), tundra (Grant et al., 2003, 2011b, 2015; Grant, 2015), croplands (Grant et al., 2007b, 2011a), and other permafrost-associated habitats (Grant and Roulet, 2002; Grant et al., 2017a, b). All *ecosys* model structures are unchanged from those described in these earlier studies.

2.5 Experimental design

To evaluate the effects of climate on model predictions, we conducted four sets of simulations in each of the three peatland types at the Stordalen Mire from 1901 to 2010. The climate data from 1901 to 2001 were used for model initialization (i.e., spin-up), and those from 2002 to 2010 were used for analysis. The 110-year simulations were performed to ensure the simulation was equilibrated with local climate (Grant et al., 2017a).

The meteorological conditions for all the simulations were based on the hourly data extracted from the GSWP3 climate reanalysis dataset (Sect. 2.3). The monthly mean bias of the GSWP3 for this location was calculated by comparing it to the air temperature and precipitation measured at the ANS, for the years 1913 to 2010 (Sect. 3.1). The full series of air temperature and precipitation extracted from GSWP3 were then bias-corrected using the monthly mean bias calculated from 1913 to 2010; we label this model scenario CTRL. Our bias correction was conceptually similar to the one used in Ahlström et al. (2017), in which the bias-corrected climate forcing fields were the ESM outputs adjusted by the corresponding bias calculated from observations in a reference period.

The simulation results from CTRL should represent the reliability of applying *ecosys* at the Stordalen Mire because CTRL is driven by the best local climate description. We first evaluated predicted thaw depth, water table depth, and CO₂ and CH₄ exchanges using the CTRL simulation (Sect. 3.2 to 3.4). In the second set of simulations, BIASSED-COLD, the biased GSWP3 air temperature data were used, and we only corrected the GSWP3 precipitation. Deviations between CTRL and BIASSED-COLD reflect the biased air temperature's effects on responses across the thaw gradient. In the

third set of simulations, BIASSED-WET, we bias-corrected the air temperature extracted from GSWP3, which allows us to quantify the effects of biased precipitation. Finally, we used the meteorological conditions directly extracted from GSWP3 to drive our fourth set of simulations, BIASSED-COLD&BIASED-WET, which reveals the uncertainty range of subarctic peatland simulation associated with the local biases in GSWP3 climate forcing.

While the three peatland types share the same climate conditions, they differ in soil hydrologic conditions and vegetation characteristics (Sect. 2.1; Fig. 1). The bulk density and porosity profiles were set to the values reported in Rydén et al. (1980), who suggested a decreasing trend of bulk density and an increasing trend of porosity from palsa (0.12 Mg m⁻³ at surface; 92 %–93 % within the upper 10 cm) to bog and fen (0.06 Mg m⁻³ at surface; 96 %–97 % within the upper 10 cm). The peatland soil carbon-to-nitrogen (CN) ratios and pH values were assigned according to Hodgkins et al. (2014), who documented an increasing trend of pH from palsa (4.0), to bog (4.2), to fen (5.7), and a decreasing trend of soil organic matter CN ratio from bog (46 ± 18), to palsa (39 ± 24), to fen (19 ± 0.4). Common values of field capacity (0.4) and wilting point (0.15) were used for the three peatland types (Deng et al., 2014). The soil property and vegetation parameters used in our simulation for the three peatland types are summarized in Tables S1 and S2 in the Supplement, respectively.

3 Results and discussion

3.1 GSWP3 climate comparison to observations

As described in Sect. 2.3, we extracted meteorological conditions at the Stordalen Mire from the GSWP3 climate reanalysis dataset. The closest GSWP3 grid cell was centered at 68.0° N and 19.0° E, which covers the Stordalen Mire and the ANS. The annual mean air temperature and precipitation calculated at this GSWP3 grid cell were −3.65 °C and 683.88 mm yr⁻¹, respectively, for the years 1913 to 2010. A cold bias (−3.09 °C) was identified in the GSWP3 an-

nual mean air temperature during the 1913 to 2010 period, although a very high correlation coefficient ($r = 0.99$) was found when compared with the ANS measurements (Fig. 2a). Both time series exhibit an overall warming trend from the early 20th century to the present ($0.01\text{ }^{\circ}\text{C yr}^{-1}$), with an even larger warming trend from 1980 to 2010 ($0.05\text{ }^{\circ}\text{C yr}^{-1}$ (ANS) and $0.04\text{ }^{\circ}\text{C yr}^{-1}$ (GSPW3)).

Similarly, the GSPW3 annual total precipitation data correlate well with ANS measurements ($r = 0.80$) but have a wet bias of 380 mm yr^{-1} between 1913 and 2010 (Fig. 2b). An increasing trend in annual total precipitation was recorded in both time series from the early 20th century to present (0.47 mm yr^{-2} (ANS) and 1.07 mm yr^{-2} (GSPW3)), although a decreasing trend was found from 1980 to 2010 (-0.56 mm yr^{-2} (ANS) and -2.39 mm yr^{-2} (GSPW3)).

The seasonal cycle of the GSPW3 monthly mean air temperature also matches that measured at the ANS, with a very high correlation coefficient ($r = 0.99$; Fig. 3a). The underestimation bias and interannual variability of GSPW3 air temperature are greater in winter (maximum underestimate in December, at $-4.52\text{ }^{\circ}\text{C}$ with interannual variability of $3.53\text{ }^{\circ}\text{C}$) and smaller in summer (minimum underestimate in July, at $-1.52\text{ }^{\circ}\text{C}$ with interannual variability of $1.65\text{ }^{\circ}\text{C}$), respectively.

The magnitude and interannual variability of the GSPW3 monthly mean precipitation are comparable between winter and summer, while the ANS measurements exhibit stronger seasonality with lower magnitudes during winter. Despite the differences found in seasonal patterns, a high correlation coefficient ($r = 0.64$) was found between the monthly mean precipitation extracted from GSPW3 and the ANS measurements. The overestimation of monthly mean precipitation was greatest in December ($43.25\text{ mm month}^{-1}$) and smallest in August ($18.75\text{ mm month}^{-1}$).

These comparisons suggest that GSPW3 air temperature and precipitation data reasonably capture measured seasonal and long-term trends over the past decades, but are biased cold and wet compared to observations, especially during winter. Similar cold and wet biases exist in CRUNCEP and ECMWF climate reanalysis datasets during our 2003 to 2007 study period (Fig. S1 in the Supplement). The annual mean air temperature and precipitation at the Stordalen Mire for the years 2003 to 2007 were $-2.49\text{ }^{\circ}\text{C}$ and 795.09 mm yr^{-1} ; $-2.46\text{ }^{\circ}\text{C}$ and 708.60 mm yr^{-1} ; and $-2.28\text{ }^{\circ}\text{C}$ and 765.67 mm yr^{-1} in the GSPW3, CRUNCEP, and ECMWF climate reanalysis datasets, respectively.

3.2 Model testing

3.2.1 Thaw depth

We first evaluated *ecosys* against observations using bias-corrected climate forcing (i.e., the CTRL simulation). Predicted thaw depth agrees well with measurements collected

from 2003 to 2007 for all examined peatland types (Fig. 4), with a correlation coefficient of 0.95, 0.87, and 0.41 in the palsa, bog, and fen, respectively. Both simulations and observations show that the rate of thaw depth deepening in the summer varies with peatland type (i.e., relatively slow, moderate, and rapid in the palsa, bog, and fen, respectively).

Predicted and observed maximum thaw depths (i.e., ALD) in the intact permafrost palsa were between 45 and 60 cm in September. In the partly thawed bog, the simulated thaw depth is slightly shallower than that observed before August. The simulated bog thaw depth exceeds 90 cm by the end of August, which matches the time when measured thaw depth reaches its maximum. In contrast, the thaw depth exceeds 90 cm nearly 1 month earlier in the fen. The patterns of thawing permafrost presented here are consistent with Deng et al. (2014), who simulated the same site using the DNDC model.

3.2.2 CO₂ exchanges

The daily net ecosystem exchange (NEE) simulated in the CTRL simulation reasonably captures observed seasonal dynamics from 2003 to 2007 for all the examined peatland types (Fig. 5). The simulations and observations generally showed net CO₂ uptake (with some episodic CO₂ emissions) during summer and release during winter. The observations and simulations also showed large CO₂ emissions in the palsa site during the fall of 2004. Simulated fall CO₂ bursts in the three sites in other years could not be confirmed because of a lack of observations during these periods. Similar to the patterns reported in Raz-Yaseef et al. (2016), some episodic CO₂ emission pulses were simulated as surface ice thaws in spring, but there were no measurements to confirm those events. The correlation coefficients of the simulated and observed daily NEE ranged from 0.58 to 0.60, and most of the discrepancies between the simulations and observations were within the ranges of NEE variability measured at different subsites (automated chambers) within the same peatland type. The simulated CO₂ uptake rates in the bog were greater than the observations in summer, which could be due to overestimated plant biomass or overestimated CO₂ uptake rate per plant biomass. However, we currently do not have data to examine the cause of this overestimation because the CO₂ flux derived from automated chambers only represents the aggregated results of all controlling factors.

As described in Sect. 2.2, simulated CO₂ exchanges were evaluated for 3-hourly and daily time steps when quality-controlled measurements were available (R^2 values and relative root mean squared errors (RRMSEs) shown in Table 2). Simulated NEE is in reasonable agreement with the 3-hourly NEE measurements, with RRMSEs ranging from 8.4 % to 19.1 %. Model comparisons with observations were generally poorer at daily time steps, although the calculated RRMSEs were comparable to those reported in Deng et al. (2014). We suspect these differences resulted from uncertainty in de-

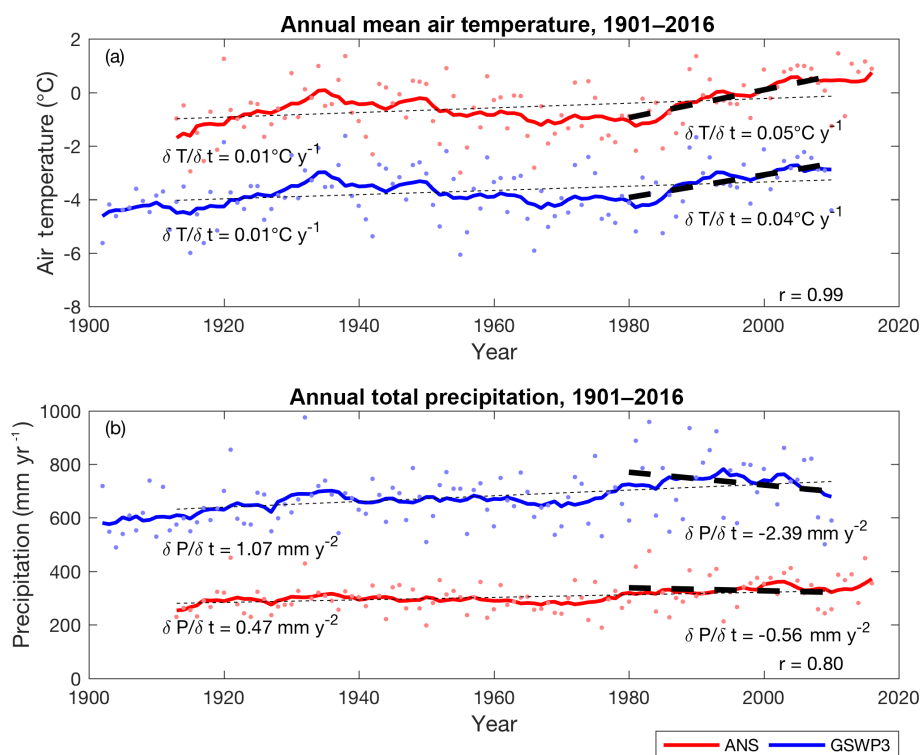


Figure 2. Time series of air temperature (a) and precipitation (b) measured at ANS (red; years 1913–2016) and extracted from GSWP3 (blue; years 1901–2010). Dots are the annual means, and solid lines are the decadal moving averages of the corresponding annual means. Thin and thick dashed lines are the trends for the years 1913–2010 and the years 1980–2010, respectively. The inset r values are the correlation coefficients calculated between the two time series.

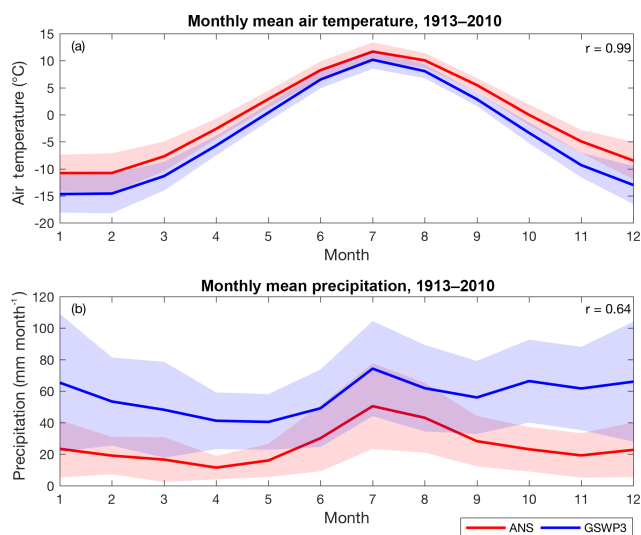


Figure 3. Monthly mean air temperature (a) and precipitation (b) measured at ANS (red) and extracted from GSWP3 (blue). The shaded area is the interannual variability for the corresponding dataset, represented by the standard deviations calculated for each month. The inset r values are the correlation coefficients calculated between the two time series.

termining an accurate observed daily NEE that is representative of the entire peatland type. This may be due to (1) limited coverage of daily data points (less than 14 % across the study period; Table 1) due to a lack of continuous quality-controlled 3-hourly measurements and (2) the large variability of daily NEE ranges measured at different subsites within the same peatland type (Fig. 5). Our results thus indicate that NEE is affected by thaw stage (Bäckstrand et al., 2010; Deng et al., 2014) and fine-scale spatial heterogeneity of the system. More detailed measurements with higher spatial and temporal resolutions within the same peatland type would be necessary to characterize the effects of this type of heterogeneity.

3.2.3 Water table depth and CH_4 exchanges

Simulated water table depth generally captures observed seasonal patterns measured in the bog and fen sites from 2003 to 2007 (Fig. 6a, c). During summer, the predicted bog water table depth fluctuates around the ground surface (−7 to −1 cm), and the predicted water table depth is at or above the ground surface in the fen. Water table depths simulated by *ecosys* are generally higher than measured in the bog, where measured water table depths are often below the ground surface with greater seasonal variability. Simulated fen water ta-

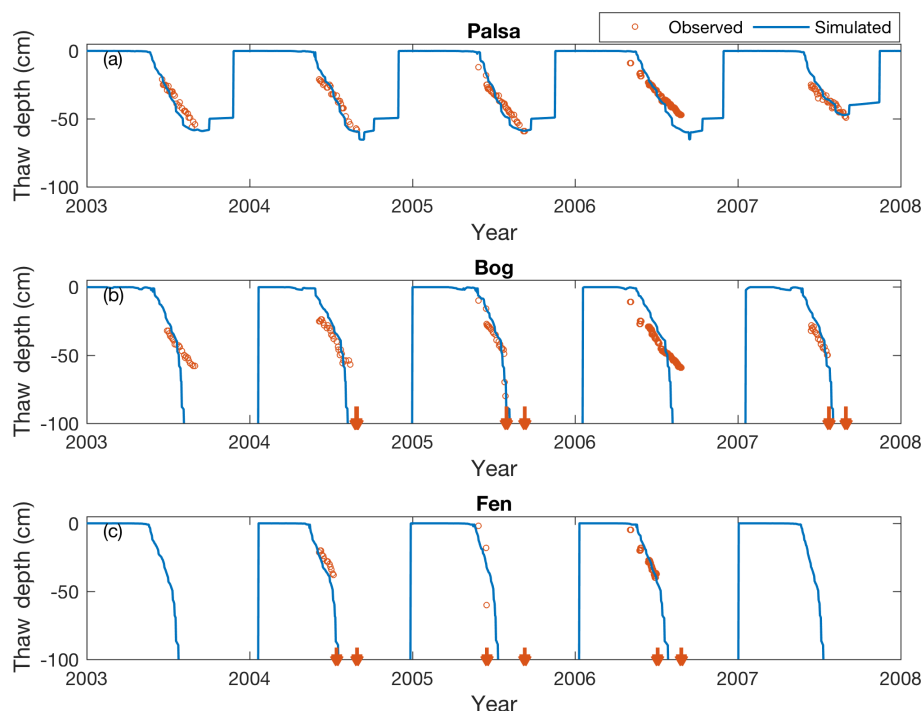


Figure 4. Simulated (solid lines) and measured (open circles) seasonal dynamics of thaw depth in the palsa (a), bog (b), and fen (c) sites from 2003 to 2007. Downward arrows indicate the time when measured thaw depth exceeds 90 cm for a measurement year.

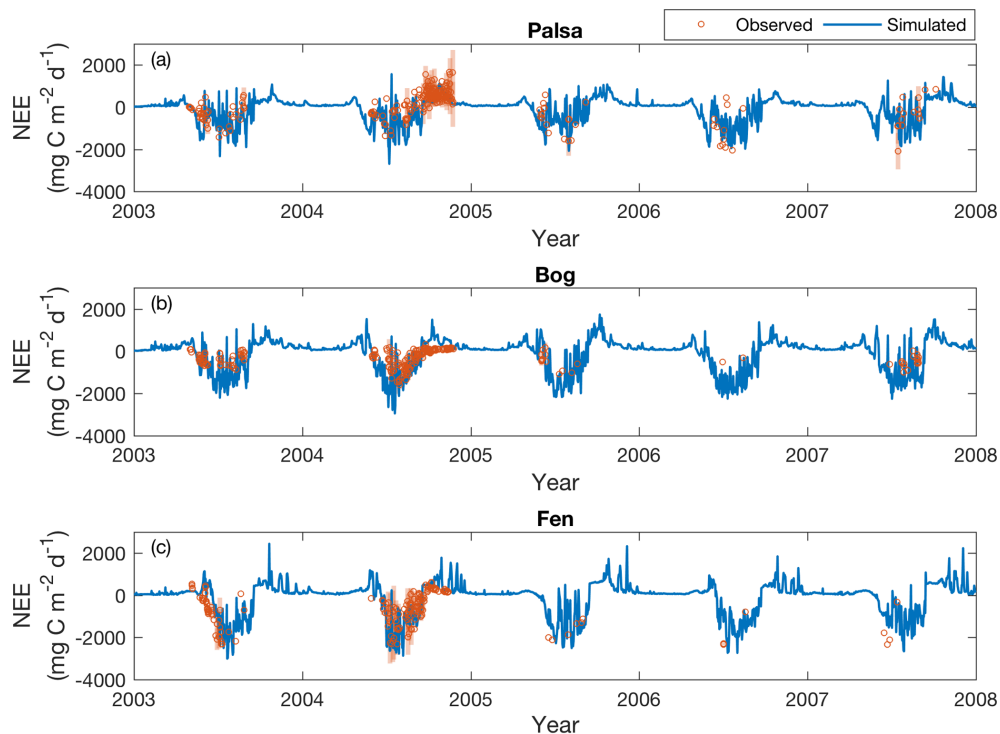


Figure 5. Simulated (solid lines) and measured (open circles) daily CO_2 exchanges (NEE) in the palsa (a), bog (b), and fen (c) sites from 2003 to 2007. Shaded bars are the standard deviations of daily NEE measured across subsites under each peatland type. Positive and negative values indicate effluxes from and influxes to the site, respectively.

Table 2. Evaluation of the 3-hourly and daily CO₂ and CH₄ exchanges simulated in the palsa, bog, and fen sites.

Sites	C component	3-hourly		Daily	
		R^2	RRMSE (%)	R^2	RRMSE (%)
Palsa	CO ₂	0.48	13.4	0.36	18.3
Bog	CO ₂	0.63	19.1	0.44	35.8
	CH ₄	0.31	16.3	0.47	22.3
Fen	CO ₂	0.64	8.4	0.43	25.5
	CH ₄	0.44	11.1	0.54	16.9

Note that RRMSE is the relative root mean squared error.

ble depths have better overall fit to observations, being higher (~ 5 cm) than measurements in 2003 and 2004, close to measurements in 2005 and 2006, and slightly deeper (~ 2 cm) than measurements in 2007. These differences in modeled and observed water table depth could be driven by the limitations of our one-dimensional column simulation that could not resolve topographic effects and thus hinder the variations of water table depth, which is a particular issue in simulating the dynamic water table of the bog. For example, no excessive water could be transported to the neighboring grids to deepen local water table depth under our current model configuration. A multidimensional simulation that includes realistic topographic effects could help improve the representation of water table dynamics, and estimates of the measurement uncertainty would help facilitate the assessment of simulation bias.

Simulated and measured daily CH₄ exchanges correlate reasonably well in the bog ($r = 0.49$) and well in the fen ($r = 0.65$) across the study period (Fig. 6b, d). Both the simulations and observations have stronger CH₄ emissions during summer, with peak emissions in late summer. Some episodic CH₄ emission pulses (Mastepanov et al., 2008) were simulated during shoulder seasons, and the simulated amount of post-growing season CH₄ emissions agrees well with those measured in 2007.

Most of the discrepancies between simulated and observed CH₄ emissions were within the variability of measurements across subsites within the same peatland type. The 3-hourly and daily RRMSEs ranged from 11.1 % to 22.3 % (Table 2), and the daily RRMSEs were comparable to results presented in Deng et al. (2014). Our results show that model evaluation of CH₄ emissions with finer temporal resolution observations is not necessarily superior to evaluation with coarser temporal resolution, as compared to the NEE counterpart, which could be related to comparatively lesser CH₄ emission variability measured across subsites within the same peatland type (Fig. 6b, d).

3.3 Variability across the permafrost thaw gradient

Thaw rate and ALD increase along the thaw gradient (i.e., palsa to bog to fen), and landscape variations are generally greater than simulated interannual variability (Fig. 7a). Maximum carbon uptake also increases along the thaw gradient, and variations across the landscape are comparable with simulated intraseasonal and interannual variabilities (Fig. 7b). The simulated mean seasonal cumulative NEE values were calculated based on the seasonality identified in Bäckstrand et al. (2010) to help facilitate the intercomparison of carbon budgets estimated at the Stordalen Mire and to better capture the actual seasonality recorded at the study site. The results show that the magnitude of mean growing season CO₂ uptake is highest in the fen and lowest in the palsa (Table 3). The same rank applies to the magnitude of mean CO₂ emissions over the non-growing season, although differences across the thaw gradient are smaller.

CH₄ emission rates increase significantly along the thaw gradient, and the palsa site emissions are negligible (Fig. 7c). Mean cumulative CH₄ emissions simulated in the fen are much higher than those in the bog, and most CH₄ emissions occur during the growing season (Table 3). The higher CH₄ emissions in the fen can be attributed to its faster seasonal thaw rate (Fig. 7a) and a water table depth close to the surface (Fig. 6c). Seasonal cumulative NEE and CH₄ emissions from observations could not be accessed due to the lack of continuous quality-controlled carbon flux measurements during our study period (Table 1).

3.4 Climate sensitivity of permafrost thaw

3.4.1 Thaw responses to climate

Our results indicate that the ALD currently simulated in the bog and fen is around 108 and 130 cm, respectively. However, the maximum depth of our thaw depth measurements is 90 cm, which makes it difficult to evaluate our model performance on ALD simulation. Our results highlight the need to acquire measurements at deeper depth to resolve whether there is no permafrost currently remaining in the bog and fen or there is a talik with permafrost developed deeper than the simulated ALDs. Such information could be important in predicting microbial activity and thermokarst in permafrost peatlands (Schoor et al., 2015), but it may not significantly alter the effects of climate forcing uncertainty discussed in our study.

For each of the four sets of simulations with different climate forcing (Sect. 2.5), simulated mean ALD from 2003 to 2007 is always greatest in the fen and lowest in the palsa (Fig. 8). This consistent trend along the thaw gradient indicates that ALDs are largely regulated by their distinct ecological and hydrological conditions because all three sites had the same climate forcing in each set of simulations (i.e., CTRL, BIASED-COLD, BIASED-WET, and BIASED-

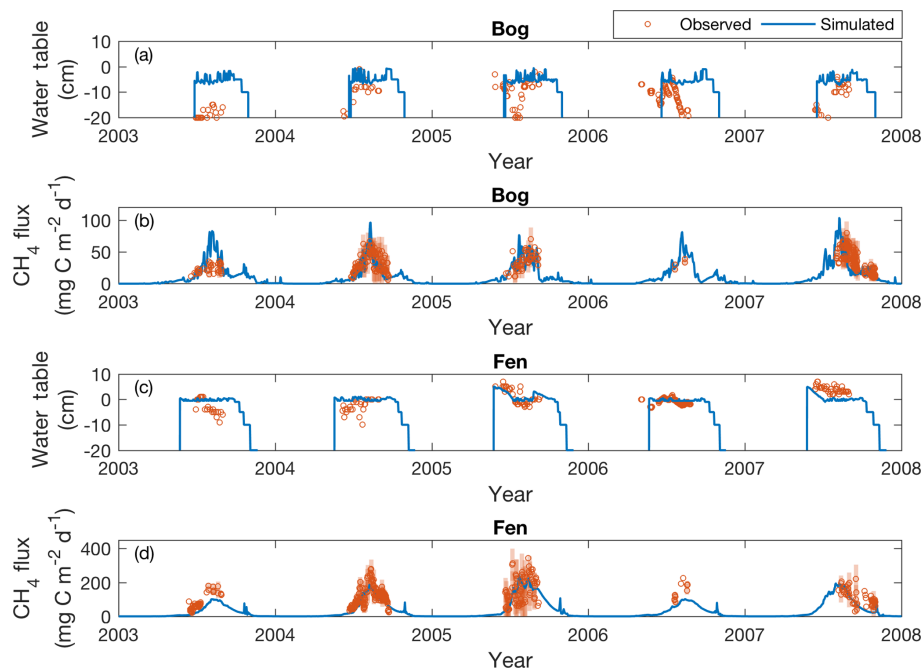


Figure 6. Simulated (solid lines) and measured (open circles) water table depths and daily CH₄ emissions at the bog and fen from 2003 to 2007. Shaded bars are the standard deviations of the daily CH₄ emissions measured across the subsites under each peatland type.

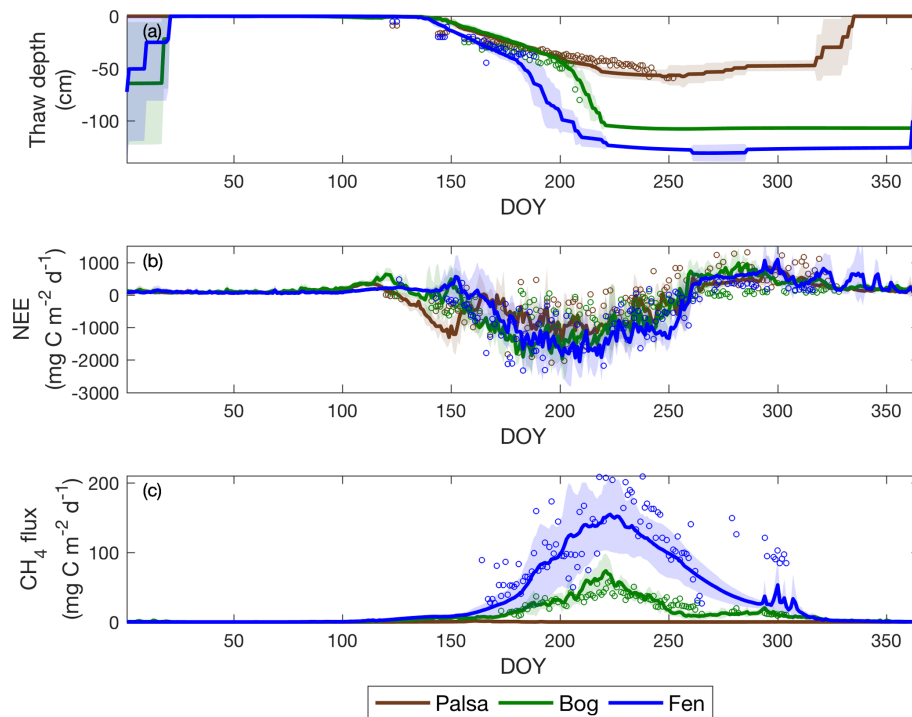
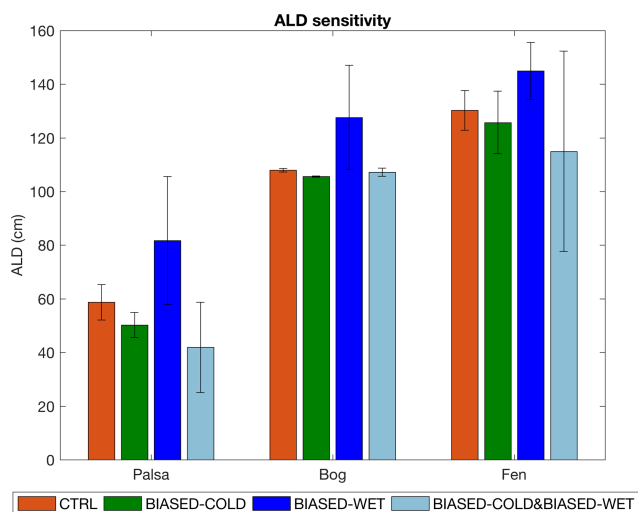


Figure 7. Daily thaw depth (a), daily NEE (b), and daily CH₄ (c) exchanges for the three sites from 2003 to 2007. Solid lines and open circles are the simulated and measured interannual means for each day of the year, respectively. The shaded area is the simulated interannual variability for the corresponding dataset, represented by the standard deviations calculated for each day of the year. Positive and negative carbon flux values indicate effluxes from and influxes to the site, respectively.

Table 3. Means and standard deviations of cumulative CO₂ and CH₄ exchanges simulated in the palsa, bog, and fen during the period 2003 to 2007.

Sites	C flux component	Growing season: days 119–288		Non-growing season: days 1–118 and 289–365	
		Mean	Standard deviation	Mean	Standard deviation
Palsa	CO ₂	−72.70	19.10	38.89	4.09
	CH ₄	0.04	0.02	0.01	0.002
Bog	CO ₂	−79.59	21.46	42.89	2.16
	CH ₄	3.52	0.45	0.42	0.11
Fen	CO ₂	−88.65	7.26	44.41	6.13
	CH ₄	10.86	3.95	0.78	0.18

Note that all gas exchanges are in units of g C m^{−2}.

**Figure 8.** Simulated ALD in the palsa, bog, and fen for four sets of climate forcing (Sect. 2.5). Bars and error bars are means and standard deviations calculated from 2003 to 2007, respectively.

COLD&BIASED-WET). Therefore, the palsa, bog, and fen have different resilience against the changes in climate forcing, and this type of ecosystem resilience plays an important role in determining ALD under changes in climate conditions.

Effects of climate on simulated ALD are similar across peatland types (Fig. 8). With increased precipitation (BIASED-WET vs. CTRL), simulated ALD generally becomes deeper with greater interannual variability because the increased snowpack depth keeps the soil warmer with lower soil ice content during winter. This effect is less prominent in the comparison between experiments BIASED-COLD and BIASED-COLD&BIASED-WET because the cold biases in these two experiments (Sect. 3.1) constrain thaw depth development. For example, summertime soil heating in some of the simulation years was not strong enough to thaw the

soil ice between 20 and 40 cm completely in the BIASED-COLD&BIASED-WET run, resulting in shallower ALDs simulated in the palsa and fen even with the snowpack warming effect. The simulated ALD also becomes deeper with higher air temperature (CTRL vs. BIASED-COLD; BIASED-WET vs. BIASED-COLD&BIASED-WET) in all the examined peatland types. This response is more evident in the comparison between experiments BIASED-WET and BIASED-COLD&BIASED-WET, probably driven by their wet biases (Sect. 3.1) that facilitate thaw depth deepening (via increased thermal conductivity and advective heat transport; Grant et al. 2017a). Similar dependencies between ALD and climate were shown in Åkerman and Johansson (2008) and Johansson et al. (2013), based on multiyear measurements and snow manipulation experiments.

Therefore, the combined cold and wet biases in the GSWP3 climate reanalysis dataset could counteract their individual effects on simulated ALD development at the Stordalen Mire. Our results indicate a 28.6 %, 0.7 %, and 11.7 % underestimation of ALD simulated in the palsa, bog, and fen, respectively, when applying the GSWP3 climate reanalysis data over this region without proper bias correction (BIASED-COLD&BIASED-WET vs. CTRL). Our sensitivity analysis suggests that projected warming and wetting trends (Collins et al., 2013) could significantly increase ALD in the Arctic, since increases in precipitation and air temperature can both contribute to ALD deepening.

3.4.2 Carbon budget responses to climate

Simulations with the four climate forcing datasets (Sect. 2.5) indicate annual mean (from 2003 to 2007) CO₂ sinks and CH₄ sources, except the weak CO₂ emissions simulated in the fen in experiment BIASED-COLD&BIASED-WET due to reduced sedge productivity driven by increased temperature and oxygen stresses (Fig. 9a, b). Our results also indicate that differences in annual CO₂ and CH₄ exchanges across the

four climate forcing datasets for a single peatland type are as large as those across peatland types for a single climate forcing dataset (Fig. 9a, b). These large CO₂ and CH₄ exchange climate sensitivities demonstrate that the peatland's dynamical responses to climate have stronger effects on the carbon cycle than on ALDs (Fig. 8).

With bias-corrected precipitation, increased air temperature (CTRL vs. BIASED-COLD) leads to stronger CO₂ uptake and greater CH₄ emissions in all the examined peatland types (Fig. 9a, b), mainly because enhanced sedge growth facilitates carbon cycling under a warmer environment (results not shown). This air temperature sensitivity affects CO₂ and CH₄ exchanges within the same peatland type without significantly changing ALD (Fig. 8). For both experiments, CO₂ uptake and CH₄ emissions are greatest in the fen and lowest in the palsa, consistent with the measurements reported in Bäckstrand et al. (2010) for the same period. Based on the Coupled Model Intercomparison Project Phase 5 (CMIP5) ESM simulations, arctic annual mean surface air temperature is projected to increase by 8.5 ± 2.1 °C over the 21st century (Bintanja and Andry, 2017). This projected air temperature increase is more than double the air temperature difference between site-observed and GSWP3 temperatures, which could significantly enhance CH₄ emissions regardless of palsa degradation into bog and fen.

On the other hand, wet biases (BIASED-WET and BIASED-COLD&BIASED-WET) increase CH₄ emissions in the palsa; wetter and colder conditions result in as much CH₄ release as the current fen, while wetter conditions alone drive palsa emissions comparable to the current bog (Fig. 9b). The large precipitation sensitivity found in palsa CH₄ emissions could have strong effects on palsa carbon cycling because arctic precipitation is projected to increase by 50 %–60 % towards the end of the 21st century (based on CMIP5 estimates; Bintanja and Andry, 2017). The comparison between experiments BIASED-WET and BIASED-COLD&BIASED-WET shows that in the palsa, increased air temperature strengthens CO₂ uptake and weakens CH₄ emissions. This shift is primarily driven in the model by increased shrub and moss productivity under the warmer environment, which facilitate CO₂ uptake while drying out the soil and reducing CH₄ emissions (results not shown). In the bog and fen sites, increased air temperature under wet bias strengthens both the simulated CO₂ uptake and CH₄ emissions (BIASED-WET vs. BIASED-COLD&BIASED-WET), due to enhanced sedge growth under the warmer environment that facilitates carbon cycling in the experiment BIASED-WET. The low CH₄ emissions in bog and fen simulated in experiment BIASED-COLD&BIASED-WET are driven by increased temperature and oxygen stresses that greatly reduce heterotrophic respiration (CH₄ production) and sedge cover (aerenchyma transport).

We assessed the integrated effects of the changes in CO₂ and CH₄ exchanges identified in the full suite of simulations in terms of the net carbon balance (NCB) and net emis-

sions of greenhouse gases expressed as CO₂ equivalents (net greenhouse gas balance, NGGB). NCB was defined as the sum of the annual total CO₂ and CH₄ exchanges. NGGB was defined in a similar fashion as the NCB but considers the greater radiative forcing potential of CH₄ than CO₂ (28 times over a 100-year horizon; Myhre et al., 2013) when calculating the annual total. The calculated NCB values are mostly negative because the stronger CO₂ uptake dominates the weaker CH₄ emissions (Fig. 9c). The results suggest that all the examined peatland types serve as net carbon sinks under current climate (CTRL), consistent with the estimates reported in Deng et al. (2014) and Lundin et al. (2016). We find a 24, 36, and 38 g C m⁻² yr⁻¹ underestimation of NCB simulated in the palsa, bog, and fen sites, respectively, due to the cold and wet biases in the GSWP3 climate reanalysis dataset (BIASED-COLD&BIASED-WET vs. CTRL). NGGB is affected more strongly by CH₄ emissions (Fig. 9d) due to its larger radiative forcing potential. NGGB values are positive over the bog and fen, suggesting that these sites have positive radiative forcing impacts despite being net carbon sinks. NGGB simulated in the palsa is generally negative (i.e., a net sink from the atmosphere) due to lower CH₄ emissions, except for the simulation conducted without any climate bias correction (correcting only air-temperature-increased CH₄ emissions but not enough to compensate for the significantly higher CO₂ sink). Our results indicate that the simulated NGGB would be biased by 298, -66, and -252 g CO₂ eq. m⁻² yr⁻¹ in the palsa, bog, and fen, respectively, without proper bias correction for the GSWP3 climate reanalysis dataset (BIASED-COLD&BIASED-WET vs. CTRL). Using the GSWP3 products directly thus effectively eliminates the positive radiative forcing from the expanding bog and fen while creating a potentially dramatically inaccurate positive radiative forcing from the shrinking palsa.

3.4.3 Climate sensitivity vs. landscape heterogeneity

Climate sensitivity and landscape heterogeneity are defined here as variability across the four climate forcing datasets for a single peatland type and variability across three peatland types with bias-corrected climate (CTRL), respectively. We estimated carbon cycle variability associated with climate sensitivity and landscape heterogeneity to quantify the corresponding uncertainty in our annual carbon cycle assessments from 2003 to 2007. Our results indicate that differences in simulated annual mean CO₂ exchanges and NCB from climate sensitivity are greater than that from landscape heterogeneity (Fig. 9a, c); i.e., annual CO₂ uptake strength is more sensitive to climate forcing uncertainty than to peatland type representation. In terms of the simulated annual mean CH₄ emissions and NGGB, our results indicate that variability from climate sensitivity is comparable to those from landscape heterogeneity (Fig. 9b, d). Therefore, bias-corrected climate and realistic peatland characterization are

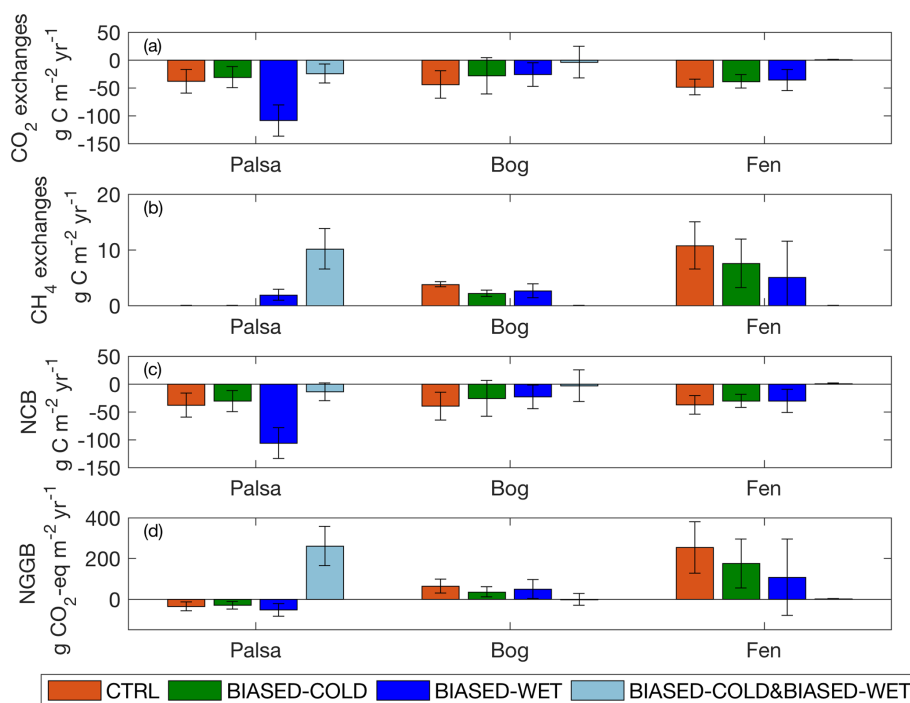


Figure 9. Annual CO₂ exchanges (a), CH₄ exchanges (b), net carbon balance (c), and net greenhouse gas balance (d) simulated in the palsa, bog, and fen, under each set of simulations. Bars and error bars are the means and standard deviations calculated from 2003 to 2007, respectively. Positive and negative values indicate effluxes from and influxes to the site, respectively.

both necessary to reduce the uncertainty in representing carbon cycling dynamics and their radiative forcing effects.

In addition to their effects on carbon cycle predictions, changes in climate conditions also affect permafrost degradation and thus induce changes in areal cover of peatland types. Malmer et al. (2005) showed that there were -0.95 , 0.24 , and 0.62 ha areal cover changes (-10.3% , 4.0% , and 46.3% percentage changes) from 1970 to 2000 in the palsa, bog, and fen, respectively, at the Stordalen Mire. By applying the annual mean CO₂ and CH₄ exchanges simulated with bias-corrected climate from 2003 to 2007, the areal cover changes from 1970 to 2000 alone would lead to -44 kg C yr⁻¹, 76 kg C yr⁻¹, and 2076 kg CO₂ eq. yr⁻¹ changes in annual mean CO₂ exchanges, CH₄ exchanges, and NGGB, respectively, at the Stordalen Mire. The changes in landscape-scale carbon cycle dynamics indicate that the radiative warming impact of increased CH₄ emissions is large enough to offset the radiative cooling impact of increased CO₂ uptake at the Stordalen Mire, consistent with the estimates reported in Deng et al. (2014). The areal cover changes across peatland types could persist or accelerate under the projected warming and wetting trends in the Arctic (Collins et al., 2013; Bintanja and Andry, 2017), which could stimulate CH₄ emissions and produce a stronger radiative warming impact.

4 Conclusions

We evaluated the climate bias in a widely used atmospheric reanalysis product (GSWP3) at our northern Sweden Stordalen Mire site. We then applied a comprehensive biogeochemistry model, *ecosys*, to estimate the effects of these biases on active layer development and carbon cycling across a thaw gradient at the site. Our results show that *ecosys* reasonably represented measured hydrological, thermal, and biogeochemical cycle processes in the intact permafrost palsa, partly thawed bog, and fen. We found that the cold and wet biases in the GSWP3 climate reanalysis dataset significantly alter model simulations, leading to biases in simulated active layer depths, net carbon balance, and net greenhouse gas balance by up to 28.6% , 38 g C m⁻² yr⁻¹, and 298 g CO₂ eq. m⁻² yr⁻¹, respectively. The net carbon balance simulated with bias-corrected climate suggests that all the examined peatland types are currently net carbon sinks from the atmosphere, although the bog and fen sites can have positive radiative forcing impacts due to their higher CH₄ emissions.

Our results indicate that the annual means of active layer depth, CO₂ uptake, and CH₄ emissions generally increase along the permafrost thaw gradient at the Stordalen Mire under current climate, consistent with previous studies in this region. Our analysis suggests that the palsa, bog, and fen differ strongly in their carbon cycling dynamics and have

different responses to climate forcing biases. Differences in simulated CO₂ and CH₄ exchanges driven by uncertainty from climate forcing are as large as those from landscape heterogeneity across the examined permafrost thaw gradient. Model simulations demonstrate that the palsa site exhibits the strongest sensitivity to biases in air temperature and precipitation. The wet bias in GSWP3 could erroneously increase predicted CH₄ emissions from the palsa site to a magnitude comparable to emissions currently measured in the bog and fen sites. These results also show that increased precipitation projected for high-latitude regions could strongly accelerate CH₄ emissions from the palsa area, even without degradation of palsa into bog and fen. Future studies should thus recognize the effects of climate forcing uncertainty on carbon cycling in addition to tracking changes in carbon budgets associated with areal changes in permafrost degradation.

Data availability. The data presented in this study are available at the IsoGenie Database (<https://isogenie-db.asc.ohio-state.edu/datasources/T1/textbackslash#modeling>, last access: 15 February 2019). The ecosys executable program can be obtained from the Ecosys Modelling Project (<https://ecosys.ualberta.ca/download-instructions-2>, last access: 15 February 2019). The ecosys source code is available at <https://github.com/jinyuntang/ECOSYS> (last access: 15 February 2019).

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Author contributions. KYC and WJR designed the study. PMC synthesized field measurements and RFG developed the ecosys model. KYC performed the analyses and led the writing of the paper. All authors contributed thoughtful discussions and insights to the study, and all authors contributed to the editing of the paper.

Competing interests. The authors declare that they have no conflict of interest.

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